

1 **Is it worthwhile to invest in learning? A stormwater management case study with** 2 **green infrastructure using Bayesian-based optimization**

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8 **Key Points:**

- 9 • We propose an adaptive green infrastructure planning framework that combines stochastic
10 tastic optimization and hydrological simulation
- 11 • The Bayesian-based model considers learning-by-doing and that experience learned is
12 transferable between locations
- 13 • We quantify the economic value of learning that can support the use of adaptive planning
14 approaches

15 **Abstract**

16 To cope with the uncertainty of green infrastructure planning at city scale, many cities take an
17 adaptive approach and use learning-by-doing to improve understanding of the urban systems.
18 However, whether that learning is worth it has been a challenge to adaptive management practi-
19 tioners. In this paper, we propose an evaluation and planning framework for green infrastructure
20 (GI) to address this issue and demonstrate its use by an application to the Wingohocking water-
21 shed, Philadelphia, PA, USA. The framework allows users to specify possible knowledge gains
22 from near-term actions and assess the impacts of this learning on subsequent decisions, which
23 enables evaluation of the net benefits of alternative investment plans. In the case study, we con-
24 sider two types of learning: learning to reduce uncertainty and learning to improve performance.
25 This learning can happen through investments or knowledge transfer from experience at other
26 locations. Estimates of cost, performance, and deterioration over time of GI and the prediction of
27 possible knowledge gains are based on the literature and expert opinions. The results propose
28 optimal investment strategies over a 25-year planning horizon and describe tradeoffs between the
29 risk of poor performance and reductions in expected annual stormwater runoff. Finally, by cal-
30 culating differences in expected total costs between non-adaptive, passive adaptive, and active
31 adaptive decision-making, we quantify the economic value of learning and adaptability.

32 **Plain Language Summary**

33 How much can stormwater plans be improved by investing in learning? Is that investment
34 worthwhile? This paper presents an adaptive green infrastructure (GI) investment planning and
35 evaluation framework to address these questions, and an application to the Wingohocking sew-
36 ershed in Philadelphia, PA, USA. The proposed framework includes two components: the eval-
37 uation of GI's capability to reduce stormwater, and the optimization of a portfolio of near-term
38 and subsequent investments. The evaluation quantifies stormwater reduction capacity for GI
39 types at different locations and the uncertainty associated with them, whereas the optimization

40 also considers deterioration in the performance of installed GI and improved GI cost-efficiency
41 of future installations due to improved designs, materials, and installation. Results of modeling
42 experiments show the optimal timing and type of GI investments, the expected stormwater re-
43 duction, and the associated risk levels. The results highlight that assumptions concerning deterio-
44 ration and learning can change which near-term GI investments are optimal. Furthermore, we
45 calculate the value of adaptability as the difference between the cost of non-adaptive (no learning)
46 and passive adaptive (trial-and-error) management, and the value of learning as the difference
47 between passive and active adaptive (active experimentation) management solutions.

48 **1. Introduction**

49 Green infrastructure (GI) or Low Impact Development (LID) utilizes engineering design with
50 soil and vegetation to mimic natural hydrological processes to remove pollutant, detain runoff
51 and harvest rainwater for non-potable uses, which has been considered as a more sustainable
52 solution to urban stormwater pollution comparing to “gray infrastructure” (i.e. conventional
53 engineering solutions, such as underground storage tunnels) (Askarizadeh et al., 2015; Copeland,
54 2014; Dhakal & Chevalier, 2017). GI consists of one or more independent distributed systems,
55 called stormwater management practices (SMPs), which are designed to treat stormwater on-site
56 and could be adjusted for the site conditions. SMPs have evolved into a broad spectrum of de-
57 signs according to the typology of the urban impervious surface (Askarizadeh et al., 2015; Lee et
58 al., 2012). For example, rain gardens and infiltration trenches are often installed to treat runoff
59 from roads and other paved areas; green roofs and rain barrels are for treating rooftop runoff; and
60 permeable pavers can replace impermeable surfaces such as sidewalks, parking lots, and play-
61 grounds.

62 The effectiveness of SMPs can vary with the designs, the spatial alignment, the climate and the
63 underlying watershed characteristics (Avellaneda et al., 2017; Dietz, 2007; Jackisch & Weiler,
64 2017; Jarden et al., 2016; Rossman & Huber, 2016). Studies have pointed out that the flexibility
65 and diversity of SMPs can complicate GI investment planning for their cost and performance un-
66 certainty, and this has become a barrier for GI adoption (Copeland, 2014; Dhakal & Chevalier,
67 2017). For example, in the analysis of Wright et al., (2016), runoff reductions vary from 10% to
68 70% depending on SMPs, while cost per cubic meter of reduction ranges from \$3 to almost
69 \$600 depending on SMPs types and land use.

70 Another layer of complexity comes from the maintenance of GI (Asleson et al., 2009;
71 Avellaneda et al., 2017; Eckart et al., 2017; Freni et al., 2010). Cities sometimes share the
72 maintenance responsibility of GI with the communities and residents by retrofit and community
73 engagement programs (Eckart et al., 2017; Jarden et al., 2016). These programs are desirable for
74 cost savings, but they could also increase the difficulty for assuring maintenance quality and the
75 uncertainty of SMPs’ long-term efficacy. Depending on the quality of the maintenance, SMPs
76 may fail during storms or become less effective over time. The deterioration in SMP
77 performance could result in increasing of disturbance of water uses and violations of regulatory
78 requirements. In contrast, traditional centralized engineering solutions, such as underground
79 storage tunnels, can provide massive storage to reduce peak flow, which are usually operated and
80 maintained by professionals and are less likely to fail during storms or suffer the deterioration in
81 performance over time. As a result, many cities implements GI only at small scale or by pilot
82 projects (US EPA, 2010), even though GI has been proven to be a viable alternative to conven-

83 tional engineering approaches as well as providing ancillary benefits that enhance residents'
84 quality of life (Copeland, 2014; Dietz, 2007; Wise et al., 2010).

85 Adaptive management (AM) is a framework for resolving key uncertainties that has been applied
86 mostly in natural resource and water management (Holling, 1978; Medema et al., 2008; Rist et
87 al., 2013; Williams & Brown, 2014). However, the literature of AM has pointed out that most
88 AM projects failed to evaluate costs and benefits of the monitoring and research plans, which
89 may improve the scientific understanding but not necessarily contribute to decision making
90 (Failing et al., 2004; Williams, 2011; Williams & Johnson, 2015). More recently, researchers
91 have recognized the need to evaluate the value of learning to justify the costs and resources re-
92 quired for monitoring and research actions, and they have applied the concept of value of perfect
93 information (VPI) (also called Expected Value of Perfect Information) to assess the expected
94 improvement resulting from AM (Johnson et al., 2017; Probert et al., 2011; Runge et al., 2011;
95 Williams & Johnson, 2015). Although traditional VPI approach can provide information about
96 the best case learning scenario, the real-world data seldom provides perfect information and the
97 state of nature could also be non-stationary, further complicating the analysis (Runge et al.,
98 2011; Williams & Brown, 2014).

99 The model of this paper extends the simple framework we have previously proposed (Hung &
100 Hobbs, 2019) for adaptive GI investment planning. Unlike the studies mentioned earlier that as-
101 sume perfect information and only look at expected value, the framework can quantify tradeoffs
102 between expected benefit and the risk of undesirable outcomes, test assumptions about what and
103 how we can learn, and assess the value of learning resulting from near-term investments. Alt-
104 hough the framework has these features that stormwater managers desired, its potential is not
105 fully explored in our previous paper other than the expected value-risk tradeoff. This paper
106 shows an application of the framework with the emphasis on how learning works (modeling) and
107 how it can improve the outcome (optimal objective value).

108 Specifically, this paper interviews experts about their thoughts on learning and SMPs' cost un-
109 certainty, evaluates SMP efficacy by hydrological simulation, and assesses the value of learning
110 imperfect information, accounting for non-stationarity of the future states. We compare the opti-
111 mal solution with learning with traditional planning approaches (one-time master planning), with
112 passive adaptive management (plan without considering in learning and the subsequent ability to
113 adapt later, if there is a surprise), and also with active adaptive management (consideration of
114 investments in learning, and how the resulting information can be used to optimally adapt plans).
115 The research questions we have are the following.

- 116 • Where, when, and how much should be invested in what types of SMP?
117 • How does the inclusion of GI performance deterioration and learning processes change near-
118 term and subsequent optimal investments?
119 • Is it worthwhile to invest in learning?

120 Learning is defined as the updating of prior distributions by Bayes' law based on the knowledge
121 gains from implementing the investments. Investments can yield different levels and types of
122 "learning." More specifically, in the case study where we have three subcatchments to account
123 for spatial heterogeneity, we model learning by updating distribution parameters when invest-
124 ments in one location exceed predetermined thresholds or by sharing experience learned at other
125 locations. Theoretically, learning can lead to perfect information with a high investment thresh-

126 old, but it is often not cost-effective if not infeasible for financial and technical issues. Therefore,
127 it is more important to explore learning imperfect information and the tradeoffs between invest in
128 learning and invest in immediate stormwater benefits.

129 Figure 1 shows the conceptual diagram of the proposed framework, which consists of a hydro-
130 logic simulation and an adaptive GI planning model. The adaptive GI investment planning con-
131 siders multiple subcatchments and decisions about siting GIs, budget of the GI program.

132 The remainder of this paper is organized as follows. Section 2 introduces the case study area as
133 well as the hydrologic model used to evaluate SMP performance in reducing stormwater runoff
134 and the associated uncertainty. In Section 3, we summarize assumptions about the dynamics of
135 performance and cost (performance deterioration due to aging, improved characterization of per-
136 formance and cost due to learning, and technological improvements resulting from cumulative
137 investment) and modification of the model formulation in Hung and Hobbs' (2019). Section 4
138 discusses the results of the optimization of adaptive investment strategies and the economic val-
139 ue of learning and adaptability. Finally, Section 5 presents our summary and conclusions .

140 **2. Evaluation of SMP Cost-effectiveness**

141 In this section, we focus on integrating various sources of uncertainty in SMP cost-effectiveness.
142 The purpose of this section is not to develop a calibrated model to evaluate the uncertainty of
143 SMP performance in the study area as the data for calibration and validation are not available,
144 but to demonstrate the use of modeling tools to evaluate the uncertainty using our best
145 knowledge. The improvement in modeling techniques is viewed as a kind of learning in the
146 adaptive planning scheme.

