

Impact of Lagrangian Sea Surface Temperature Variability on Southern Ocean Phytoplankton Community Growth Rates

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Key Points:

- Sea surface temperature (SST) variability differs between the Lagrangian and Eulerian reference frames
- SST variability decreases phytoplankton community growth rates for multiple generations
- The Q₁₀ growth model does not fully capture the effects of sea surface temperature variability on phytoplankton growth

Abstract

Ocean phytoplankton play a critical role in the global carbon cycle, contributing ~50% of global photosynthesis. As planktonic organisms, phytoplankton encounter significant environmental variability as they are advected horizontally across the upper ocean. The impact of this variability on phytoplankton growth rates has not been quantified and is not captured by many current biogeochemical models. Here, we systematically investigated the impact of different rates and magnitudes of sea surface temperature (SST) variability on phytoplankton community growth rates using surface drifter observations from the Southern Ocean ($> 30^{\circ}\text{S}$) and a phenotype-based ecosystem model. Moderate SST changes of $3\text{--}5^{\circ}\text{C}$ over 7–21 days ($\sim 4\text{--}13$ generations for a typical growth rate of 0.5 day^{-1}) produced the largest time lag between the temperature change and the biological response. Shorter term SST variability (< 7 days) had little impact on the phytoplankton community growth rates. The impact of SST variability was not captured by the Q_{10} -based model of community growth leading to an overestimation of community growth rates, particularly in dynamic, strong frontal regions of the Southern Ocean. Furthermore, we demonstrated that the nature of variability encountered in a Lagrangian reference frame (following trajectories of surface water parcels) differed from that within an Eulerian reference frame, which resulted in significant effects on phytoplankton dynamics. Our results quantify the temporal scales of SST variability relevant for phytoplankton in the Southern Ocean and take a first step towards including the impact of variability and biological response times into numerical models.

Plain Language Summary

Ocean phytoplankton are fundamental to the global carbon cycle. To understand the sensitivity of the global carbon cycle to shifts in climate, we need to understand the impact of environmental variability on phytoplankton growth rates. Phytoplankton encounter environmental variability (e.g. sea surface temperature (SST) changes) in a Lagrangian reference frame. Here, we quantified the variability in SST encountered by phytoplankton *in situ* using surface drifters and investigated the impact of this variability on phytoplankton community growth rates using a phenotype-based ecosystem model. We also compared SST variability in the Lagrangian reference frame to the Eulerian reference frame using high-resolution satellite data. We found significant differences between the two reference frames with larger SST changes in the Lagrangian than in the Eulerian reference frame, and that these differences impacted phytoplankton community structure and growth rates. SST changes $> 0.07^{\circ}\text{C/day}$ induced a lag time between SST changes and the biological response. The impact of SST variability was not captured by the Q_{10} model that is typically used by global biogeochemical models. Our results quantify the temporal scales of SST variability relevant for Southern Ocean phytoplankton and provide a first step towards implementing the biological response to variability into numerical models.

1. Introduction

The ocean is a highly dynamic environment and, in some regions, is expected to become more variable with rising global temperatures (Boyd et al., 2016). Previous work investigating the impact of anthropogenic warming on marine planktonic ecosystems has primarily focused on shifts in the annual mean and/or seasonal conditions. However, recent work has highlighted the importance of considering changes in sub-annual temperature variability, in addition to changes in the mean state (Doblin & van Sebille, 2016; Kroeker et al., 2020). Such short-term variability may be particularly important for phytoplankton dynamics as both mean environmental conditions and sub-seasonal variability affect plankton physiology (Kremer et al. 2018). To better understand how phytoplankton might respond to future environmental changes, such as warming, it is important to characterize the variability these phytoplankton encounter *in situ* and determine the influence of that variability on physiology and community structure.

Phytoplankton are largely passive drifters and, so a Lagrangian, rather than Eulerian, reference frame is most appropriate when considering environmental variability. Both mean conditions and variability (magnitude and rate of change) can differ markedly between the two reference frames (e.g. Doblin and van Sebille 2016; Hellweger et al. 2016). As a result, the temperature regime that sets the thermal acclimation history of a phytoplankton community may differ from prevalent local conditions. For example, Hellweger et al. (2016) modeled the advection of communities across large temperature gradients, such as those along a western boundary current, and showed that the optimum growth temperature (T_{opt}) for the community can be considerably different from the *in situ* temperature. However, the impact on the community growth rate and the extent of differences in the characteristics of Lagrangian versus Eulerian variability – and the circumstances under which these differences have the largest impact on phytoplankton community dynamics – remain unclear.

In global marine ecosystem models, temperature-dependent phytoplankton growth rates are often parameterized via a Q_{10} model based on thermal performance curves generated across multiple temperatures under constant conditions. The Q_{10} parameterization assumes rapid phenotypic shifts by the community such that, as the temperature changes, the community rapidly shifts its optimal growth temperature (Eppley, 1972). Previous modeling studies have demonstrated that the Q_{10} growth parameterization does not capture phytoplankton population dynamics, such as the time-lag between changes in sea surface temperature (SST) and the corresponding change in optimum growth temperature of the community (Hellweger et al. 2016), and results in an over-estimation of phytoplankton community growth rates (Moisan et al., 2002). Additionally, in a lab culture study, Bernhardt et al., (2018) demonstrated that due to non-linearities, fluctuating conditions will alter the shape of the thermal performance curve relative to a constant environment. An improved understanding of temperature-dependent growth rates and how to parameterize these relationships in global models is critical as they have been shown to be one of

the main sources of uncertainty for future climate predictions among global biogeochemical models (Laufkötter et al., 2015).

It has been suggested that phytoplankton responses to changes in temperature depend on previous thermal history (previous acclimation temperature), as well as the magnitude of change and rate of change (Kremer et al., 2018; Pittera et al., 2014). However, results from lab-based studies remain inconclusive. Some studies found an overall decrease in community growth rate in a thermally variable environment relative to a stable environment (Bernhardt et al., 2018; Qu et al., 2019; Wang et al., 2019), while others found higher growth rates under variable conditions (Schaum et al., 2018), and some found that thermal variability did not impact community growth rates (Kling et al., 2019; Qu et al., 2019). The lack of consensus concerning the impact of variability on phytoplankton growth rates may be due to the different magnitudes and rates of change used by the different studies, which ranged from $\sim 1.5^{\circ}\text{C}/\text{day}$ (Schaum et al., 2018) to as high as $10^{\circ}\text{C}/\text{day}$ (Bernhardt et al., 2018).

The direction of SST change, either increasing or decreasing, can also impact the community growth rate due to asymmetry in the thermal reaction norm (Moisan et al., 2002). For phytoplankton in the Southern Ocean, the shape of the reaction norm can be broad and symmetrical, but for most species tends to be asymmetric, often with skewed tails towards lower temperatures (Boyd, 2019). The growth response to changes in SST will depend on the starting SST relative to the optimum growth temperature (T_{opt}) and whether the SST change is increasing or decreasing (Figure 1). When the starting temperature is along the skewed tail and there is an increase in SST, the growth rate will increase (orange arrow, Figure 1). If the starting SST is near T_{opt} , an increase in SST will move beyond the optimum growth and growth rates will decrease (red arrow, Figure 1). Phytoplankton with skewed reaction norms living in water with temperatures at or near T_{opt} can persist through larger decreases in SST than they can increases in SST. Depending on the starting SST, the change in growth rate associated with changing SSTs can either be linear or exponential and the rate of change in growth rate will depend on the acclimation rate and type of acclimation or plasticity (i.e. detrimental or beneficial; Kremer et al., 2018) of the phytoplankton. A detrimental response is one in which the initial growth rate after the temperature shift is greater than the acclimated rate and acclimation to the new temperature results in a decrease in growth rate. A beneficial response is one in which growth rates fall below the acclimated growth rate immediately after the temperature shift but then increase to the acclimated growth rate. When SST changes are slower than the phytoplankton acclimation rate, the instantaneous growth rate will be equivalent to the acclimated growth rate. When the rate of SST change is faster than the rate of acclimation, the instantaneous growth rates could be higher or lower than the acclimated growth rate, depending on the type of response, detrimental or beneficial, respectively (Kremer et al., 2018).