147 **2.1. Evaluation of SMP Performance in Reducing Stormwater**

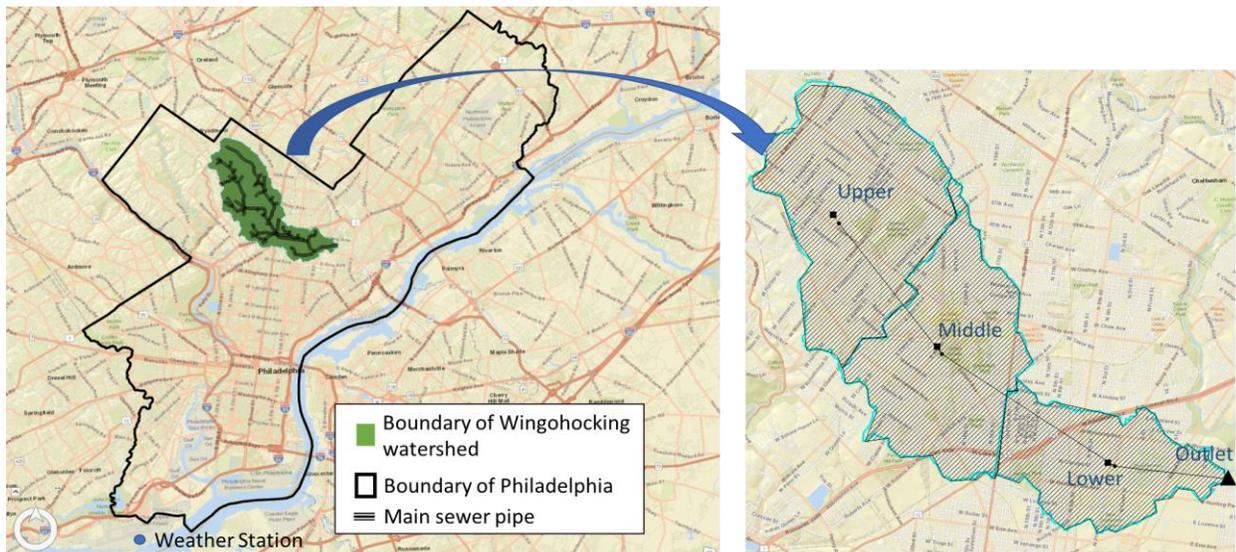
148 Among existing rainfall-runoff simulation products, USEPA's Storm Water Management Model
149 (SWMM) is one of a few models with functions for modeling GI (called LID in SWMM)
150 (www.epa.gov/water-research/storm-water-management-model-swmm). SWMM is widely ap-
151 plied to urban stormwater management studies around the world (Avellaneda et al., 2017; Dong
152 et al., 2017; McGarity, 2013; Palla & Gnecco, 2015; Petrucci & Tassin, 2015; Sebti et al., 2016).
153 Although it is a highly parameterized hydrologic model, its GI module makes it a convenient tool
154 for preparing SMP performance and uncertainty estimates for the adaptive optimization frame-
155 work of Section 3. For simplicity, we only focus on runoff generation processes and SMP re-
156 sponse to precipitation events.

157 For most input parameters for the SWMM simulations, values needed for the simulations cannot
158 be obtained by direct measurement but are instead inferred from observed hydrographs or based
159 on numbers reported in the literature (Bates & Campbell, 2001; Stow et al., 2007). Unfortunate-
160 ly, for our study area, we have neither direct measurements nor runoff hydrographs to calibrate
161 the model, which limits our ability to make precise predictions regarding the dynamic
162 interactions between GI, the urban sewershed, and the climate. However, it is more important to
163 have a consistent assessment of possible SMPs using the best information available, given that
164 the main value of the investment planning models is to screen candidate investments, to identify
165 the most attractive near-term installation, and to understand how assumptions concerning
166 learning, adaptation, and performance changes could impact those recommendations. Therefore,
167 the evaluation is based on the most relevant parameter values we could find in the literature, the
168 GI/LID design guides (Hinman & Wulkan, 2012; Philadelphia Water Department, 2015;

169 Schueler & Claytor, 2009), and consultation with a local expert (a GI design and installation contractor with extensive experience in Philadelphia) (S. Szalay, personal communication, 2018),
 170 There are methods to improve parameterization in the literature (Dong et al., 2017; Muleta et al.,
 171 2013; Sadegh & Vrugt, 2014) if monitoring data are available, but their use is beyond the scope
 172 of this study.
 173

174 2.1.1. Study Area

175 The study area is located at North Philadelphia where the original Wingohocking Creek watershed
 176 was before it was integrated into the sewer system. Figure 2 shows the location and coverage
 177 of the Wingohocking sewershed, which has a total area of 2,076 ha and a length of approximately 10 km,
 178 and the location of the sewer overflow outlet to Frankford Creek.



179

180 **Figure 2.** Boundaries of Philadelphia and the Wingohocking watershed, the main sewer pipes,
 181 and three subcatchments

182 Commonly used modeling approaches tend to involve significant geographical aggregation of
 183 the study area because including all processes in a fine scale would result in complex models
 184 with high data needs and computing requirements (Jefferson et al., 2017). Due to the limited data
 185 available, we simplified the processes by dividing the Wingohocking sewershed into three
 186 subcatchments, called Upper, Middle, and Lower, based on the distance to the overflow outlet
 187 and the layout of the main sewer pipes (Figure 2). For the same reason, the hydrologic model
 188 simplifies the sewer system to only one conveyance channel which collects only stormwater runoff
 189 (i.e., wastewater generation and treatment are not included), under the assumption that runoff
 190 dominates flows in the system in the events of greatest concern to planners. These simplifications
 191 expedite computation and allow us to estimate output uncertainty by simulating a large
 192 number of scenarios based on random samples of model inputs and parameters from assumed
 193 distributions.

194 2.1.2. Watershed Characteristics

195 The Wingohocking watershed is a highly developed area (100% developed). Land uses in the
 196 Upper subcatchment are mostly residential, while the Middle and Lower subcatchments are pri-

197 marily residential mixed with commercial and industrial uses (Table 1). The Lower and Middle
 198 also have higher vacant properties (6% and 7%, respectively) than the Upper (3%). The impervi-
 199 ousness is about 55% and is fairly uniformly distributed across Wingohocking sewershed. The
 200 average slope is about 6% and is slightly lower in the Lower sewershed (4.7%), where the origi-
 201 nal Wingohocking Creek transitioned from the Piedmont Region to the Atlantic Coastal Plain.
 202 The average catchment widths of the subcatchments are estimated on the digital map by the
 203 ArcGIS measure tool. The characteristics of the three subcatchments calculated are summarized
 204 in Table 1, and data for the calculation are from publicly available geospatial data, of which the
 205 sources are listed in the Appendix.

206 **Table 1.**
 207 *Subcatchment Characteristics and Land Use*

	Slope (%)	Imper- viousness (%)	Area (ha)	Width (m)	Land use (%)			
					Residential	Commercial	Industrial	Others
Upper	6.2	53.9	850	3,800	50%	11%	5%	34%
Middle	6.5	56	733	3,800	37%	17%	7%	39%
Lower	4.7	55.6	493	3,000	23%	10%	19%	53%

208 The soil in Wingohocking sewershed is mostly (>95%) categorized as urban land in USDA's soil
 209 survey, meaning that the soil is largely covered by impervious materials and its permeability is
 210 unknown. However, the soil on Atlantic Coastal Plain (the physiographic region where most of
 211 Philadelphia is located) generally is highly permeable (Markewich et al., 1990). Infiltration of
 212 permeable surfaces is modeled by the Green-Ampt method (Chow et al., 1988) in our SWMM
 213 simulation, where the saturated infiltration rate is chosen based on loam (25 mm/hr). The de-
 214 pression storage for the pervious and impervious areas are 4 mm and 1.3 mm, respectively.

215 The conveyance channel collecting stormwater is modeled as a 9 m-wide by 3 m-high rectangu-
 216 lar channel with a length of 7.8 km and a Manning roughness coefficient of 0.013. For other pa-
 217 rameters for SWMM modeling not discussed here, we use default values from the SWMM user
 218 manuals (Rossman, 2015; Rossman & Huber, 2016).

219 2.1.3. SMP design parameters

220 The five distinct types of SMPs evaluated in this paper include the following. *Rain gardens*
 221 (RGs) are vegetated SMPs that detain stormwater to infiltrate and recharge groundwater. *Infil-*
 222 *tration trenches* (ITs) and *permeable pavement* (PP) are both non-vegetated infiltration SMPs.
 223 ITs are often installed for retaining stormwater from transportation right-of-ways, while PP holds
 224 stormwater directly on its surface and increases permeability to the underlying soil, relative to
 225 the impermeable pavement. RG, IT, and PP are characterized as infiltration SMPs because they
 226 divert a portion of surface runoff to soil moisture and groundwater. *Green roofs* (GRs) and *rain*
 227 *barrels* (RBs) can retain or reduce stormwater from rooftops and, therefore, are characterized as
 228 roof SMPs. However, a GR utilizes the evapotranspiration to deplete water after storms, while
 229 an RB is generally designed to delay but not reduce discharges, unless it drains primarily onto
 230 permeable surfaces.

231 We assume that all SMP installations are in parallel so that the stormwater reductions are addi-
 232 tive if multiple installations are made. In actual GI plans, some SMPs may be installed down-
 233 stream of others (e.g., rain gardens downslope from permeable pavement); we assume that such

234 installations in series are the exception rather than the rule. By the assumption of additivity, the
 235 SMP installations can be combined into single synthetic SMP to represent the overall perfor-
 236 mance. Then we can apply the hydrologic model to assess the performance of the SMPs, in
 237 which the parameter values are assumed uncertain due to the variability of the designs, and the
 238 quality of the installations. The design parameters of SMPs and the assumptions regarding their
 239 uncertainty are presented in Supporting Information (Text S1).

240 2.1.4. Evaluation of SMP Efficacy and Its Uncertainty

241 We apply Monte Carlo method coupled with SWMM to estimate distributions of the stormwater
 242 reduction for each SMP type. Recent studies have applied the same method to evaluate
 243 SWMM's prediction uncertainty by comparing monitoring data with the modeling results
 244 (Avellaneda et al., 2017) and SMP efficacy uncertainty under future climate and land use change
 245 uncertainty (Dong et al., 2017). Our work is distinguished from these analyses in that our esti-
 246 mation focuses on the overall performance uncertainty of individual SMP type at watershed scale
 247 concerning the design variability and installation quality among the installations., whereas their
 248 analyses emphasis on the uncertainty from model selection and calibration contributed to SMP
 249 performance. Although our analysis does not include the model and calibration uncertainty due
 250 to lack of data, the results can still provide a basis for comparison to inform investment decision
 251 making. Model uncertainty can be incorporated in our framework when sufficient data become
 252 available, but it is beyond the scope of this paper.

253 We chose reduced stormwater volume as the performance metric for simplicity and because we
 254 lack sufficient information to model the CSO generation process (e.g., data of sewer network and
 255 characteristics, capacities of wastewater treatment plants and their operation rules). In addition,
 256 we use the same parameter distributions for SMP efficacy in all three subcatchments due to our
 257 limited knowledge of site conditions in the subcatchments. These assumptions are not a limita-
 258 tion of our framework for the following reasons: lack of data is common in stormwater manage-
 259 ment for the high costs of monitoring and the missing sewer system information. Moreover, our
 260 framework can accommodate different metrics and differentiated assumptions for each
 261 subcatchment, which could, for example, be implemented after the initial (first stage) invest-
 262 ments are made.

263 The Monte Carlo simulation is described as follows. In order to quantify the mean and standard
 264 deviation (SD) of stormwater reductions for each SMP, we draw one sample from each of the
 265 parameter distributions, simulate the resulting annual stormwater reduction with one-year precip-
 266 itation data (from Jan. 1 to Dec 31), and repeat this process 30 times for each year from 1980 to
 267 2013 using SWMM. The precipitation data applied in this analysis are presented in Supporting
 268 Information (Text S2).

269 The simulation assumes that a 200 ha impervious area is treated by the SMP and that the random
 270 parameters are statistically independent. The simulation results are summarized in Table 2,
 271 where the unit is in m/yr (m^3 stormwater reduction per m^2 SMP per year). The coefficient of var-
 272 iation (CV) is a measure of dispersion (risk) relative to the mean.

273 **Table 2**
 274 *Statistical Summary of SMP Annual Stormwater Reductions*

SMP	Upper			Middle		Lower	
	Mean	SD	CV ($\frac{\mu}{\sigma}$)	Mean	SD	Mean	SD

	(m/yr)	(m/yr)		(m/yr)	(m/yr)	(m/yr)	(m/yr)
Rain Garden (RG)	14.71	2.27	0.15	14.75	2.22	14.57	2.20
Infil. Trench (IT)	19.15	3.21	0.17	19.01	3.09	19.08	3.07
Permeable Pavement (PP)	2.03	0.49	0.24	2.04	0.48	2.06	0.49
Rain Barrel (RB)	33.36	3.08	0.09	33.27	2.95	33.41	3.09
Green Roof (GR)	0.66	0.08	0.12	0.67	0.08	0.62	0.07

275 From Table 2, we can see that the SMPs' performances are only slightly different in the three
 276 subcatchments, which is the result of watershed geometry and rainfall pattern. Generally
 277 speaking, SMPs performs at locations of which the treated area collects water faster, if no
 278 overflows (SMPs's storage are not full during the storms) and at locations of which the treated
 279 area collects water slower, if SMPs overflow. The former is the case of RG and IT in the Middle
 280 subcatchment and the latter is the case of PP and RB in the Lower subcatchment. Since the
 281 effects only happen at site scale, the difference of the means and standard deviations of the SMPs'
 282 performance in the three catchments should not be viewed as real differences in distribution but
 283 rather the errors from our modeling assumption. Therefore, we assume that the SMP efficacy
 284 distributions in the Middle and Lower subcatchments are the same as the results in the Upper
 285 Subcatchment.

286 Moreover, we can see that IT can provide the highest expected annual stormwater reduction per
 287 m²-installation, but it is also relatively risky, as its CV being the second-highest among the SMPs.
 288 PP is the riskiest (the highest CV) and the second lowest in stormwater reduction per m²-
 289 installation. For the roof SMPs, RB is better than GR for the higher stormwater reduction per
 290 m²-installation and lower uncertainty (lower CV).

291 2.2. Evaluation of Cost Uncertainty

292 The previous section described uncertainty in SMP performance in terms of annual water storage
 293 per unit area (m/yr). This section describes the derivation of uncertainty in SMP cost (\$/yr/m²).
 294 The ratio of performance to cost (in m³/\$) and its uncertainty can then be derived.

295 Our cost analysis is based on the cost information summarized in the Center for Neighborhood
 296 Technology's "Green Values – National Stormwater Management Calculator"
 297 (greenvalues.cnt.org/national/cost_detail.php, accessed June 2018), including capital and
 298 maintenance costs, installation lifespans, and cost uncertainty ranges of the SMPs. Because this
 299 cost information came from various sources, the values are adjusted to accommodate the local
 300 conditions based on the discussion with our Philadelphia expert (S. Szalay, personal communica-
 301 tion, 2018).

302 The cost uncertainty needed for basin planning is the uncertainty in the average cost of many in-
 303 stallations rather than the cost variability of one installation. So instead of using the cost range
 304 summary from the calculator, we adjust those ranges in order to obtain plausible ranges of aver-
 305 age cost based on two criteria:

- 306 1. the construction complexity of the SMP; and
- 307 2. the portion of the installation cost that serves the purpose of stormwater control. There are
 308 other costs that are sometimes incurred (such as the cost of landscaping) to enhance GI's so-
 309 cial benefits; these are highly variable and are not directly related to the purpose of the instal-
 310 lation, and so for a basin-level analysis are not included.

311 Table 3 shows our estimates of the five SMPs' capital costs, including their lifespans and the
 312 ranges of the annualized capital costs, and the annual maintenance costs. The values of the aver-
 313 age capital and maintenance costs are chosen based on the ranges in CNT's cost summary and
 314 expert judgment (S. Szalay, personal communication, 2018). The annualized costs are calculated
 315 by assuming a 5% interest rate and their lifespans. We can see that RB has the highest annual-
 316 ized capital cost per m^2 -installation because the surface area of RB provides a 91cm (3 ft) storage
 317 depth and the cost estimation is based on the aggregation of rain barrels, each with 227 liters
 318 storage (60 gallons). With economies of scale, the cost could be lower, which is not directly
 319 modeled as a function of investment size; however, such effects are at least partially captured by
 320 our modeling of learning as a function of total investment since learning can reduce SMP costs.

321 **Table 3**
 322 *SMP Capital and Maintenance Costs and Uncertainty Ranges*

SMP	Avg. Capital Cost (\$/m ²)	Life Span (yr)	Assumed Annualized Capital Cost Range			Assumed Annual Maintenance Cost Range		
			Lower Bound (\$/m ² /yr)	Mean (\$/m ² /yr)	Upper Bound (\$/m ² /yr)	Lower Bound (\$/m ² /yr)	Mean (\$/m ² /yr)	Upper Bound (\$/m ² /yr)
RG	100	30	15.1	21.6	28.1	3.5	5	6.5
IT	150	20	13.9	19.9	25.9	5.6	8	10.4
PP	80	15	6.7	8.3	10	3.2	4	4.8
RB	200	20	25.2	26.5	27.9	1	1	1.1
GR	80	30	13.8	17.3	20.7	3.2	4	4.8

323
 324 The SMPs' annual maintenance cost ranges (Table 3) are assigned based on the frequency and
 325 complexity of maintenance. For example, the infiltration SMPs (RG, IT, and PP) require fre-
 326 quent inspection to avoiding clogging (Avellaneda et al., 2017; Fletcher et al., 2013), while RB
 327 can be inspected less often and only rarely needs replacement (US EPA, 2013). However,
 328 maintenance of the RB often requires manual emptying of storage after storms (other designs
 329 may come with a drain hose at the bottom and discharge slowly). We assume that the attention
 330 paid to RB maintenance and emptying would affect rain barrels' capacity to store stormwater,
 331 which adds to the uncertainty in performance.

332 The annualized costs per installation area in Table 4 combine the annualized capital costs and
 333 maintenance costs under the assumption that the two costs are perfectly correlated (unexpectedly
 334 high capital costs are likely to be accompanied by higher maintenance costs as well). Table 4 al-
 335 so shows the annualized cost per treated area (the annualized costs per installation area divided
 336 by the average drainage area ratio), which are the costs for treating the impervious surface. It ap-
 337 pears that the annualized cost per treated area is mostly determined by the SMP's treated area to
 338 surface area ratio (drainage area ratio). RG and PP are expensive because of their low treatment
 339 ratios, whereas RB becomes the cheapest SMP because of its high treatment ratio even though
 340 they are the most expensive SMP per m^2 -installation.

341 **Table 4**
 342 *SMP Annualized Costs per m^2 -installation and per m^2 -treated*

SMP	Annualized Cost per Installation Area			Avg. Drainage	Annualized Cost per Treated Area		
	Lower	Mean	Upper		Lower	Mean	Upper Bound

	Bound (\$/m ² /yr)	Bound (\$/m ² /yr)	Bound (\$/m ² /yr)	area ratio	Bound (\$/m ² /yr)	Bound (\$/m ² /yr)	Bound (\$/m ² /yr)
RG	18.6	26.6	34.6	24	0.78	1.11	1.44
IT	19.5	27.9	36.3	30	0.65	0.93	1.21
PP	9.9	12.3	14.8	1.5	6.57	8.21	9.85
RB	26.2	27.5	28.9	108	0.24	0.25	0.27
GR	17.0	21.3	25.5	1	17.03	21.29	25.55

343

344 We calculate the distribution of each SMP's cost-effectiveness using Monte Carlo simulation by
 345 drawing samples from the annualized cost (per treated area) distributions (assumed to be uniform
 346 between the lower and upper bounds of Table 4) and from the results of SWMM performance
 347 simulations (Table 2). Costs and stormwater reductions are assumed to be independent. In
 348 addition, we assumed that the costs in the Middle and Lower subcatchments are 5% and 10%
 349 less than the values shown in Table 4 because of the lower property values in those areas, which
 350 provides an incentive to invest in the Lower and Middle over the Upper subcatchments. Table 5
 351 shows the statistical summary of the SMP's cost-effectiveness in m³/\$ of the subcatchments,
 352 where IT is the most cost-effective SMP while GR is the least. The values of CV indicate that
 353 PP is the most uncertain SMP and RB is the least risky one. The CV values of the SMPs are the
 354 same for the subcatchments because the cost adjustment would cancel out.

355 **Table 5**

356 *SMP Cost-effectiveness and Uncertainty per Unit Installation Areas by Subcatchment (based on*
 357 *Tables 2 and 4)*

SMP	Upper			Middle		Lower		Stormwater Reduction (m ³ /ha)	
	Mean (m ³ /yr)	SD (m ³ /yr)	CV	Mean (m ³ /yr)	SD (m ³ /yr)	Mean (m ³ /yr)	SD (m ³ /yr)	Per ha of Drainage Area	Per ha of Instal- lation
RG	0.437	0.106	0.243	0.460	0.112	0.486	0.118	4,851	147,090
IT	0.537	0.127	0.236	0.565	0.134	0.597	0.141	4,994	191,523
PP	0.127	0.036	0.283	0.134	0.038	0.141	0.040	10,427	20,258
RB	0.927	0.087	0.094	0.976	0.092	1.030	0.097	2,318	333,581
GR	0.024	0.004	0.167	0.025	0.004	0.027	0.004	5,110	6,620

358

359 From the values of the SMP cost-effectiveness per ha installation (next to the last column in Ta-
 360 ble 5), we can see that a low-cost investment strategy would be to install IT for impervious
 361 ground surfaces and RB for impervious roofs. However, the relatively low amount of storm-
 362 water reduction per ha of drainage area for IT and RB means that if a large amount of reduction
 363 is desired, these measures will not be as effective as other approaches (namely PP and GR, re-
 364 spectively). For instance, for a given roof drainage area, GRs will reduce more stormwater than
 365 RBs (as indicated by the last column of Table 5), even though RBs are more cost-effective (in
 366 m³/yr).