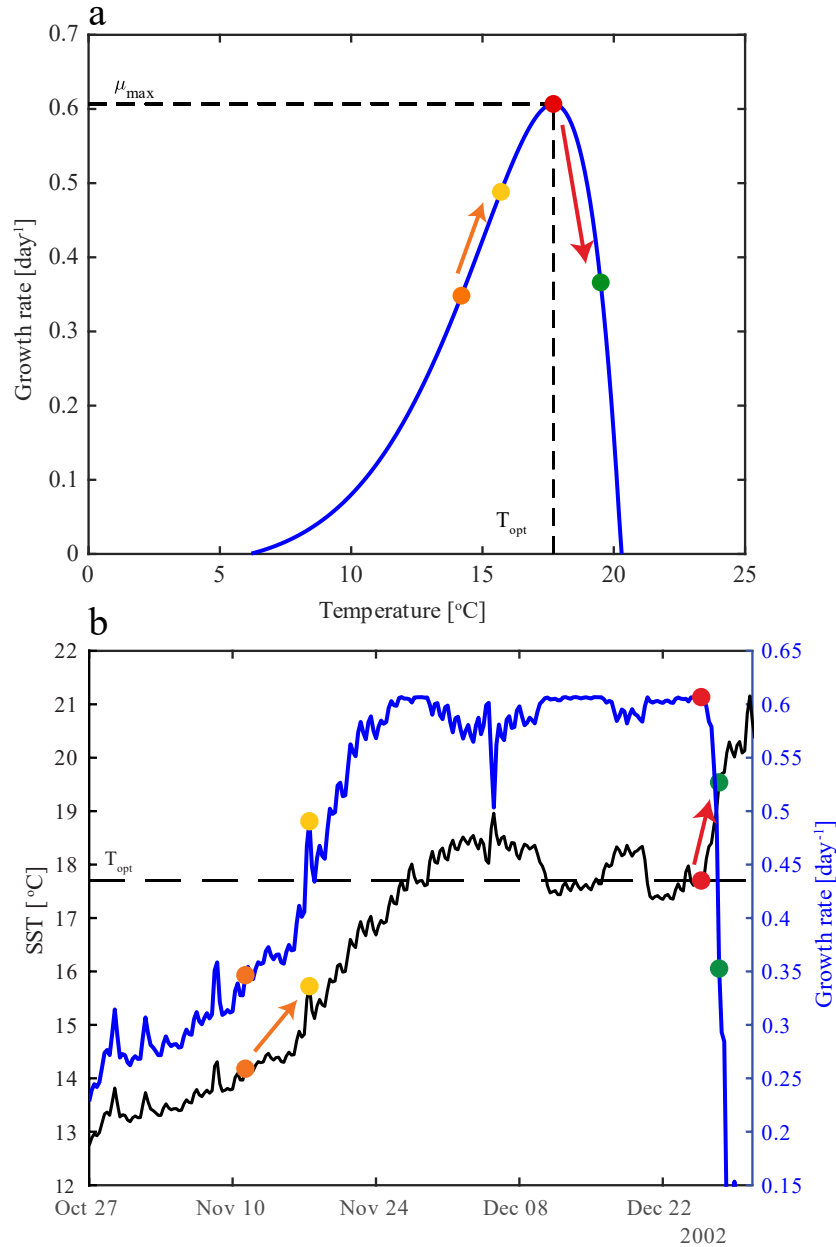


Figure 1. The impact of SST variability on individual phenotype growth rate. (a) The temperature related growth response for a phenotype with a skewed shaped reaction norm. The values for the optimum growth temperature (T_{opt}) and the corresponding maximum growth rate (μ_{\max}) are shown with dashed lines. (b) The 90-day SST profile of an example drifter trajectory (black) and the associated changes in phenotype growth rate (blue). The orange and red arrows in the top panel indicate the change in the phenotype growth rate associated with the corresponding changes in SST in the bottom panel.

Here, we provide a systematic assessment of the effect of different magnitudes and rates of change of temperature on community growth rates and how those might relate to the variability encountered by phytoplankton communities *in situ*. Specifically, using *in situ* and remote sensing SST observations from the Southern Ocean (south of 30°S), we constrained the relevant frequencies and magnitudes of temperature variability that phytoplankton encounter in the Lagrangian reference frame. This southern hemisphere region encompasses some of the lowest (0.2°C) and highest (1.5°C), non-seasonal SST variability globally (Deser et al., 2010) which provided a large range over which to investigate the impact of temperature variability on phytoplankton community growth rates. We used a suite of idealized simulations of temperature shifts and a numerical model to provide mechanistic insight into how this variability might impact community growth rates. We found that relatively small perturbations (< 2 °C over 7 days), which are most common *in situ*, did not substantially impact community growth rates and that moderate changes (4-6°C over 21-45 days) had the largest and longest lasting effect on community growth rates. These moderate changes resulted in a temporary decrease in community growth rate that lasted up to 20 generations, as the community responded to the new temperature. Applying our numerical model to *in situ* SST data from the Southern Ocean, we found that the effect of temperature variability on phytoplankton community growth rates was present everywhere in Southern Ocean with the largest impact occurring in regions dominated by meso- and sub-mesoscale activity.

2. Methods

The impact of SST variability on phytoplankton community growth rates was studied by combining SST observations, both *in situ* and from remote sensing products, and a phenotype-based ecosystem model. Here, we focused on the impact of SST variability on phytoplankton community growth rates and therefore did not consider growth limitations due to other sources of variability such as nutrients, light, and mixed layer depth (Rohr et al., 2020a, 2020b). Further work is needed to investigate these multi-driver impacts.

2.1 Southern Ocean drifter profiles

Lagrangian SST data were obtained from 422 Southern Ocean surface drifters from the Global Drifter Program with 6-hourly SST data. Float data south of 30°S from July 1999 – April 2016 was downloaded from the Drifter Data Centre at the Atlantic Oceanographic and Meteorological Laboratory (accessed 11/2018). The lifetime of the drifters ranged from 91 days to 5.8 years with a median duration of 521 days. Each drifter was segmented into 90-day trajectories to provide consistency in the dataset. We used only segments that had less than 10% of missing data. This resulted in 2,190 90-day trajectories (Figure 2a).

The surface drifter trajectories were used to examine SST variability within a Lagrangian reference frame over different temporal periods. Seasonal dynamics dominated the longer time frames (90 days) while smaller-scale dynamics were important for the shorter windows. We chose to include seasonal dynamics in the analysis rather than filter them out as they were important sources of SST variability encountered by phytoplankton. While surface drifters may have been subjected to some physical movements that phytoplankton do not encounter (e.g., lateral transfer across fronts due to wind rather than subduction and mixing), they provided the best *in situ* dataset for studying Lagrangian variability in surface temperature. To minimize this impact, we limited our analysis to the most frequently measured scales of variability within the drifter record (Section 3.1).

To estimate the magnitude of Lagrangian variability in our study region, we applied moving windows of 7, 21, 45, and 90 days to each of the 90-day trajectories and recorded the absolute value of the maximum range of SST recorded by the drifter during each window ($\Delta\text{SST}_{\text{max}}$) and the time (Δt_{max}) over which the temperature change occurred. To assess the distribution of variability across different window sizes, we aggregated the data into 1°C bins for $\Delta\text{SST}_{\text{max}}$ and Δt_{max} bins of 0-7, 8-21, 22-45, and 46-90 days. For example, a 2.4°C change that occurred over 14 days was recorded in the 2-3°C and 8-21-day bin. To investigate the potential impact of small-scale noise, we also created smoothed splines of each of the 90-day SST profiles using a cubic smoothing spline (*csaps* in Matlab with a smoothing parameter of 0.00001). We then repeated the $\Delta\text{SST}_{\text{max}}$ and Δt_{max} analysis on the spline data.