371 **3. Modeling GI Investment Planning and Learning**

372 Our GI investment planning model is an extension of the basic method proposed by Hung and
 373 Hobbs (2019). The basic approach accounts for the following characteristics of the GI decision
 374 framework:

- 375 • optimization of stormwater reduction over multiple years subject to monetary budget limita-
 376 tions and the acceptable risk level specified by users;
- 377 • uncertainty in SMP performance and cost that changes from one investment decision stage to
 378 the next as a result of learning and investment; and
- 379 • adaptive investment planning framework, including multiple investment decision stages (e.g.,
 380 years 0 and 5) and consideration of the value of learning for improving decisions in later
 381 stages.

382 The method is structured as a two-stage stochastic program with recourse, in which uncertainties
 383 are characterized by discrete scenarios with probabilities. The uncertain coefficients are the
 384 SMP cost-effectiveness, which follow some distributions representing the current understanding
 385 of the SMPs cost-effectiveness. Based on the learning assumptions, the distribution parameters
 386 would be updated if the first-stage investments exceed the learning thresholds. For example, the
 387 multi-stage learning model with technology improvement assumes that, if learning criteria are
 388 met, the mean values of the cost-effectiveness distributions will increase by $\gamma\%$, and the standard
 389 deviation would be reduced by $\beta\%$. Our GI investment planning model is developed based on
 390 this variant.

391 The differences of our method compared to the original method presented in Hung and Hobbs
 392 (2019) are as follows. The highly simplified hypothetical example presented in that reference
 393 considers a single subcatchment, disregards the possibility of deterioration over time of perfor-
 394 mance of existing installations, and does not consider that learning about SMPs at one location is
 395 transferable to SMPs at other locations. In this paper, we incorporate these considerations into
 our model by introducing new random variables for deterioration and constraints for modeling
 knowledge transfer. We also test the system based on SWMM hydrological simulations and cost
 estimates for an actual case study, the Wingohocking watershed, rather than a simple hypothet-
 ical illustration.

396 **3.1. Assumptions and Problem Settings**

397 Following the Philadelphia's Green City Clean Waters (GCCW) program, the total planning
 398 horizon is set to 25 years where the first stage starts at year 0 and the second stage starts at year 5.
 399 Although the planning and review cycle of the GCCW program is every 5 years, which means 5
 400 planning stages in total, we only model the first stage decisions (the near-term decisions) and
 401 combine the rest into the second decision stage (the long-term decisions).

402 Therefore, the objective of the planning model is to maximize the expected stormwater reduction
 403 over a 25-year time horizon by making investments in the SMPs in the subcatchments while
 404 considering constraints representing the total budget (which limits investment amounts),
 405 impervious roof and ground area of each subcatchment (which limit installation opportunities),
 406 risks of realizing a low stormwater reduction (in the form of a conditional value of risk,CVAR),
 407 and learning (which use Bayes Law to update distributions of the SMP cost-effectiveness). This
 408 is essentially a multi-objective optimization where the objectives are (1) maximizing expectation

409 of the stormwater reduction over the horizon, (2) maximizing CVAR (defined as the conditional
 410 expectation of the stormwater reduction for the poorest 10% of realizations), or (3) minimizing
 411 cost. We use the epsilon constraint method of multi-objective optimization, in which the prob-
 412 lem is solved repeatedly as single-objective optimization while treating the other objectives as
 413 constraints with various user-specified upper bounds (for maximization objectives) or lower
 414 bounds (for minimization objectives) (Deb, 2014), to search for the Pareto-optimal solutions. By
 415 applying the epsilon constraint method, we can describe tradeoffs among these three objectives.

416 Uncertainty is characterized using the magnitude of the standard deviation of a distribution, and
 417 we assume that learning would cause a decrease in the standard deviation. In addition, if learn-
 418 ing contributes to technological progress, it could also lead to an increase in the expected value
 419 in the SMPs' cost-effectiveness. More details of the learning assumptions are presented below.

420 3.1.1. Basic Learning (BL) and Advanced Learning (AL)

421 In this case study, we assume that the learning for an SMP has two levels, which are triggered if
 422 cumulative investment in that SMP in the first (year 0) decision stage exceeds predefined learn-
 423 ing thresholds. The lower level of learning, basic learning (BL), has a lower investment threshold
 424 and can result in the second stage (here, year 5) realizing both (1) a reduction in uncertainty con-
 425 cerning the SMP's cost-effectiveness (standard deviation) and (2) technological progress, re-
 426 flected in an increase in an expected SMP's cost-effectiveness over what it would have been oth-
 427 erwise. Meanwhile, the higher level of learning, advanced learning (AL), has a higher investment
 428 threshold and can provide a greater boost in the cost-effectiveness of second stage SMP invest-
 429 ment and a larger uncertainty reduction. On the other hand, if the first stage investment in an
 430 SMP does not reach the BL thresholds, the cost-effectiveness distribution would remain un-
 431 changed in the second stage, called no learning (NL).

432 Based on consultations with our local expert (S. Szalay, personal communication, 2018), the as-
 433 sumptions concerning learning thresholds for each of the SMPs and the corresponding changes
 434 in the SMPs' cost-effectiveness distributions in the second stage are summarized in Table 6. Re-
 435 finement of the learning thresholds and changes in distributions could be a subject for future
 436 study based on consultation with multiple experts or statistical analysis of actual experience.

437 **Table 6**

438 *Assumptions about Changes in Mean and Standard Deviation (SD) of SMP Cost-effectiveness*
 439 *Realized in the Second Stage (year 5) as a Result of Learning, and the Investment Thresholds for*
 440 *Learning*

SMP	Basic Learning			Advanced Learning		
	Mean ad- justment (γ_{BL})	SD ad- justment (β_{BL})	Threshold (\$K)	Mean adjust- ment (γ_{AL})	SD adjust- ment (β_{AL})	Threshold (\$K)
RG	+10 %	-30 %	2,200	+30 %	-50 %	10,000
IT	+10 %	-30 %	1,000	+20 %	-50 %	4,000
PP	+10 %	-30 %	160	+20 %	-50 %	350
RB	+5 %	-30 %	30	+10 %	-50 %	60
GR	+10 %	-30 %	350	+30 %	-50 %	700

441 3.1.2. Partially Transferable Learning

442 When learning happens, the knowledge gains from one SMP in a subcatchment may be partially
 443 applicable to the other subcatchments. For example, if we learned about the rain gardens in the
 444 Lower subcatchment, we may expect the rain gardens in the Middle and Upper subcatchment to
 445 perform similarly. However, we may not know the costs of the SMPs in the Middle and Upper
 446 subcatchment if we did not invest there. Therefore, we assume the learning is only partially
 447 transferrable across locations. That is, if the investment in one SMP triggers AL in one sub-
 448 catchment, the same SMP in the other subcatchment would have BL even though the invest-
 449 ments there are below the BL investment thresholds.

450 Similarly, the knowledge transfer could happen between different types of SMPs in a subcatch-
 451 ment. For example, RG and IT are both infiltration practices treating stormwater runoff from the
 452 ground impervious area, so learning about one of them may also teach us about the other. It is
 453 not difficult to model both types of learning, but in this case study, we only model the former
 454 (learning transfers between subcatchments) for simplicity.

455 3.1.3. Performance deterioration

456 Studies have shown that the performance of installations of infiltration SMPs (RG, IT, and PP)
 457 would deteriorate over time due to clogging, with the rate depending on the inflow water quality,
 458 pretreatment, and maintenance (Bergman et al., 2011; Drake & Bradford, 2013). Deterioration
 459 may also happen to roof practices (RB and GR). Unfortunately, data to support estimates of dete-
 460 rioration rates is unavailable. Nonetheless, we believe that it is important to consider the deterio-
 461 ration process in a long-term planning problem, in part because practitioners believe that differ-
 462 ent SMPs are likely to experience different rates of deterioration. To model deterioration, we
 463 assume that the stormwater reduction (and thus cost-effectiveness) of an SMP installed in the
 464 first stage would, on average, decrease by a fraction, D_I , on average over the 5 years (denoted
 465 T_I) between the first and second decision stages, and by D_{II} over the 20 years (denoted T_{II}) fol-
 466 lowing the second stage. For second stage installations, we assume that the deterioration rate is
 467 also D_{II} . This simplification could result in a slight disadvantage for second stage installations,
 468 but sensitivity tests showed that the exact value used for those installations did not appreciably
 469 affect the solutions.

470 The local GI expert gave quantitative judgments of the distributions of the values of D_I and D_{II}
 471 shown in Table 7; we assume that uniform distributions with the means and ranges shown in Ta-
 472 ble 7. These distributions represent parameter uncertainty concerning the mean over many in-
 473 stallations over the time horizon, and not the variability among individual facilities, which would
 474 be expected to be much greater.

475 **Table 7**

476 *Assumptions Concerning Performance Deterioration of Stage I and Stage II Installations*

SMP	Stage I			Stage II		
	Lower Bound	Mean	Upper Bound	Lower Bound	Mean	Upper Bound
RG	0.9	0.95	1	0.7	0.8	0.9
IT	0.9	0.95	1	0.6	0.75	0.9
PP	0.7	0.85	1	0.5	0.6	0.7
RB	0.9	0.95	1	0.7	0.8	0.9

GR	0.95	0.975	1	0.9	0.93	0.95
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477 3.1.4. Prior and Posterior Distributions of SMP Cost-effectiveness

478 The first stage cost-effectiveness values (denoted as vector C_I in $\text{m}^3/\$$) have a prior probability
479 distribution that represents our current understanding of the SMPs' cost-effectiveness; if no
480 learning occurs, then that same distribution applies to the second stage cost-effectiveness, inde-
481 pendent of what C_I occurred (realized). But if learning occurs, then the prior distribution would
482 be updated to a posterior distribution of C_{II} . For the mathematical formulation, we generate sce-
483 narios to represent the uncertainty, as explained in Section 3.1.5. The updated distribution in sce-
484 nario s is the posterior distribution, conditioned on the scenario s , which is denoted C_{IINS} , C_{IIBS} ,
485 or C_{IIAS} , if no learning, BL, or AL happens, respectively. Depending on which s occurs, the pos-
486 terior mean cost-effectiveness may be high or low; when this deviates from the prior expected
487 value, this indicates that learning has occurred that indicates that the performance is either better
488 or worse than what was originally expected.

489 For ease of computation, when learning occurs, the distribution of the posterior expected values,
490 across s , is assumed to be adequately approximated by a normal distribution. (If no learning oc-
491 curs, then the posterior expected value is just the prior expected value.) For the cases of BL and
492 AL, this distribution may deviate from the prior if it is also assumed that there is a technological
493 improvement as a result of learning, in which case, the mean cost-effectiveness is increased. Ta-
494 ble 6 shows that the mean adjustments are 10% (γ_{BL}) for BL and 30% (γ_{AL}) for AL for most
495 technologies, except for the simpler technologies of rain barrels and infiltration trenches, where
496 less improvement is expected.

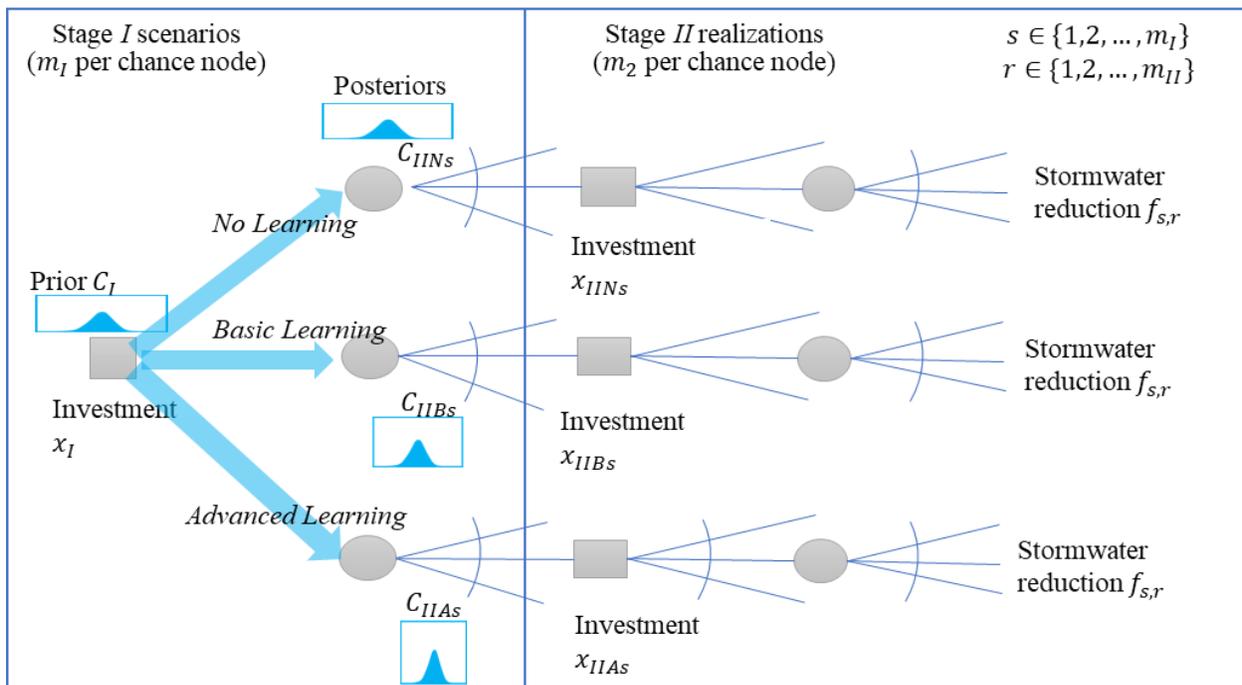
497 Each posterior has a reduced standard deviation equal to $((1 + \beta)\sigma)$, where β is the uncertainty
498 adjustment (negative values; either β_{BL} or β_{AL} in Table 6, depending on the amount of invest-
499 ment), and σ is the standard deviation of the prior distribution. The standard deviation of the ex-
500 pected posterior value is $\sqrt{(1 - (1 + \beta)^2)}\sigma$, which can be derived from the Law of Total Vari-
501 ance (Hung and Hobbs, 2019)

502 To represent the prior distributions, we simply use the sets of Monte Carlo samples ("sampling
503 distribution") generated for the cost-effectiveness assessment of the SMPs in Section 2, based on
504 the assumed prior means and standard deviations (Table 5). The samples are saved and reused in
505 the scenario generation process. By recycling the samples, we need to make neither assumptions
506 about the distributions of the SMPs' cost-effectiveness nor to perform distribution fitting. The
507 posterior distributions are also sampling distributions where the means and standard deviations
508 are adjusted based on the learning assumptions, which are explained next.

509 3.1.5. Scenario Generation Procedure

510 Our mathematical formulation below (Section 3.2.) requires m_I scenarios of SMP cost-
511 effectiveness in the first stage and, for each first stage scenario, m_{II} scenarios of second stage
512 cost-effectiveness for each of the learning cases (NL, BL, and AL), where m_I and m_{II} are user-
513 specified. Figure 3 provides a decision tree to illustrate the structure of the model and the scenar-
514 ios, where the squares are the decision nodes, the circles are the chance nodes, the lines emitting
515 from chance nodes are the scenarios, and the end of each branch is an outcome ($f_{s,r}$, $s \in m_I$ and
516 $r \in m_{II}$ are the indices of the first and second stage scenarios, respectively). At this point, let us
517 simplify the explanation by imagining that the decision tree represents just the investment deci-

518 sion for one SMP on one subcatchment. First, a decision is made about the amount of Stage I invest-
 519 ment (year 0, left-most node). The state of knowledge at that point in time about that SMP's
 520 effectiveness is represented by the prior distribution. Following the investment, then only one of
 521 the three learning cases can happen for that SMP; if investment in the first stage is high enough,
 522 then either BL or AL can occur, otherwise NL occurs. The first chance node then represents
 523 possible realizations (m_I distinct scenarios s) of C_I for that SMP as well as what is learned (i.e.,
 524 whether one learns that the SMP's effectiveness per \$ is likely to be higher or lower than the pri-
 525 or expected value). The decision tree shows that the next thing to happen is that a Stage II deci-
 526 sion about investment in that SMP (year 6 in our case study, represented by the second set of de-
 527 cision nodes). The state of knowledge at that point in time is the posterior distribution of effec-
 528 tiveness; if learning takes place then the variance of possible outcomes is narrowed relative to
 529 the prior, and depending on which scenario s occurred (i.e., what exactly is learned), the ex-
 530 pected value changes relative to the prior. Following that investment, the actual effectiveness
 531 $f_{s,r}$ of the SMP is realized, depending on which of the m_{II} second stage scenarios r occurs. At
 532 this point, we know what decisions have been made in the two stages and the resulting precise
 533 impact on stormwater.



534

535 **Figure 3.** Decision tree representation of the prior, posteriors, and scenarios, in the adaptive GI
 536 investment planning model

537 In contrast to the decision process just described, the optimization model actually considers five
 538 SMPs and three subcatchments. Consequently, the investment decisions at each stage are a vec-
 539 tor length 15, one element per SMP and catchment. Similarly, a scenario of C_I or C_{II} for a par-
 540 ticular stage also consists of a vector of length 15 (15 values of cost-effectiveness, one for each
 541 combination of SMP technology and subcatchment). The probabilities of these scenarios s and r
 542 reflect any assumptions about joint distributions across SMPs and subcatchments. Posterior joint
 543 distributions reflect the learning that occurs for each SMP, based on the first stage investments,
 544 as well as assumptions about how learning about one SMP affects learning about others. The

545 total number of paths in the scenario tree equals $m_I * m_{II}$ multiplying by the combinations of
 546 learning outcomes (3^{15}). The first stage scenario is indexed by $s \in S_I$ and the second stage sce-
 547 nario is indexed by $r \in S_{II}$. The mathematical formulation in Section 3.2 is more complicated for
 548 the risk and physical constraints and the consideration of transferable learnings.

549 The scenario generation procedure is as follows. For each first-stage scenario s , SMP i , and sub-
 550 catchment j :

- 551 • Generate one cost-effectiveness realization for each SMP from the prior distributions
 552 (resample from the sampling distributions), denoted by the vector C_{IS} , of which the elements
 553 are the cost-effectiveness realizations of SMP i and subcatchment j , denoted $C_{IS(i,j)}$
- 554 • Generate one sample (vector) from the expected value distributions of BL and AL each, de-
 555 noted C_{IBS} and C_{IAS} , respectively. This is assumed independent to the C_{IS} realization. This
 556 assumption is based on the idea that learning of the expected value may not necessarily come
 557 from just data. It could be the information extracted from modeling and monitoring data, per-
 558 sonal experience or something else. For example, the stormwater manager may learn that the
 559 poor performance in the first-stage installations was the results of some design flaws, which
 560 could be corrected easily to improve the performance, or that the high performance in the
 561 first stage installations was due to siting on the best sites and would expect a decline in aver-
 562 age performance with more installation in the second stage.
- 563 • Generate m_{II} samples from the prior for no learning case, denoted $C_{IINS,r}$
- 564 • Generate m_{II} samples from the posterior distribution of the BL case (denoted $C_{IIBS,r}$), where
 565 the mean is equal to C_{IBS} and the standard deviation is equal to $\sqrt{(1 - (1 + \beta_{BL})^2)\sigma}$
- 566 • Generate m_{II} samples from the posterior distribution of the AL case (denoted $C_{IIAS,r}$), where
 567 the mean is equal to C_{IAS} and the standard deviation is equal to $\sqrt{(1 - (1 + \beta_{AL})^2)\sigma}$
- 568 • This sampling procedure does not need to make assumptions about the posterior distributions
 569 but simply resample and adjust the mean and variance based on the learnings. This allows the
 570 user to work with empirical distributions and data for the distributions in real-world cases
 571 that are often difficult to characterize.

572 3.2. Mathematical Formulation

573 In this section, we present only our modification of the original formulation in the following or-
 574 der: decisions, objectives, learning constraints, risk constraints, and physical constraints. Please
 575 refer to Hung and Hobbs (2019) for the complete formulation.

576 3.2.1. The Decisions

577 The decisions are the annualized investments (\$/yr) in the SMPs (denoted by x_I for the invest-
 578 ment decisions in the first stage and x_{II} for the decisions in the second stage). x_I and x_{II} are de-
 579 cision vectors that contain 15 elements representing the investment in the five SMPs at the three
 580 locations. The investment in SMP i at subcatchment j in the first stage is denoted $x_{I,(i,j)}$ and the
 581 investment in the second stage, scenario s , is denoted $x_{IIS,(i,j)}$. The sets of SMPs and the sub-
 582 catchments are denoted as $SMP = \{RG, IT, PP, RB, GR\}$ and $Sub = \{Upper, Middle, Lower\}$.