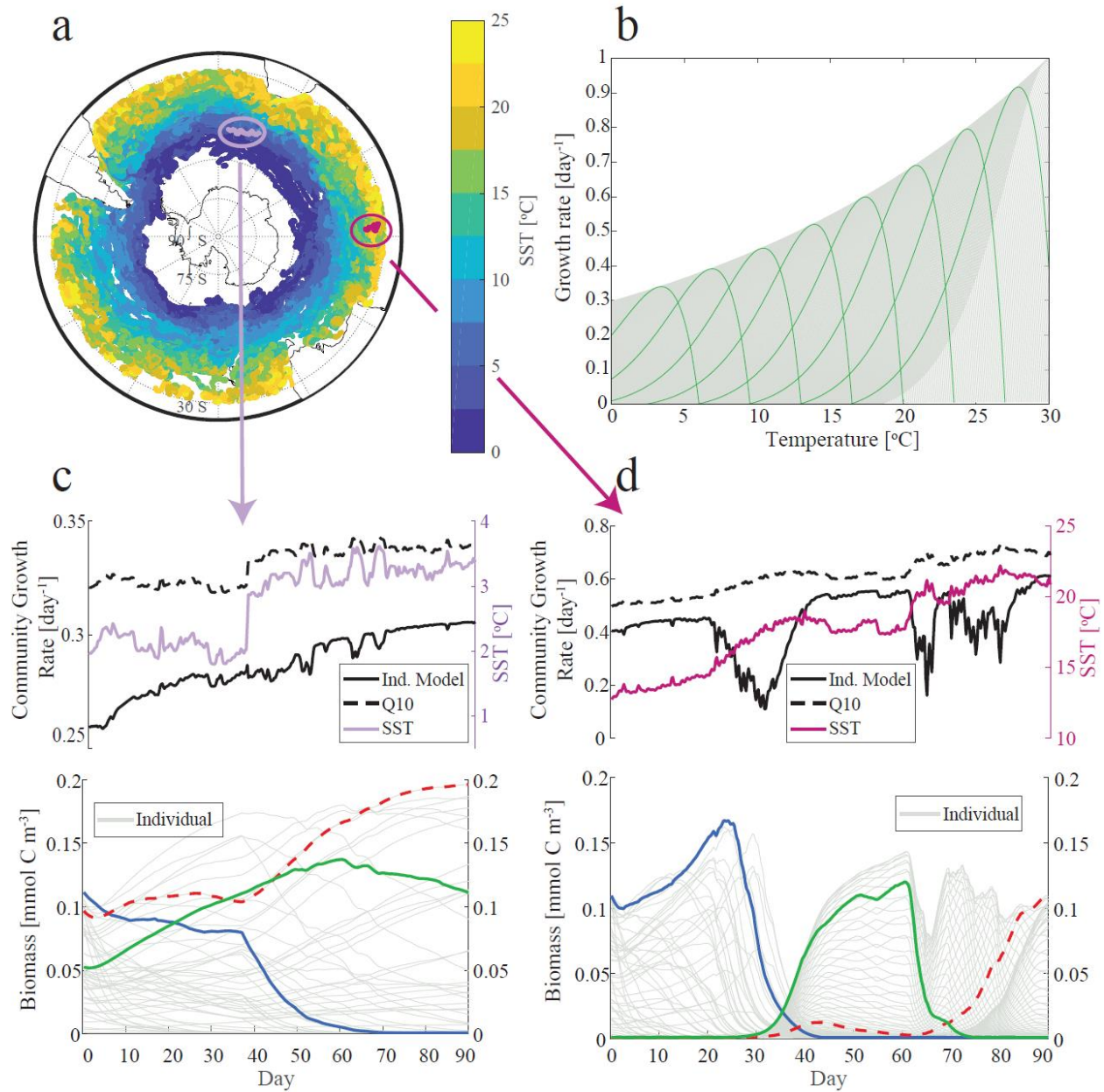


Figure 2. a) Map of all 90-day drifter trajectories ($n = 2190$) colored by SST. Two example trajectories are highlighted in purple and magenta. b) Reaction norms for each of the 319 phenotypes in the ecosystem model. The grey lines represent all the phenotype reaction norms and the green lines are example phenotypes to highlight the reaction norm shape. c and d) Example trajectories and their resulting model outputs. The top panels show the SST (colors), the community growth rate simulated by Q_{10} method (dashed line), and the community growth rate from our phenotype-based model (solid line). The bottom panel shows the growth rate through time of each phenotype (grey lines). The blue line follows the phenotype with the highest initial biomass, the red dashed line follows the phenotype that has the highest biomass at the end of the 90 days, and the green line follows the phenotype that has a T_{opt} equal to the mean SST of the trajectory.

2.2 Remote sensing SST

To compare the SST variability in the Lagrangian reference frame to the variability that would be captured in the Eulerian reference frame, we used high-resolution (0.01° horizontal resolution and 1 day temporal resolution) satellite SST data from GHRSSST Level 4 MUR Global Foundation Sea Surface Temperature Analysis (v4.1) (JPL MUR MEaSUREs Project, 2015; accessed Oct. 2018). This dataset spanned 2003-2014 which overlaps with 71% of our 90-day drifter trajectories. For each 90-day drifter segment between 2003-2014, we extracted 90 days of satellite SST data for the latitude and longitude of the final location of the drifter, where the 90 days corresponded to the dates of the drifter segment. We performed the same $\Delta\text{SST}_{\text{max}}$ and Δt_{max} variability analyses for the satellite data as the surface drifter trajectories (described in Section 2.1).

2.3 Idealized SST profiles

We complemented the observed SST trajectories with idealized SST trajectories to mechanistically understand the impact of the rate and magnitude of SST change on community growth rates. Specifically, a suite of trajectories ($N = 64$) was generated with both increasing and decreasing SST trends ranging from $\Delta\text{SST} = 2^\circ$ to $\Delta\text{SST} = 9^\circ\text{C}$ (in increments of 1°C) over 7, 21, 45, and 90 days. This range was chosen based on our Lagrangian variability analysis. To minimize initialization bias, SST was held constant for the first 30 days before increasing/decreasing. After the SST change, the SST was again held constant until the 200th day. The final temperature for all idealized trajectories was 15°C . The impact of the final temperature on model results was analyzed with a set of sensitivity experiments. The final SST had no significant impact on the results when the results were reported in terms of generation, rather than absolute days as this normalized the effect of higher growth rates at warmer temperatures (Supplemental Information S1).

2.4 Phenotype-based Ecosystem Model

To estimate the impact of variable temperature on phytoplankton community growth rates, we used a phenotype-based ecosystem model. The model consisted of 319 phytoplankton phenotypes that were identical except for the optimum growth temperatures (T_{opt}). Temperature dependent growth rate (μ , day^{-1}) was defined as a function of T ($^\circ\text{C}$) (Thomas et al., 2012):

$$\mu(T) = ae^{bT} \left[1 - \left(\frac{T - T_{\text{opt}}}{w/2} \right)^2 \right] \quad \text{Eq. 1}$$

where T_{opt} was the optimal growth temperature. The value of b controlled the shape of the reaction norm, a (day^{-1}) scaled the reaction norm, and w ($^\circ\text{C}$) defined the width of the reaction norm. We used two sets of reaction norms: a symmetrical, or broad, curve where $b = 0$ ($^\circ\text{C}^{-1}$) and

a skewed reaction norm where $b = 0.3$ ($^{\circ}\text{C}^{-1}$). Both curves had a thermal breadth of 14°C ($w = 20$ $^{\circ}\text{C}$), consistent with observed reaction norms for many polar species (Boyd, 2019). Sensitivity tests were performed with thermal breadths of 10.5°C ($w = 15$ $^{\circ}\text{C}$) and 20.5°C ($w = 29$ $^{\circ}\text{C}$) (Supplemental Information). The results from these sensitivity tests did not differ substantially from the simulations with a thermal breadth of 14°C .

The parameter a scaled the reaction norms at T_{opt} to the Eppley curve (Eppley, 1972) where maximum growth rates ranged between $0.28 - 1.0$ day^{-1} for -1.8°C to 30°C , consistent with experimental data (Boyd, 2019). Specifically, a_i was defined for each phenotype i as:

$$a_i = 0.2963e^{0.0405T_{opt}} \quad \text{Eq. 2}$$

This resulted in a Q_{10} relationship of 1.5, consistent with the apparent Q_{10} from Sherman et al. (2016). We generated 319 phenotype curves for both the broad and skewed reaction norms with T_{opt} ranging from -1.8 $^{\circ}\text{C}$ to 30 $^{\circ}\text{C}$ increasing by 0.1 $^{\circ}\text{C}$ (Figure 2a).

The biomass of each phytoplankton phenotype P_i was calculated at each time-step as the integral of:

$$\frac{dP_i}{dt} = \mu_i(T)P_i - m(T)P_i^2 \quad \text{Eq. 3}$$

where $\mu_i(T)$ was the temperature-dependent growth rate for phenotype i from Equation 1. $m(T)$ was the temperature-dependent quadratic mortality rate ($\text{m}^3 \text{mmol C}^{-1} \text{day}^{-1}$) where:

$$m(T) = 0.1 * a \quad \text{Eq. 4}$$

Here we used the same temperature dependent Eppley curve (Eq. 2) to scale mortality with temperature using SST instead of T_{opt} where $a = 1$ day^{-1} for $\text{SST} = 30^{\circ}\text{C}$. We imposed a minimum biomass (0.001 mmol C m^{-3}) so that no phenotype went locally extinct, akin to the “everything is everywhere” principle (Hutchinson, 1961). Sensitivity tests were performed with the minimum biomass set to 0.0001 mmol C m^{-3} . The minimum biomass threshold did not affect the overall patterns but did increase both the magnitude of the difference from the community growth rates obtained using the Q_{10} model and the time to acclimation (memory length, Section 3.2) for both broad and skewed reaction norms (Supplemental Information). Imposing this minimum biomass purposefully introduced mass into the system which was accounted for by adjusting the biomass of each phenotype to keep the total community biomass at the concentration it would have been without the minimum biomass criteria. Specifically, the total change in biomass without the minimum biomass phenotypes was calculated using the biomass weighted community growth rate (λ) in place of $\mu(T)$ in Eq. 3, where λ was defined as:

$$\lambda = \sum \mu_{i,t} \frac{P_{i,t}}{P} \quad \text{Eq. 5}$$

where, $\mu_{i,t}$ was the growth rate of the i th phenotype at time t for all phenotypes with biomass greater than the minimum, $P_{i,t}$ was the biomass of the i th phenotype whose biomass was greater than the minimum at time t , P was the sum of the biomass of all phenotypes with biomass greater than the minimum at time t .

Several different models for mortality and grazing were tested including linear mortality, constant mortality, a dynamic zooplankton population, and a simple ecosystem model with constant grazing pressure (see section S.4 in Supplemental Material). All model versions resulted in qualitatively similar results which demonstrated that the community dynamics were not particularly sensitive to the top-down control formulation in the model (Section S.4 in Supplemental Material). Here, we present the quadratic mortality as it was the simplest model with smooth (non-oscillatory) solutions.

The ecosystem model was forced with each of the 2,190 drifter segment trajectories (see Figure 2c-e for examples), the corresponding smoothed splines, the idealized SST profiles, and the satellite-derived SSTs. The initial biomass of phenotypes with a T_{opt} within $\pm 2.5^\circ\text{C}$ of the starting SST value were randomized to simulate previously accumulated biomass with phenotypes outside this range set to the minimum biomass. Simulations that used idealized SST profiles were performed 100 times with different initial biomass conditions to account for stochasticity in the model initialization.

3. Results

3.1 SST variability

We characterized *in situ* SST variability using the surface drifter SST data. The average ΔSST_{\max} values ranged from $0.9^\circ\text{C} \pm 0.7^\circ\text{C}$ (1σ) for the 7-day window, which corresponded to $0.13^\circ\text{C}/\text{day}$ change over the 7 days, to $4.2^\circ\text{C} \pm 2.0^\circ\text{C}$ (1σ) for the 90-day window or $0.05^\circ\text{C}/\text{day}$ change (Figure S8, Table S1). The latter was consistent with the expected seasonal SST cycle for the Southern Ocean (Reynolds & Smith, 1994). The SST variability of the drifters was highly correlated with ΔSST_{\max} (Figure S9).