583 The second-stage investment decisions (x_{IIs}) in scenario s consist of the decisions under three
 584 learning cases: no learning (x_{IINs}), basic learning (x_{IIBs}), and advanced learning (x_{IIAs}). That is,
 585 $x_{IIs} = (x_{IINs}, x_{IIBs}, x_{IIAs})$. The investments are assumed non-negative.

$$586 \quad x_I, x_{IINs}, x_{IIBs}, x_{IIAs} \geq 0, \forall s \in S_I \quad (1)$$

587 One caveat is that if investments are made in the first stage, some SMPs may reach their service
 588 life before the end of the planning horizon. We assume that the replacement would be installed
 589 immediately at the end of an SMP's service life and the average cost-effectiveness would remain
 590 unchanged throughout the planning horizon. This allows us to model the investment problems as
 591 a two-stage programming (now and later) to avoid the curse of dimensionality. The replacement
 592 problem can be a future direction, but it is beyond the scope of this paper.

593 3.2.2. The Objective – Maximizing Expected Annual Stormwater Reduction

594 The objective (f) is the expected annual stormwater reduction as shown in Eq. 2, where $E[C_I]$ is
 595 the vector of the expected values of the SMPs' annual stormwater reduction per \$ investment
 596 based on our current understanding (i.e., the prior); C_{IBs} and C_{IAs} are the vectors of our predic-
 597 tion on the SMPs' expected performance for basic learning and advanced learning, respectively
 598 (i.e., the posteriors); D_I and D_{II} are fractions representing the average loss of the SMPs' capacity
 599 in stormwater reduction in the first stage and the second stage, respectively; and T_I and T_{II} are the
 600 time horizons of the first and second stages, respectively.

$$601 \quad Max f = \frac{T_I}{T_I+T_{II}} E[D_I C_I] x_I + \frac{T_{II}}{T_I+T_{II}} (E[D_I D_{II} C_I] x_I + E[D_{II} (C_I x_{IINs} + C_{IBs} x_{IIBs} + C_{IAs} x_{IIAs})]) \quad (2)$$

602 Multiplying the objective by the planning horizon ($\frac{1}{T_I+T_{II}}$), we can get the annual average storm-
 603 water reduction over that time period. That is, the objective is essentially to maximize the total
 604 stormwater reduction over the planning horizon. As a result, deferring investments to the second
 605 stage would mean an opportunity cost is incurred, in the form of giving up of stormwater reduc-
 606 tions for the years in the first stage. Another way to interpret the objective is to view $\frac{T_{II}}{T_I+T_{II}}$ as a
 607 discount factor for the worth of future benefits. This formulation allows the user to manipulate
 608 the values of T_I and T_{II} and assess how the decisions change with the discount factor.

609 3.2.3. Learning Constraints

610 Eqs. 3 and 4 are the constraints for modeling the investment – learning relationship, where
 611 $Th_{(i,j)}^{BL}$ and $Th_{(i,j)}^{AL}$ are the learning thresholds of SMP i at subcatchment j for basic learning and
 612 advanced learning, respectively; $L_{NL,(i,j)}$, $L_{BL,(i,j)}$, and $L_{AL,(i,j)}$ are binary variables that indicate
 613 whether learning happens or not; and M is an arbitrarily large number (e.g., 10^8).

$$614 \quad \begin{cases} (a): L_{N,(i,j)} + L_{B,(i,j)} + L_{A,(i,j)} = 1 \\ (b): -x_{I,(i,j)} + Th_{(i,j)}^{AL} L_{A,(i,j)} \leq 0 \\ (c): x_{I,(i,j)} - M L_{A,(i,j)} \leq Th_{(i,j)}^{AL} \\ (d): x_{I,(i,j)} - M (L_{B,(i,j)} + L_{A,(i,j)}) \leq Th_{(i,j)}^{BL} \\ (e): -x_{I,(i,j)} + Th_{(i,j)}^{BL} (L_{B,(i,j)} - \sum_{k \in Sub, k \neq j} L_{A,(i,k)}) \leq 0 \\ (f): L_{A,(i,j)} - L_{A,(i,k)} - L_{B,(i,k)} \leq 0, \quad k \in Sub, k \neq j \end{cases}, \forall i \in SMP, j \in Sub$$

$$615 \quad (3)$$

$$616 \quad \begin{cases} x_{IINs,(i,j)} - ML_{NL,(i,j)} \leq 0 \\ x_{IIBs,(i,j)} - ML_{BL,(i,j)} \leq 0 \\ x_{IIAs,(i,j)} - ML_{AL,(i,j)} \leq 0 \end{cases}, \forall i \in SMP, j \in Sub \text{ and } \forall s \in S_I \quad (4)$$

617 The constraints of Eq. 3 forcing only one of the situations (NL, BL, and AL) can happen to an
 618 SMP, based on the first-stage investment. The corresponding binary learning variable would be
 619 set to 1, and the rest would remain 0. In addition, as discussed in Section 3.1.1, we assume that
 620 for an SMP i in the subcatchment j , if AL happens, it will also update the cost-effectiveness dis-
 621 tribution of SMP i in other subcatchments with BL. This transferrable learning is modeled as Eq.
 622 3(e) and 3(f).

623 Once we know which case (NL, BL, or AL) we are in, the corresponding second stage decision
 624 variables would be activated (allows to be non-zero), and the decision variables of other situa-
 625 tions would be set to 0 (Eq. 4).

626 3.2.4. Risk Metric

627 The risk metric applied in this formulation is the conditional value at risk (CVaR). The CVaR
 628 value of a random variable, f , is the expected value of the left tail below the value of the α -th
 629 quantile, denoted $CVaR_\alpha(f)$, and the value at the α -th quantile is called the value at risk (VaR),
 630 denoted $VaR_\alpha(f)$. For example, a $CVaR_{0.1}(f)$ value of 1-million-ton stormwater reduction
 631 means that the average of the lowest 10% outcomes is 1 million tons. Therefore, a higher CVaR
 632 value is more desirable. More discussion of this risk metric and its mathematical properties is
 633 available in Artzner et al. (1999) and Rockafellar & Uryasev (2000). The mathematical defini-
 634 tions of VaR and CVaR are as follows.

$$635 \quad \text{VaR: } VaR_\alpha(f) = \text{ArgMin}_y \{ \text{Prob}(f \leq y) \geq \alpha \}$$

$$636 \quad \text{CVaR: } CVaR_\alpha(f) = \frac{1}{\alpha} \int_0^\alpha VaR_t(f) dt$$

637 We adopted the method proposed by Krokmal et al. (2001) for our formulation, where the
 638 CVaR can be calculated using linear constraints in an optimization. The method is based on
 639 discretizing the random distributions by drawing sample sets. Each sample set represents a sce-
 640 nario with probability $1/m$, where m is the number of the sample sets.

641 The mathematical formulation is as follows.

$$642 \quad \begin{cases} z_{s,r} \geq \tau - f_{IIs,r}, \forall s \in S_I, r \in S_{II} \\ \tau - \frac{1}{\alpha * m_I m_{II}} \sum_{s=1}^{m_I} \sum_{r=1}^{m_{II}} z_{s,r} \geq CVaR_\alpha, \end{cases} \quad (5)$$

$$643 \quad f_{IIs} = D_{Is} C_{Is} x_I + D_{IIs} (D_{Is} C_{Is} x_I + C_{IINs,r} x_{IINs} + C_{IIBs} x_{IIBs} + C_{IIAs,r} x_{IIAs}), \forall s \in S_I, r \in S_{II} \quad (6)$$

644 where $z_{s,r}$ is the auxiliary variable for CVaR calculation, τ is a variable representing VaR_α , $f_{IIs,r}$
 645 is a linear objective function with the coefficient set equal to the sample set of cost-effectiveness
 646 values in scenario (s,r) , and $CVaR_\alpha$ is the target CVaR value specified by the user (i.e., the least
 647 acceptable outcome having a chance of α).

648 3.2.5. Total Budget and Impervious Area

649 We assume a total budget (denoted B) in the study area for the 25-year planning horizon. The
 650 impervious areas of the three subcatchments are divided into the ground (60%) and roof (40%)

651 areas, which is approximated by visual inspection and randomly sampling points on the aerial
 652 photographs on Google Map. These numbers could be updated if further information is available
 653 for the sewershed. Eq. 7 is the total budget constraint; for each scenario, this limits the sum of
 654 annual expenditures x on each SMP within each subcatchment over the time horizon to B . Eqs. 8
 655 and 9 are the constraints imposed by the amount of impervious surface of the ground and rooftop
 656 area, respectively, upon the amounts of SMP investment of each type that can be made. The no-
 657 tation includes the following: $G = \{RG, IT, PP\}$ is the set of the ground SMPs; $R = \{RB, GR\}$ is
 658 the set of the roof SMPs; $T_{A,(i,j)}$ is drainage area ratio; $E_{c,(i,j)}$ is the prior expected cost per m^2 per
 659 year of SMP i at subcatchment j ; and $A_{G,j}$ $A_{R,j}$ are the total impervious surface area and total
 660 impervious roof area, respectively, in subcatchment j .

$$661 \sum_{i \in SMP} \sum_{j \in Sub} \{(T_I + T_{II})x_{I,(i,j)} + T_{II}(x_{IINS,(i,j)} + x_{IIBS,(i,j)} + x_{IIAS,(i,j)})\} - B \leq 0, \forall s \in S_I \quad (7)$$

$$662 \sum_{i \in G} \frac{T_{A,(i,j)}}{E_{c,(i,j)}} (x_{I,(i,j)} + x_{IINS,(i,j)} + x_{IIBS,(i,j)} + x_{IIAS,(i,j)}) - A_{G,j} \leq 0, \forall j \in Sub, s \in S_I \quad (8)$$

$$663 \sum_{i \in R} \frac{T_{A,(i,j)}}{E_{c,(i,j)}} (x_{I,(i,j)} + x_{IINS,(i,j)} + x_{IIBS,(i,j)} + x_{IIAS,(i,j)}) - A_{R,j} \leq 0, \forall j \in Sub, s \in S_I \quad (9)$$