Using the ΔSST_{\max} analysis, we were able to quantify the most common types of variability encountered *in situ* in terms of both the magnitude of change and the rate of change (Figure 3a). Due to the difference in the number of data points generated by the moving windows, the data are presented as the distribution function (fraction) then converted to percent for that window length such that the binned ΔSST_{\max} data sum to 100% over each window length. A 7-day window was most likely to have a ΔSST_{\max} of 2°C or less (82%), 14% of the data had a ΔSST_{\max} envelope of 3°C , and ~3% of the trajectories recorded a ΔSST_{\max} of 4°C . Over a 21-day window,

most trajectories had a $\Delta\text{SST}_{\text{max}}$ of 2-3°C (combined account for 86% of data) and ~10% of the trajectories had a $\Delta\text{SST}_{\text{max}}$ of 4-5°C. As the window length increased, trajectories were more likely to frequently have larger $\Delta\text{SST}_{\text{max}}$. For the 45-day window lengths, $\Delta\text{SST}_{\text{max}}$ of 2-4°C were most common (83% of data), and for 90-day window lengths, $\Delta\text{SST}_{\text{max}}$ of 3-5°C were most common (61% of data). $\Delta\text{SST}_{\text{max}}$ reached as high as 9°C for the 90-day windows but accounted for only 2.5% of the data in that window.

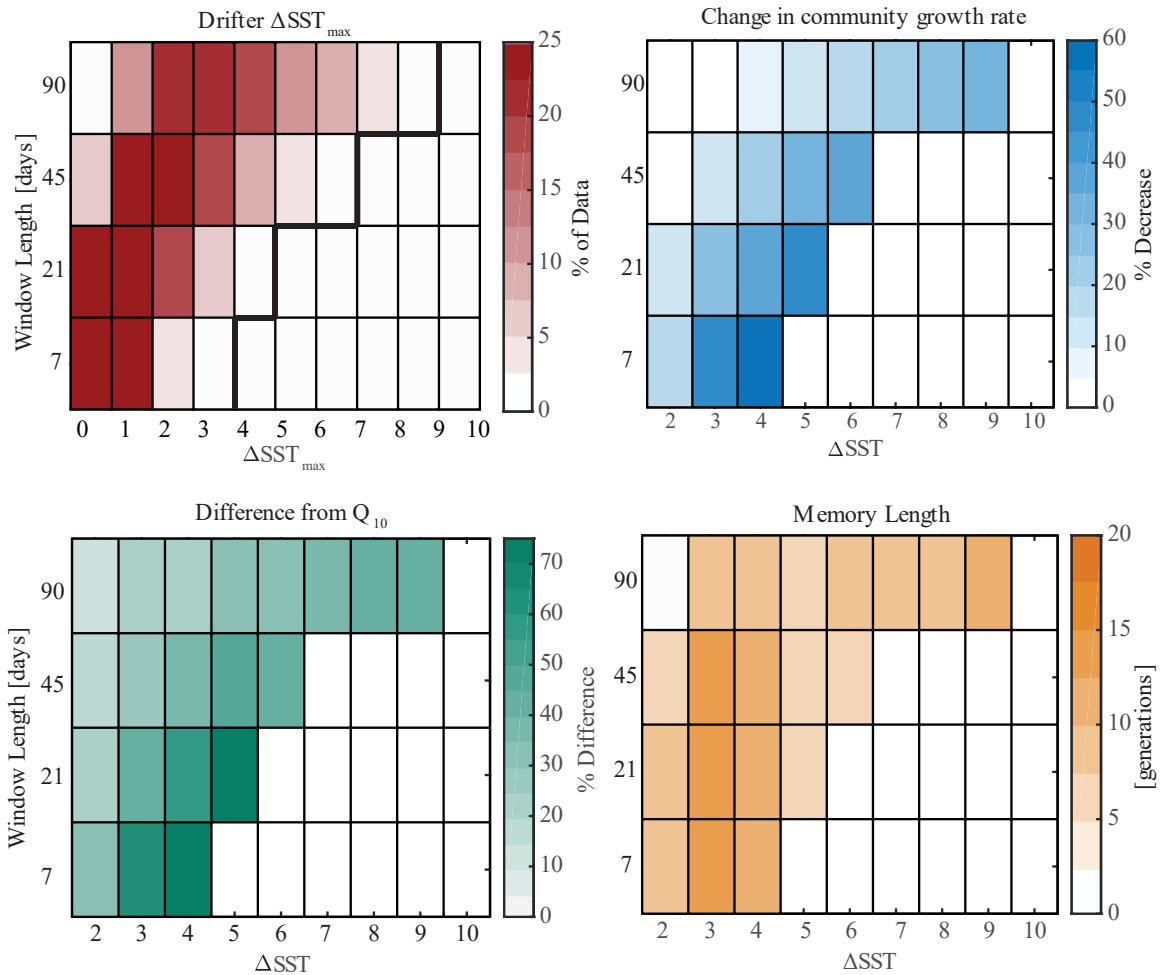


Figure 3. SST variability and the impact on growth rate and memory length for a skewed reaction norm under increasing ΔSST conditions (see Supplemental Material for decreasing ΔSST and broad reaction norm results). Panel (a) shows the frequency of $\Delta\text{SST}_{\text{max}}$ changes in the drifter data for different sized windows. Data are presented as total percentage of data that fall within each window length bin such that each row sums to 100%. Data to right of the thick black line are below a 2.5% occurrence rate and are excluded from the other analyses. Panel (b) plots the decline in community growth rate in the phenotype model that results from the SST moving out of the thermal niche of the original population (see Methods and Figure S12). Data that are greyed out represent ΔSST and window length combinations that were not supported by the results from panel (a). Panel (c) shows the percent difference between the Q_{10} parameterized growth rate and the phenotype modeled community growth rates at the point where SST

stabilizes (see Figure S12 for example). Panel (d) plots the memory effect length associated with SST changes in the idealized simulations. This represents the time it takes for the community growth rate to be within 5% of the steady state growth rate at the final SST from the first time-step that SST is constant (See Figure S12 for example).

Results from the $\Delta\text{SST}_{\text{max}}$ analysis for the satellite data showed that while the Lagrangian and Eulerian reference frames largely recorded similar SST changes, the Lagrangian reference frame was more likely to capture large $\Delta\text{SST}_{\text{max}}$ (Figure S10). For example, over a 7-day window, $\Delta\text{SST}_{\text{max}}$ in a bin around 3°C was almost twice as likely to occur in the Lagrangian (14%) than in the Eulerian (8%) reference frame. For the 21-day window, a $\Delta\text{SST}_{\text{max}}$ of 3-4°C was more likely to occur in the Lagrangian (35%) reference frame than in the Eulerian (23%). Similarly, for the 45-day window, $\Delta\text{SST}_{\text{max}}$ of 2-4 °C were common in both, and larger changes of 4-5°C were more common in the surface drifters (28%) than the satellite data (19%). Across the full 90 days, $\Delta\text{SST}_{\text{max}}$ of 3-5 °C dominated the variability for both reference frames but the surface drifters were twice as likely to record $\Delta\text{SST}_{\text{max}}$ beyond this range (29% of data) compared to the satellite data (15%). Overall, while patterns of variability were similar, the Lagrangian reference frame was more likely to record larger SST changes than the Eulerian. The larger SST changes recorded in the Lagrangian reference frame impact phytoplankton community growth rates in a different manner than the smaller SST changes that overlap between the two reference frames (see Sections 3.2 and 3.3 below). Simulating phytoplankton growth with environmental variability recorded in the Eulerian reference frame, as is common among many global biogeochemical models, omits biologically relevant variability.

For most of the SST data recorded by the drifters, the rate of SST change was slower than the expected phytoplankton acclimation rates. Acclimation rates for the Southern Ocean diatom *F. cylindrus* range from 0.3°C/day to 0.6°C/day (Robert Strzepek, personal communication). For the drifter trajectories, only 8% of all the days analyzed ($n = 197,100$ days) recorded daily rates of SST change higher than 0.3°C/day and less than 2% recorded daily rates of change larger than 0.6°C/day (Figure S11). Because SST rates of change were typically slower than the phytoplankton acclimation rate, we hypothesized that the rate of acclimation would not play a major role in the community response. To simplify model dynamics, we ran our model with rapid acclimation such that each phenotype responded directly to SST changes. See Section 4.1 for discussion about situations in which acclimation may be important.

3.2 Idealized Simulations

We used idealized simulations to develop a mechanistic understanding of how variability impacts community growth rates. First, we used the idealized SST profiles to assess the impact of changing SSTs on community growth rates by calculating the biomass-weighted community growth rate (Eq. 5) at each timestep of the simulation. We found that SST rates of change larger

than 2-3°C in 45-90 days (or 28-57 generations at the stabilized growth rate of 0.5 day⁻¹), which corresponded to a rate of SST change of 0.02-0.07 °C/day, caused community growth rates to decrease, regardless if SSTs were increasing or decreasing. Once SSTs stabilized at the final value, community growth rates increased and eventually stabilized. We tracked the time required for the community growth rate to reach $\pm 5\%$ of the stable value from the timestep at which SSTs stabilized. This time lag or as we call it here, “memory effect”, was then converted to generations rather than days as this allowed us to understand the relative impact of temperature change on phytoplankton using a common currency such that our results are not growth rate dependent. We then compared the biomass-weighted community growth rates from the phenotype model to the community growth rates calculated using the Q_{10} equation to test the impact of rapid phenotype shifts on phytoplankton growth rates. We report only results for those magnitudes of temperature change and durations of change that we observed in the drifter data set (Figure 3a), specifically: $\Delta\text{SST} = 2\text{-}4^\circ\text{C}$ in 7 days (0.29-0.57 °C/day), $\Delta\text{SST} = 2\text{-}5^\circ\text{C}$ in 21 days (0.10-0.24 °C/day), $\Delta\text{SST} = 2\text{-}6^\circ\text{C}$ in 45 days (0.04-0.13 °C/day), and $\Delta\text{SST} = 2\text{-}9^\circ\text{C}$ in 90 days (0.02-0.1 °C/day). Uncertainty estimates reported here are the result of stochastic variability from 100 simulations with different initial conditions. The results from the full range of ΔSSTs are shown in the Supplemental Information (Figures S13 and S14).

3.2.1 Impact of variable SSTs on community growth rates in the phenotype model

For small, gradual SST changes of 2-3°C in 45-90 days (0.02-0.07 °C/day), the community growth rates changed linearly with the SST changes during the period of SST transition and then stabilized once SST stopped changing. That is, the distribution of phenotypes within the community changed at the same rate as the SST such that the T_{opt} of the most dominant phenotype closely matched the SST. As a result, the temporal response in the community growth rate from the phenotype model was similar to the growth rate from a null model using a community Q_{10} parameterization which assumes rapid phenotype shifts in the community in response to changing SST conditions. In all other simulations, when SST changed by more than 3°C, independent of the time over which this change occurred, a different pattern was observed. Growth rates initially increased or decreased depending on the sign of the SST change, but then began to decrease rapidly (see Figure S12 for example). The growth rates continued to decrease until a minimum value was reached. This corresponded to when SSTs stabilized, after which community growth rates increased to their final value and remained there for the rest of the simulation. Growth rates decreased as much as $70\% \pm 1\%$ (1σ) during this low growth period (Figure 3b) and a SST increase of 4°C in 7 days (0.57°C/day) resulted in the largest change (Figure S13). While the absolute percent change in growth rate was partly due to our model formulation, the qualitative results were robust and seen in the other ecosystem models we tested (Figure S7).

Shifts in community growth rates were determined both by changes in the growth rates of individual phenotypes (i.e. shifts along a reaction norm) and shifts in the community

composition (i.e. abundance of different phenotypes). The decrease in community growth rates that accompanied both increases and decreases in SSTs was caused by the SSTs extending beyond the thermal optimum of the initial community such that the bulk of the biomass was growing slowly. During this period, the individual phenotypes with elevated growth rates only made up a small fraction of the community and so did not contribute significantly to the community growth rate. The community growth rates then rebounded as these high growth phenotypes increased their biomass and eventually became the dominate biomass group. Faster rates of SST change moved the community out of the thermal optimum of the initial community more quickly than smaller rates of change, and therefore larger and faster Δ SSTs resulted in greater decreases in community growth rates. However, the high growth individuals were able to dominate the community more quickly due to the high loss rates for the slow (or no) growth individuals and so the community growth rates rebounded more quickly than in the cases of moderate SST change. For rapid SST changes, the rate and type of acclimation response could potentially play a role in the shifts in community growth rates depending on the nature of the plastic response (see *Discussion*).

3.2.2 Memory Effect

The overall magnitude and direction (increasing or decreasing) of SST change, combined with time under transient conditions, and shape of the reaction norm (broad vs skewed) all played a critical role in determining the length of the memory effect – defined here as the amount of time for the community growth rate to stabilize. Critically, our results indicated that the most common Δ SST changes (Figure 3a) were associated with the longest memory effects (Figure 3d). Nearly all Δ SST values tested were sufficient to create a memory effect of longer than 2 generations, and moderate changes of 3-4°C over 7-45 days or 4-28 generations (0.07-0.57 °C/day) resulted in the longest memory effects with slight differences between the reaction norm shapes (Figure 3d, Figure S12). These moderate changes incurred the longest memory effect (up to 22 generations) for both reaction norm shapes (decreasing SST, Figure S14) or up to nearly five times longer than the duration of the Δ SST transient. Larger SST changes (5-6 °C) that occurred over 45 or 90 days or 28-57 generations (0.06-0.13 °C/day) tended to have slightly shorter memory effects (~8-19 generations) than the moderate changes (3-4 °C; 0.03-0.09 °C/day) that occurred over the same time frame (~10-23 generations). This difference was not statistically significant. Longer memory effects for moderate SST changes resulted from dominant phenotypes in the previously acclimated community being able to grow in the new environment, albeit at a reduced rate. This made it harder for the phenotypes optimally suited for the new environment to become more abundant, which resulted in larger memory effects. Communities that underwent large and rapid temperature changes showed the largest short-term decline in community growth rates (Figure S13) but also rebounded more quickly (i.e. had a shorter memory effect) than communities experiencing moderate changes in SST (Figure S14).

The sign of the SST change also impacted the response time of the community. Decreasing Δ SSTs had longer memory effects by an average of 7 generations compared to increasing Δ SSTs (t-test, 95% CI) for the skewed shaped reaction norms. The longer memory effect was due to the long tail on the decreasing side of the reaction norm, which allowed the phenotypes in the initial community to grow during decreasing SST conditions (Figure S14). For reaction norms that were symmetrical about the optimum growth temperature, the direction of Δ SST did not matter, and the memory lengths were statistically the same for increasing and decreasing Δ SSTs (t-test, 95% CI) (Figure S14).

3.2.3 Impact of assuming rapid phenotype shifts via Q_{10} model

Using the Q_{10} equation (only a function of SST) to calculate community growth rates resulted in perfect environmental tracking, or effectively (unrealistic) rapid phenotype shifts such that the T_{opt} of the community perfectly followed SST (Figure S12). However, as discussed in Section 3.2.1, the timescales required for rearrangement of the community composition resulted in decreased community growth rates in response to SST variability. To compare the Q_{10} community growth rates to the phenotype-based model growth rates, we calculated the percent difference between phenotype-based model and the Q_{10} model relative to the Q_{10} model (as $[(Q_{10} - \text{phenotype}) * 100] / Q_{10}$) at the time step when SSTs stabilize, which typically corresponds to the time of minimum growth in the phenotype-based model (Figure 3c and S12). Community growth rates derived from the Q_{10} model were always larger than those simulated by the phenotype model, which was consistent with previous work that suggested that SST variability will result in lower growth rates than when temperature variability is not considered (Moisan et al. 2002; Bernhardt et al. 2018). However, the difference from Q_{10} was not constant but rather a function of SST variability. As the Δ SST increased over a given window length, so did the difference between the phenotype model and the Q_{10} model. The largest departures from Q_{10} occurred for Δ SSTs of 4°C and 5 °C over 7 and 21 days, respectively, with up to 80% lower simulated community growth rates for the phenotype model. The magnitude of this difference was consistent with the modeling study of Moisan et al. (2002) who found that the Q_{10} growth rate model over-estimated individual phenotype growth rates and population growth rates. The smallest difference between the models was for Δ SST = 2 °C over 90 days which resulted in 2.5-5% lower growth rates. Generally, larger Δ SSTs and faster rates of change (changes occurring over a few generations) resulted in larger differences between the models.

3.3 Southern Ocean Drifter Trajectories

The idealized simulations allowed for a mechanistic explanation of the impact SST changes have on phytoplankton community growth rates. However, in the ocean, SST change is much more complicated as phytoplankton are exposed to a large variety of rates and duration of SST changes. We used Southern Ocean drifter trajectories to investigate the impact of *in situ* SST variability on community growth rates.

Using the Southern Ocean drifter trajectories, we found that, like the idealized simulations, the phenotype model resulted in lower average community growth rates compared to the Q_{10} model over the entire 90-days (Figure 4). Like the idealized simulations, as drifter SST variability increased, so did the difference between the phenotype model and the Q_{10} model. The mean percent difference between the phenotype model and the Q_{10} model for the trajectories ranged between 9.0% - 39.8% for the skewed reaction norms. A similar pattern was observed for the broad shaped reaction norms, but the magnitude of the difference was smaller and ranged from just 0.6% to 24.6% different (Figure S15). Trajectories with higher mean SSTs were affected less by SST variability than trajectories with lower SSTs. This was due to faster growth rates at higher temperatures, which allowed more rapid responses to SST changes and shorter memory lengths in terms of days (when normalized to generations there was no difference).