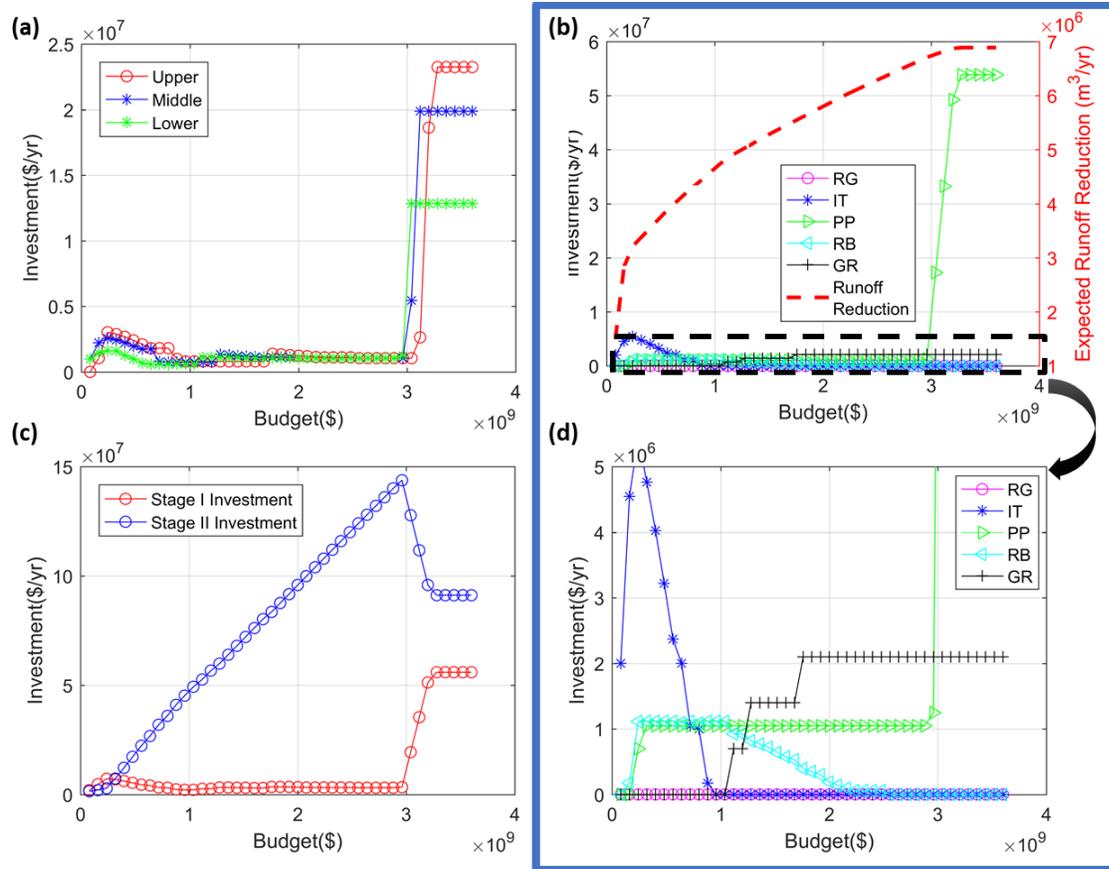
664 4. Results and Discussion

665 In this section, we describe tradeoffs between the three objectives—total stormwater reductions,
 666 total cost, and risk—by changing constraints on two of the objectives (budget and CVaR/risk) and
 667 noting their effect on the optimal decisions and resulting stormwater reductions. Increasing the
 668 budget gives us an estimate of the incremental cost of stormwater reductions using GI, while in-
 669 creasing the risk target will show how becoming more averse to the possibility of particularly
 670 bad stormwater performance has a cost in terms of worsening the expected (probability-
 671 weighted) performance. As we will show, changing the risk target favors different SMPs, and
 672 affects the attractiveness of investment in learning and deferring investment until more is known
 673 about the performance of SMPs.

674 4.1. Effect of the Budget on Stormwater Reductions, Optimal GI Investment Mix, and Learning

675 In order to focus on cost-expected stormwater reduction tradeoffs, we first did a series of runs
 676 with the CVaR constraint dropped from the model. Section 4.2, which follows, instead explores
 677 how changing our risk tolerance, as measured by the CVaR constraint, affects the results.

678 We solved the model forty times with a range of budgets for the 25-year planning horizon, vary-
 679 ing from \$0.08 billion to \$3.2 billion. The impacts on decisions and stormwater are organized in
 680 Figure 4. Part (a) shows how investments allocated among subcatchments, (b) shows the
 681 investments grouped by SMP type (the left y-axis) and how much they reduce expected annual
 682 runoff (the right y-axis), (c) shows how investment commitments are divided between the first
 683 and the second stages, and (d) is an enlargement of (b) providing more detail on how investments
 684 are allocated among SMP types.



685

686 **Figure 4.** First stage investments as a function of budget; (a) optimal GI investment allocations
 687 among three subcatchments; (b) the objective (expected annual stormwater runoff reduction; red
 688 dashed line, right y-axis [from 1 million m^3/yr to 7 million m^3/yr]) and the optimal GI investment
 689 allocations among SMPs (solid lines, right left y-axis), (c) split of the budget between invest-
 690 ments in the first and second stages, and (d) expanded view of the optimal GI investment alloca-
 691 tions among the SMPs

692 Generally, the results show strongly diminishing returns as the budget is increased (Figure 4b).
 693 Of the total possible reduction of 6.9 million m^3/yr that can be achieved by spending \$3.6 billion,
 694 one-quarter of that reduction can be achieved by spending just over \$100 million, while one-half
 695 requires only \$300 million. This is because the most cost-efficient SMP types (IT and RB) are
 696 installed first until all impervious area is treated. Then, if extra money is available, the optimal
 697 solution is to replace some low-cost SMPs with more expansive and higher reduction SMPs or to
 698 invest for learning. As a result, the expected stormwater reduction gradually flattens as the mar-
 699 ginal stormwater reduction per dollar investment diminishes. Figure 4 also shows that expand-
 700 ing the budget changes the decisions in several specific respects: in terms of location, technology
 701 choice, timing, and emphasis on learning to improve technologies. We discuss each in turn be-
 702 low. In addition to numerical results, some general relationships are established as to when particu-
 703 lar strategies are best.

704 4.1.1. Location: Does It Matter Where the GI Go?

705 In Figure 4a, we can see that when the total budget is low (\$0.08 billion, the leftmost point in
 706 Figure 4a), the optimal first-stage investment portfolio is to invest in Middle and Lower sub-
 707 catchments for \$10 million/yr each (both are investments in IT from Figure 4b) because of the
 708 low SMP costs in those locations. As more money becomes available, the Upper subcatchment
 709 starts to receive IT investment. When the budget is low (< \$0.16 billion, the second point from
 710 the left in Figure 4a), early investment in the most cost-efficient SMP (IT) provides the highest
 711 expected stormwater reductions, since there will not have much money left to make investments
 712 to improve planning even if learning happens (Figure 4c). For the same reason, we can see that
 713 with the total budget reaches \$3.0 billion, first stage investment suddenly increases, first in the
 714 Lower subcatchment and then the Middle and Upper subcatchments. Occasionally, the model
 715 may suggest investments in the Upper subcatchment for learning (AL) for its largest impervious
 716 area and the higher improvement in expected stormwater reduction. By doing so, the knowledge
 717 learned in the Upper subcatchment will be partially transferred to the other subcatchments so that
 718 we can also improve the second stage investments in these areas. This effect is most obvious
 719 when the budget is between \$0.7 and \$0.8 billion in Figure 4a.

720 4.1.2. Technology: Which GI?

721 From Figure 4d, we can see that increasing the total budget shifts the mix of technologies chosen.
 722 For example, under a low budget (\leq \$0.16 billion, the first two points from the left in Figure 4d),
 723 IT dominates, but as the budget increases (beyond \$0.24 billion to \$0.9 billion), IT is then dis-
 724 placed by PP. Similarly, investment in RB is emphasized for treating impervious roof area when
 725 the budget is low but gradually replaced by GR when the total budget climbs beyond \$1.1 billion.
 726 As mentioned above, this occurs because the technology with the greatest reduction per dollar is
 727 chosen first, even if its overall effectiveness (per unit surface area) is less. But when the budget
 728 is large enough such that the watershed's impervious area is fully treated, then further reductions
 729 are only possible by replacing technology that is the cheapest in terms of \$/unit stormwater
 730 reduction with a more costly technology that yields more stormwater reduction per unit area.
 731 Among the SMPs, PP and GR have the highest stormwater reduction per treated area (Table 5,
 732 "Per ha of Drainage Area") for the ground and roof areas, respectively, as a result of having the
 733 highest ratio of storage to total impervious area draining into the SMP. Therefore, when budgets
 734 are ample (above \$3.2 billion), the optimal solutions consist of PP and GR only.

735 4.1.3. Timing: Invest All Now, or Invest in Learning and Wait?

736 From Figure 4b, we can see that the expected stormwater reduction (red dashed line; right y-axis)
 737 increases with the increase in the total budget and reaches its maximum (6.9 million- m^3 /yr) when
 738 the total budget equals \$3.2 billion, after which the budget is no longer a limiting factor because
 739 no further investment opportunities remain. Under that ample budget (Figure 4b), we can see
 740 that the PP investment (\$54 million/yr) treats all the impervious ground area in the first stage
 741 (invest all now) whereas the GR investment (\$2.1 million/yr) is only for advanced learning in the
 742 three subcatchments (invest in learning and wait). The PP investment occurs now because that
 743 GI has a relatively high deterioration rate and low potential for learning. Consequently, storm-
 744 water reductions that could result from potential improvement in its cost-effectiveness (either via
 745 reduced costs or increased effectiveness) cannot make up for the loss of the potential stormwater

746 benefits in the first stage. In contrast, GR is the opposite, and waiting for potential cost-
747 effectiveness improvements is well worthwhile.

748 We can derive a general relationship for determining whether to invest all or to invest just
749 enough for learning in the first stage by calculating the difference in objective value (Eq. 2) of
750 those two strategies. The objective values (expected stormwater reductions) of the two strategies
751 are as follows.

752 • Invest all now: $\frac{T_I}{T_I+T_{II}} E[D_I C_I] X + \frac{T_{II}}{T_I+T_{II}} E[D_I D_{II} C_I] X$

753 • Invest just enough for learning in the first stage (AL, in this derivation) and invest the rest in
754 the second stage to exploit the cost-effectiveness improvement:

755 $\frac{T_I}{T_I+T_{II}} E[D_I C_I] (Th^{AL}) + \frac{T_{II}}{T_I+T_{II}} E[D_I D_{II} C_I] (Th^{AL}) + \frac{T_{II}}{T_I+T_{II}} E[(D_{II S} C_{IAs})] (X - Th^{AL})$

756 where X is a vector of fixed investments (not a decision variable), C_I is the prior distribution of
757 SMP cost-effectiveness, Th^{AL} is the investment threshold for AL, D_I and D_{II} are random varia-
758 bles for the performance deterioration in the first and the second stages, respectively, and $D_{II S}$
759 and C_{IAs} are random samples drawn from D_{II} and C_I , respectively. If the difference between the
760 two objective values below is positive, then investing now is best:

761 $\frac{T_I}{T_I+T_{II}} E[D_I C_I] (X - Th^{AL}) + \frac{T_{II}}{T_I+T_{II}} E[D_I D_{II} C_I] (X - Th^{AL}) - \frac{T_{II}}{T_I+T_{II}} E[D_{II S} C_{IAs}] (X - Th^{AL}) > 0$ (10)

762 Because we assume independence, the average of $D_{II S} C_{IAs}$ is the product of the mean of the ran-
763 dom variables, D_{II} and C_I , and since $E[C_{IAs}]$ equals $(1 + \gamma)E[C_I]$, we can rewrite Eq. 10 as:

764 $\frac{T_I}{T_I+T_{II}} E[D_I C_I] (X - Th^{AL}) + \frac{T_{II}}{T_I+T_{II}} E[D_I D_{II} C_I] (X - Th^{AL}) > \frac{T_{II}}{T_I+T_{II}} E[D_{II}] (1 + \gamma) E[C_I] (X - Th^{AL})$
765 (11)

766 By multiplying both sides of Eq. 11 by $(\frac{T_I+T_{II}}{E[C_I](X-Th^{AL})})$ and reorganizing, we get the following
767 inequality:

768 $\frac{E[D_I]}{E[D_{II}]} \frac{T_I}{T_{II}} + E[D_I] - 1 > \gamma$
769 (12)

770 If Eq. 12 holds, then making the entire investment in the first stage and none in the second stage
771 can reduce more runoff in terms of expectation, otherwise to invest for learning would be opti-
772 mal. In our specific case, a \$3.2 billion budget results in the inequality test being satisfied for PP
773 but not for GR, as shown in Table 8, which explains why the best timing for each of those GIs
774 differs.