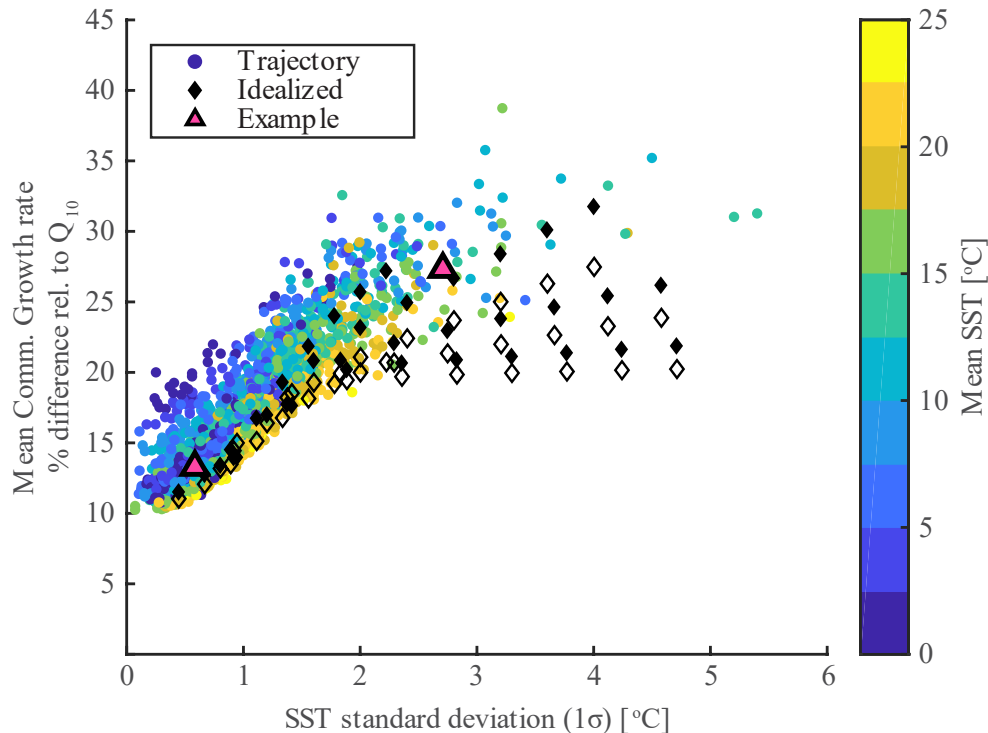


Figure 4. Impact of SST variability on community growth rate. The average percent difference in community growth rate between the phenotype model and the Q_{10} growth model from the 90-day drifter segments are plotted against the standard deviation (1σ) of the drifter SST. Each segment is colored by the mean SST. Results from the idealized trajectories are shown as black circles with filled circles denoting increasing SST trajectories and open circles denoting decreasing SSTs. Pink triangles represent the two example trajectories from Figure 2. Results shown here are for skewed shaped reaction norms, see Figure S15 for results for the broad shaped reaction norms.

To isolate the impact that short-term variability may have had on community growth rates relative to longer-term shifts, we compared the 90-day mean biomass-weighted community

growth rate of the drifter trajectories to the smoothed splines derived from the trajectories. The smoothed splines (i.e. low frequency variability) accounted for a varying percentage of the overall SST variability from a median value of 27% when looking at a 1-day window to 93% when looking at a 90-day window (Figure S16). We found that removing short-term variability had no impact on community growth rates (t-test, 95% CI; Figure S17).

4. Discussion

4.1 Impact of Acclimation

Acclimation becomes potentially important when the rate of acclimation is slower than the rate of SST change. To assess the potential impact of an acclimation timescale that was slower than the SST change on our results, we ran the phenotype-based model with a range of acclimation rates (0.2 °C/day to 0.6 °C/day (Robert Strzepek, personal communication) and Δ SST changes of 2-4°C in 7 days (0.29-0.57 °C/day) and 5°C in 21 days (0.24 °C/day) (Section S.7 in Supplemental Material). The results were generally consistent with the rapid acclimation results which suggested that acclimation timescales have a secondary impact on community growth rates compared to phenotypic shifts. The slower acclimation rates did shift the balance between phenotypic change in growth rate and the community composition shift (Section 3.2.1) and resulted in both a delay in the decrease in community growth rates and a smaller magnitude decrease (Figures 5a and S18). The memory effect increased linearly with a decrease in acclimation rate (longer acclimation time) (Figures 5b and S19).

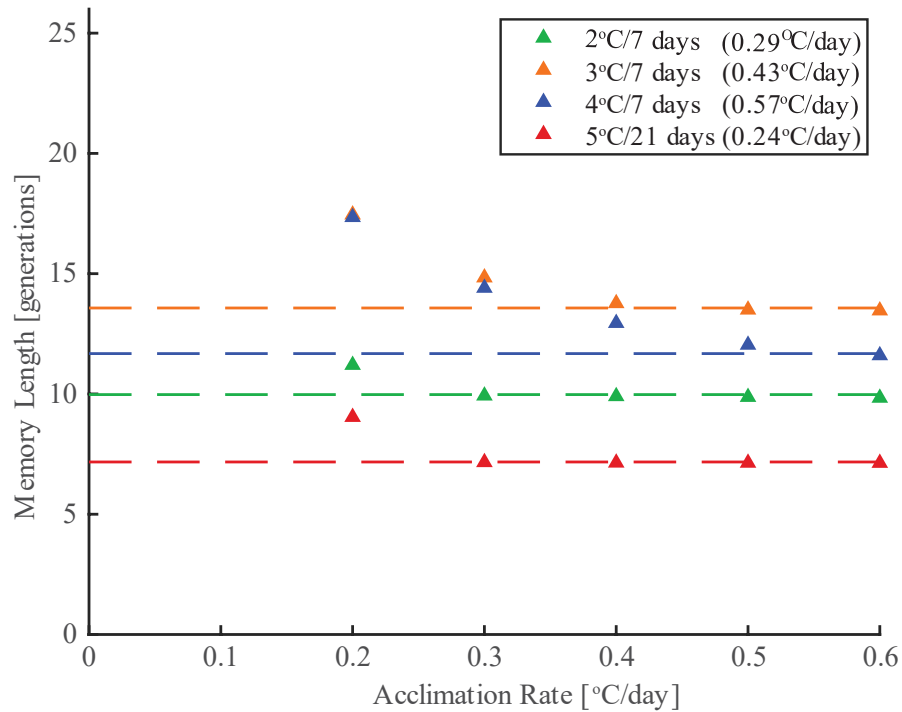


Figure 5. The impact of acclimation on memory lengths. Acclimation rates that were slower than the rate of SST change resulted in longer memory lengths than for simulations in which acclimation rate was equal to or faster than the SST rate of change.

The phenotypic plasticity represented in the model was a simplistic representation of plasticity. Specifically, we assumed that the phenotype slowly adjusted its growth rate by moving along the reaction norm at the rate defined by the acclimation timescale. As such, whether the plastic response was “beneficial” or “detrimental” (Kremer et al., 2018) depended on whether SSTs were increasing or decreasing and whether the initial SST was above or below T_{opt} . In reality, plastic responses are much more complex and nonlinear and most likely vary among species (Kremer et al., 2018). Additional work is needed to better constrain both the range or acclimation timescales and the mechanisms of phenotypic plasticity. However, our results suggest that these dynamics will only become important under rapid temperature changes which are infrequent in the ocean.

4.2 Implications for in-situ community composition

Our findings support the important role of thermal history in shaping the response of phytoplankton communities to changes in temperature. We have shown that SST variability can lower community growth rates for tens of generations following SST perturbation. This indicates that, for many regions of the ocean, the phytoplankton community will not be fully acclimated to

local conditions as a result of the mismatch between timescales of physical variability and biological response. This mismatch in timescales will be a function of the rate and magnitude of variability that phytoplankton in the water mass were previously exposed to and may be reflected in physiological properties such as optimum growth temperature or overall community growth rate.

Our results also provide an important extension on the classic principle that “everything is everywhere: but the environment selects” (Hutchinson, 1961). Even when ‘everything is everywhere’, we show that the timescale for environmental selection (community replacement) is a critical factor in determining community composition. Specifically, we hypothesize that even when the ‘optimal’ organism is present in an environment, SST variability generated by local physics, lateral advection, and seasonal trends can delay or prevent that organism from dominating the community. This hypothesis is supported by previous modeling work that has shown a time-lag on the order of weeks to a month in the phytoplankton community growth response to SST changes due to lateral advection and seasonal trends (Moisan, et al., 2002; Hellweger et al. 2016). Here, we have quantified the relationship between varying rates of SST variability and the timescale required for community replacement to impact the community composition.

Underlying the time lag of the community response to SST variability were differences in the distribution of phenotypes. When rates of change were slow, the community was able to keep up with the changing SSTs and the T_{opt} of the most abundant phenotype matched the changing SSTs (i.e. there is no memory effect and difference from Q_{10} is small). However, as SSTs became more variable, the dynamics described above (Section 3.2) resulted in a mismatch between *in situ* SST and the T_{opt} of the dominant phenotype (which resulted in larger offsets from Q_{10} and a memory effect). Further, our results indicated that the difference in the nature of SST variability between the Eulerian and Lagrangian reference frames could cause different community growth rates and phenotype distribution.

The influence of lateral advection on the phenotype distribution of the community was evaluated by comparing differences in the offset between the T_{opt} of the most abundant phenotype and the SST of the simulation. We compared the SST at the end of each 90-day segment to the phenotype T_{opt} offset (defined as SST minus T_{opt} of the most abundant phenotype at the final time step). While the final SST of the drifter segments and satellite data were not statistically different (t-test, 95% CI, Figure S20), differences in the nature of variability in the proceeding 90 days resulted in a significant difference between the final SST and the T_{opt} of the most abundant phenotype (t-test, 95% CI, Figures 6 and S21). The magnitude of the offset between SST and T_{opt} depended on the timing of SST changes throughout the 90-day profiles. Consistent with the results from the idealized simulations, when SST changes were slow, the offset between SST and

the T_{opt} of the most abundant phenotype were negligible (Figure S22 for an example) while larger and/or faster SST changes resulted in larger offsets (Figure S23 for an example). Large SST changes that occurred early in the 90-day segment allowed sufficient time for the community to respond and for the T_{opt} of the community to reflect the SST change, assuming the SST profile stayed relatively stable after the initial SST change (see drifter data in Figure S22). When SST changes occurred later in the 90-days, the community did not have sufficient time to respond which caused a larger offset between the SST at day 90 and the T_{opt} of the community (see satellite data in Figure S23). Different phenotype distributions for the Eulerian versus Lagrangian reference frames is consistent with previous results that showed advection of phytoplankton communities was a key process in shaping phytoplankton diversity (Barton et al., 2010; Clayton et al., 2013; Lévy et al., 2014).

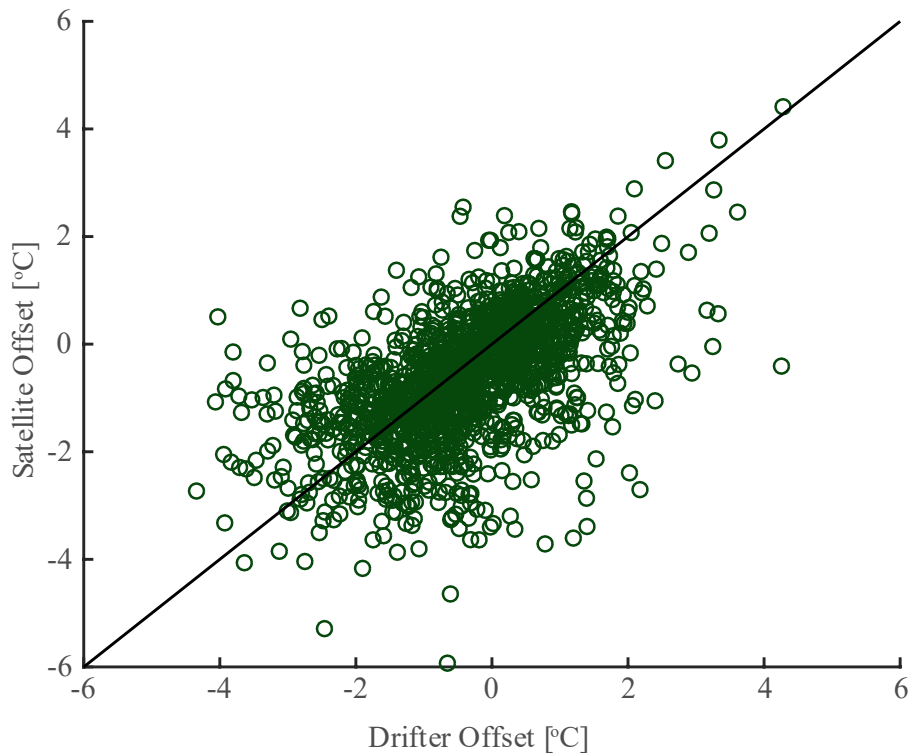


Figure 6. The impact of Lagrangian and Eulerian variability on community composition. Here we plot the difference between the T_{opt} of the most abundant phenotype at the end of each 90-day trajectory and the final SST for the drifter trajectory (x-axis) and the satellite data (y-axis). The final SSTs for the drifter and satellite data are statistically identical (t-test, 95% CI). Therefore, deviations from the 1:1 line demonstrate the impact of a Lagrangian versus Eulerian reference frame on community composition.

Our results support the hypothesis of Moisan et al. (2002) that the shape of the temperature reaction norms has a significant impact on the community response to temperature variability.

Additionally, we suggest that the nature of the environmental variability may play an important role in determining which reaction norm shapes will be most regionally competitive. Specifically, under decreasing temperatures, a phenotype with a skewed reaction norm (T_{opt} is closer to T_{max}) is predicted to have a competitive advantage over a phenotype with a broad reaction norm (T_{opt} at the center of the range), given the same range and T_{opt} . The increased competitive advantage is because the skewed reaction norm provides a larger range of temperatures less than T_{opt} under which the phenotype can grow and, consequentially, a more gradual decline in growth rate as a function of decreasing temperatures. Therefore, organisms with skewed reaction norms should be adapted to have T_{opt} values close to maximum encountered temperatures not only due to the rapid drop in growth rates for temperatures greater than T_{opt} (Thomas et al., 2012) but also due to the competitive advantage under temperatures less than T_{opt} (Moisan et al., 2002). Conversely, broad reaction norms are favored when temperatures are warming, as expected, or when temperatures are more variable. Our results have implications for the global distribution of reaction norm shapes. One would expect more skewed reaction norms in the tropics where the warmest waters are found making it easier to evolve a T_{opt} close to the maximum encountered temperature as this upper bound is constrained (Thomas et al., 2012). However, in the polar regions where the temperatures encountered by phytoplankton is much more variable, there should be selective pressure for broad reaction norms with large growth ranges beyond T_{opt} as suggested by Moisan et al. (2002) or skewed reaction norms where T_{opt} is higher than mean SSTs (Thomas et al., 2012).

4.3 Implications for simulating community growth rates in global biogeochemical models

The Q_{10} temperature-growth response is widely used in ecosystem models. The premise behind employing a Q_{10} growth equation is that each modeled functional group encompasses many species or strains with successive optimum growth temperatures. Therefore, a temperature change will merely shift the distribution of species/strains in the group such that the one with a T_{opt} matching the new temperature will dominate the community. However, as we have demonstrated, the timescale of this shift is important and is a function of the rate, magnitude, and direction of temperature change and the shape of the species/strains' thermal response curve. As a result, this succession does not occur as rapidly as represented by the Q_{10} function. In fact, community growth rates often temporarily declined until the phenotype with T_{opt} near ambient conditions overcame the previously accumulated biomass. Because of the temporary decrease in community growth rates, the Q_{10} temperature-growth response not only overestimates temperature-related community growth but does so as a function of SST, SST variability, and reaction norm shape. Critically, this indicates that adjusting the Q_{10} relationship to use a lower exponent as previously suggested (Sherman et al., 2016) will only partially capture realistic dynamics.

In addition to not capturing the full effect of SST variability on growth rates, biogeochemical models may not be adequately capturing relevant variability. The current framework of most

biogeochemical models requires them to be integrated in an Eulerian reference frame. We have shown that the Eulerian variability can differ substantially from the Lagrangian reference frame (Figure S11). While the overall magnitude of the variability was similar between the Eulerian and Lagrangian reference frames, we also found that the character of that variability was different (Figure S11) and the Lagrangian reference frame recorded more variability that generated longer memory effects in the phytoplankton community growth rates. The spatial patterns of SST variability in the Southern Ocean were similar between Eulerian and Lagrangian reference frames (Figure 7a,c) as was the effect on community growth rates (Figure 7b,d). However, due to the dampened variability in the Eulerian reference frame relative to the Lagrangian reference frame, the offset between community growth rates simulated using the Q_{10} equation and the phenotype model was also lessened. Models such as DARWIN (Follows et al., 2007) resolve phenotypes with a range of thermal reaction norms and so will capture the community growth rate dynamics presented here. However, additional work is needed to compare the variability encountered by functional group phenotypes in large-scale models integrated in an Eulerian framework to true Lagrangian variability.

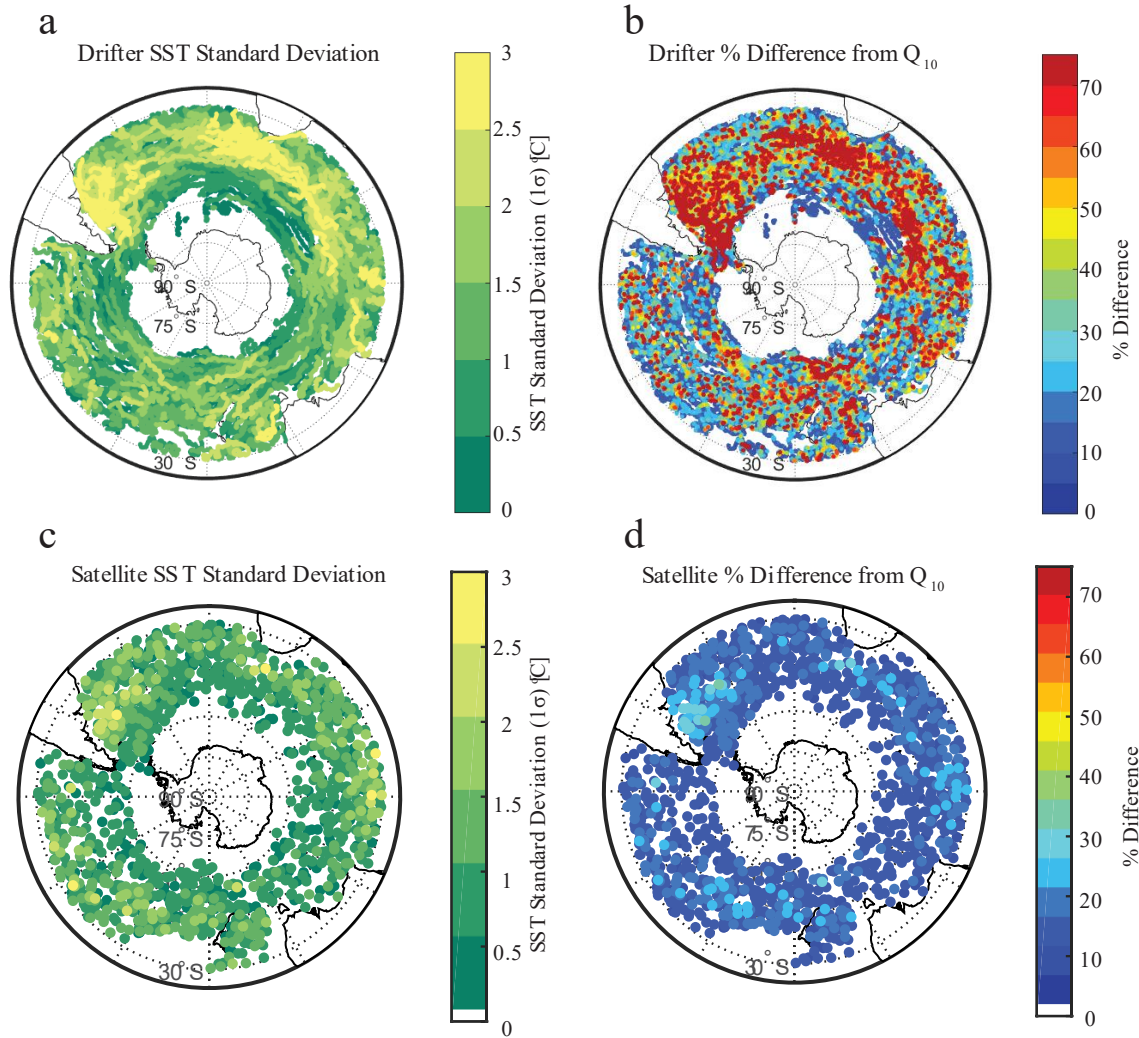


Figure 7. Distribution of SST variability (a, c) and the deviation in community growth rate from Q₁₀ (b, d) over the Southern Ocean (>30°S). Top row shows drifter data and bottom row shows satellite data. Three key regions of high SST variability stand out: Malvinas-Brazil confluence zone, the Agulhas Retroflection, and the Subtropical front. These regions have enhanced SST variability in both datasets but higher variability in the drifters. These high variability regions correspond to large differences between the phenotype model growth rates and Q₁₀.

Improving the parameterized temperature-growth relationship may be particularly important in the Southern Ocean. We used the model results described above to identify key regions within the Southern Ocean that might be most strongly impacted by temperature variability. Three particular regions stand out that exhibited the most SST variability and had the largest relative deviations from the Q₁₀ model: the Malvinas-Brazil confluence zone; the Agulhas Retroflection region; and downstream from these two along the Subtropical Front near ~45°S, 60°E (Figures 7a,b). All three regions were previously identified as highly dynamic, strong frontal regions (Artana et al., 2019; Beal et al., 2015; Graham & Boer, 2013) and shown to be important hot-

spots for phytoplankton diversity (Barton et al., 2010; Clayton et al., 2013; d'Ovidio et al., 2010; Soccodato et al., 2016). These regions represent the boundaries where cold, fresh polar water meets the warm, salty subtropical where abundant mixing via meso- and sub-mesoscale processes can occur (Clayton et al., 2017; d'Ovidio et al., 2010) and lead to the highly variable SSTs recorded by the floats. It is possible that in these highly dynamic frontal regions the floats were subjected to physical movements across the fronts that was previously thought to elude phytoplankton movements. However, recent field and modeling studies have shown that cross-front transfer and diapycnal mixing can occur due to the fine-scale physics associated with these strong fronts (Clayton et al., 2017; Mahadevan, 2016; Wenegrat et al., 2020)

Our results also showed that large SST changes were not required for temperature variations to have a lasting impact on community growth rates. While large differences between Q_{10} community growth rates and the phenotype-based model community growth rates were common in the regions mentioned above, other regions of the Southern Ocean that had moderate (2-3.5 °C, 1σ) SST variability also recorded equally large differences in community growth rate, often at least 45% smaller than Q_{10} estimates and up to 94.5% different. This was consistent with our results from the idealized simulations which showed that moderate temperature changes can have large impacts on community growth rates.

5. Conclusions

In this study we utilized idealized SST simulations and SST data from ocean surface drifters to show that synoptic SST variability on time-scales of a few days to a few weeks temporarily decreases phytoplankton community growth rates, while higher frequency variability has little impact. The time it took for the community growth rate to reflect the new environment was dependent upon the rate and magnitude of temperature change, the direction of change, and the shape of the species/strains' thermal response curve. The largest memory effects resulted from moderate changes in SST that occur over 1-3 weeks. This impact of SST variability can potentially cause a large offset between a phenotype-based temperature-dependent community growth rate and a Q_{10} based estimate and suggests that phytoplankton communities sampled *in situ* may often not be adjusted to local conditions. Given the highly variable nature of the ocean and importance of environmental variability for phytoplankton physiology, it is critical to consider the correct reference frame and the magnitude and duration of variability when studying phytoplankton dynamics. Here we demonstrated that the Lagrangian reference frame captured by drifters was, in many instances, not equivalent to the Eulerian frame and that this had significant impacts for estimating phytoplankton growth rates.

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Figure 1. The impact of SST variability on individual phenotype growth rate. (a) The temperature related growth response for a phenotype with a skewed shaped reaction norm. The values for the optimum growth temperature (T_{opt}) and the corresponding maximum growth rate (μ_{max}) are shown with dashed lines. (b) The 90-day SST profile of an example drifter trajectory (black) and the associated changes in phenotype growth rate (blue). The orange and red arrows in the top panel indicate the change in the phenotype growth rate associated with the corresponding changes in SST in the bottom panel.

Figure 2. a) Map of all 90-day drifter trajectories ($n = 2190$) colored by SST. Two example trajectories are highlighted in purple and magenta. b) Reaction norms for each of the 319 phenotypes in the ecosystem model. The grey lines represent all the phenotype reaction norms and the green lines are example phenotypes to highlight the reaction norm shape. c and d) Example trajectories and their resulting model outputs. The top panels show the SST (colors), the community growth rate simulated by Q_{10} method (dashed line), and the community growth rate from our phenotype-based model (solid line). The bottom panel shows the growth rate through time of each phenotype (grey lines). The blue line follows the phenotype with the highest initial biomass, the red dashed line follows the phenotype that has the highest biomass at the end of the 90 days, and the green line follows the phenotype that has a T_{opt} equal to the mean SST of the trajectory.

Figure 3. SST variability and the impact on growth rate and memory length for a skewed reaction norm under increasing ΔSST conditions (see Supplemental Material for decreasing ΔSST and broad reaction norm results). Panel (a) shows the frequency of ΔSST_{max} changes in the drifter data for different sized windows. Data are presented as total percentage of data that fall within each window length bin such that each row sums to 100%. Data to right of the thick black line are below a 2.5% occurrence rate and are excluded from the other analyses. Panel (b) plots the decline in community growth rate in the phenotype model that results from the SST moving out of the thermal niche of the original population (see Methods and Figure S12). Data that are greyed out represent ΔSST and window length combinations that were not supported by the results from panel (a). Panel (c) shows the percent difference between the Q_{10} parameterized growth rate and the phenotype modeled community growth rates at the point where SST stabilizes (see Figure S12 for example). Panel (d) plots the memory effect length associated with SST changes in the idealized simulations. This represents the time it takes for the community growth rate to be within 5% of the steady state growth rate at the final SST from the first time-step that SST is constant (See Figure S12 for example).

Figure 4. Impact of SST variability on community growth rate. The average percent difference in community growth rate between the phenotype model and the Q_{10} growth model from the 90-day drifter segments are plotted against the standard deviation (1σ) of the drifter SST. Each segment is colored by the mean SST. Results from the idealized trajectories are shown as black circles with filled circles denoting increasing SST trajectories and open circles denoting decreasing SSTs. Pink triangles represent the two example trajectories from Figure 2. Results shown here

are for skewed shaped reaction norms, see Figure S15 for results for the broad shaped reaction norms.

Figure 5. The impact of acclimation on memory lengths. Acclimation rates that were slower than the rate of SST change resulted in longer memory lengths than for simulations in which acclimation rate was equal to or faster than the SST rate of change.

Figure 6. The impact of Lagrangian and Eulerian variability on community composition. Here we plot the difference between the T_{opt} of the most abundant phenotype at the end of each 90-day trajectory and the final SST for the drifter trajectory (x-axis) and the satellite data (y-axis). The final SSTs for the drifter and satellite data are statistically identical (t-test, 95% CI). Therefore, deviations from the 1:1 line demonstrate the impact of a Lagrangian versus Eulerian reference frame on community composition.

Figure 7. Distribution of SST variability (a, c) and the deviation in community growth rate from Q_{10} (b, d) over the Southern Ocean ($>30^{\circ}\text{S}$). Top row shows drifter data and bottom row shows satellite data. Three key regions of high SST variability stand out: Malvinas-Brazil confluence zone, the Agulhas Retroflection, and the Subtropical front. These regions have enhanced SST variability in both datasets but higher variability in the drifters. These high variability regions correspond to large differences between the phenotype model growth rates and Q_{10} .

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