775 **Table 8**

776 *The Results of the Investment Timing Test of the SMPs (Eq. 12)*

	$E[D_I]$	$E[D_{II}]$	T_I	T_{II}	$\frac{E[D_I]}{E[D_{II}]} \frac{T_I}{T_{II}} + E[D_I] - 1$	γ_{AL}	Major Investment Timing
RG	0.95	0.8	5	20	24.7%	<	30% Stage II
IT	0.95	0.75	5	20	26.7%	>	20% Stage I
PP	0.85	0.6	5	20	20.4%	>	20% Stage I
RB	0.95	0.8	5	20	24.7%	>	10% Stage I
GR	0.975	0.93	5	20	23.7%	<	30% Stage II

777 4.1.4. Learning: When is It Worthwhile to Stimulate Technological Improvement?

778 In this section's series of model runs, we did not impose the risk constraint, so the jumps in SMP
 779 investments (e.g., GR, Figure 4d) are motivated by the potential benefits of learning for improv-
 780 ing SMP cost-effectiveness, not for reducing uncertainty. However, the uncertainty in the cost-
 781 effectiveness improvement also provides an incentive for the investor to hedge the strategy. This
 782 is because the outcome of the first stage investments can be to learn which SMPs will perform
 783 better than expected *a priori* (and thus should be attractive for second-stage investment) and
 784 which SMPs will perform more poorly (and therefore should perhaps be shunned in the second
 785 stage). If there is sufficient potential to learn, then stormwater management investments in the
 786 second stage will benefit from the improved SMPs' cost-effectiveness and the knowledge gained
 787 about which SMPs will perform better than expected. That benefit can be higher than the fore-
 788 gone stormwater reductions in the first stage resulting from delaying investment.

789 This effect is illustrated by the increased PP investment in Figure 4d that occurs when the budget
 790 expands from \$0.4 to \$0.8 billion. In these cases, the PP investment in the first stage enables the
 791 stormwater manager to learn whether PP can reduce more stormwater than IT (both are for
 792 ground impervious surface) in the second stage. If that turns out to be the case, the manager can
 793 switch to invest in PP at that time, and obtain more stormwater reductions for the dollars spent.

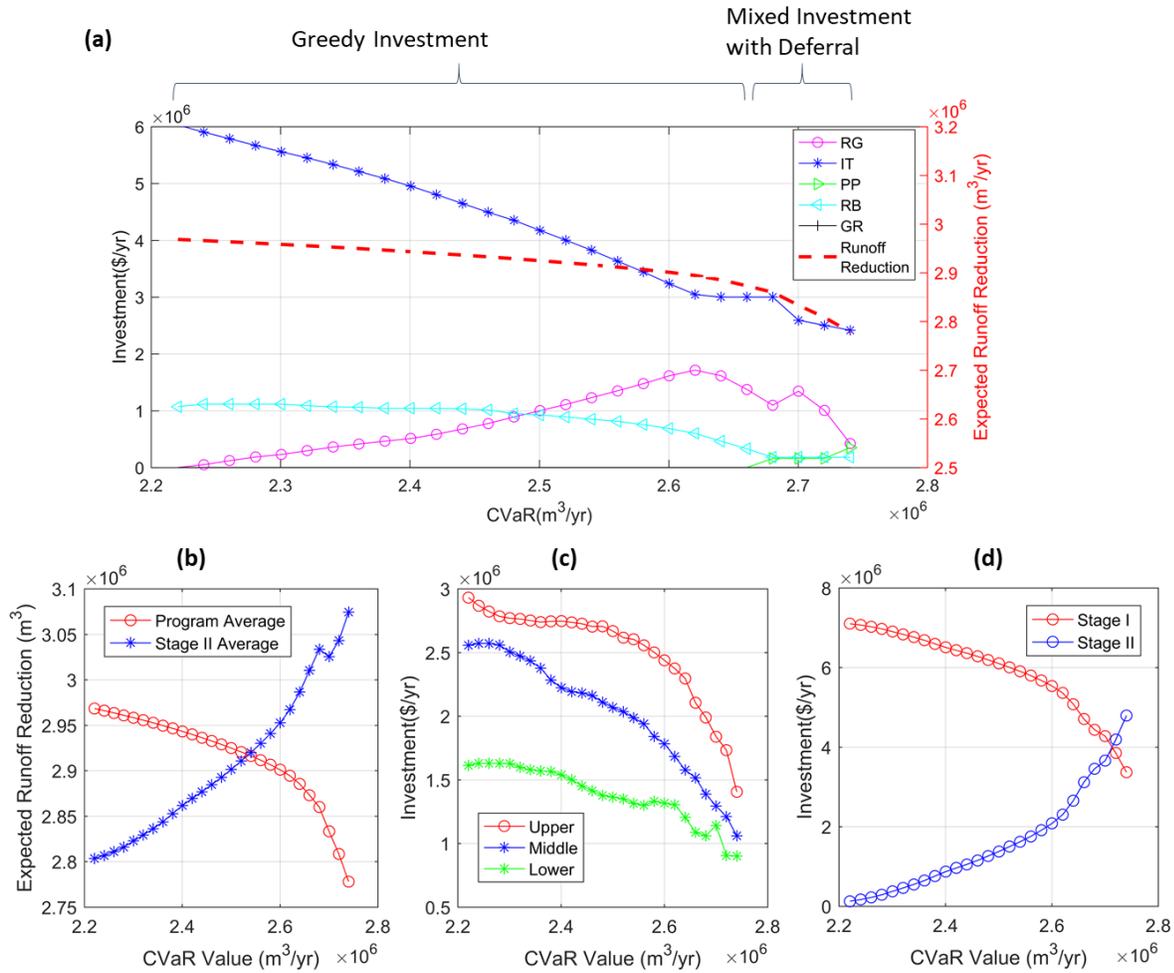
794 4.2. Effect of Risk Aversion upon Expected Stormwater Reduction Tradeoff and SMP Choices

795 Based on the above results, the best strategy under a modest budget without a risk constraint is to
 796 invest in IT and RB in the first stage. But because those SMPs have some risk of performing
 797 more poorly than expected, we anticipate that if risk aversion is considered then this strategy will
 798 change to one that has less probability of a poor stormwater performance. This hedging can be
 799 accomplished either by diversifying investment in the first stage, or by investing in learning and
 800 deferring investment until the second stage when more is known. Tightening the CVaR con-
 801 straint allows us to calculate how much the strategy shifts as we become more risk averse, and
 802 how the expected reduction as well as the risk of poor performance changes. In particular, the
 803 model quantifies how much we need to give up in expected stormwater reduction in order to re-
 804 duce the possibility of poor performance as measured by CVaR.

805 Therefore, we constrain the model results such that the average of the 10% worst stormwater
 806 outcomes (i.e., CVaR with $\alpha=0.1$) meets or exceeds a stated target, and then vary that target. The
 807 total budget is set to \$180 million, which is the amount of money needed to treat all the impervi-
 808 ous surfaces with the least-cost SMPs (IT and RB) for the 25-year planning horizon.

809 The results in Figure 5 summarizes how the investment changes as the CVaR target increased
 810 from 2.2 million to 2.74 million m^3/yr of stormwater reduction, which represent 74% to 92% of
 811 the overall expected reduction of 2.97 million m^3/yr if there is no CVaR constraint. If the CVaR
 812 target is below 2.2 million, the constraint no longer binds, and the optimal investment strategy is
 813 the risk neutral strategy – to invest in only IT and RB. The results we show in Figure 5 include:
 814 (a) the expected stormwater reduction over the 25-year planning horizon (the right y-axis in Fig-
 815 ure 5a) and the optimal mix of first stage SMP investments, (b) the expected stormwater reduc-
 816 tion over all years and the expected annual stormwater reduction in just the second stage (the last
 817 20 years, including reductions during that time yielded by both the first and second stage invest-
 818 ments), (c) how investments are allocated among the three subcatchments, and (d) how invest-

819 ment is distributed between the first and second stages (years 0-4 and 5-24, respectively). In all
 820 the solutions, the full \$180 million budget is completely spent on SMPs.



821

822 **Figure 5.** The results of simulations with different CVaR targets: (a) the expected stormwater
 823 runoff reduction and the optimal first stage investments; (b) the expected stormwater runoff re-
 824 duction and the expected runoff reduction in just the second stage; (c) investment allocations
 825 among subcatchments; and (d) split of the budget between investments in the first and second
 826 stages

827 We now focus on how investment strategies shift as the manager becomes more risk averse, as
 828 represented by an increased CVaR target. Two types of strategies are identified to manage risk,
 829 one (which we call the “greedy strategy”) dominating when risk aversion is mild, and the other
 830 (“mixed investment with deferral”) being chosen for higher CVaR targets. These two strategies
 831 are summarized in Sections 4.2.1 and 4.2.2, respectively.

832 4.2.1. First Risk-averse Strategy: Greedy Investment (CVaR between 2.2 to 2.6 million m^3/yr)

833 From Figure 5a, we can see that by changing the composition of the investment portfolio in the
 834 first stage, the CVaR value can increase from 2.2 million m^3/yr to 2.60 million m^3/yr (17% in-
 835 crease) with a minimal (0.067 million m^3/yr , or 2.3%) deterioration in the objective value (ex-

836 pected stormwater reduction). We dub this strategy “greedy investment” because the majority of
 837 the investment is made right away in cost-effective measures IT, RG and RB (Figures 5a and d).
 838 Since RG has a slightly lower standard deviation in its cost-effectiveness and less deterioration in
 839 the second stage compared to IT, the investment portfolio shifts to the less risky RG from the
 840 riskier IT as the CVaR target increases within this range. Also, a small portion of the investment
 841 (about 2 to 20%) is deferred to the second stage as another way or lessening risk.

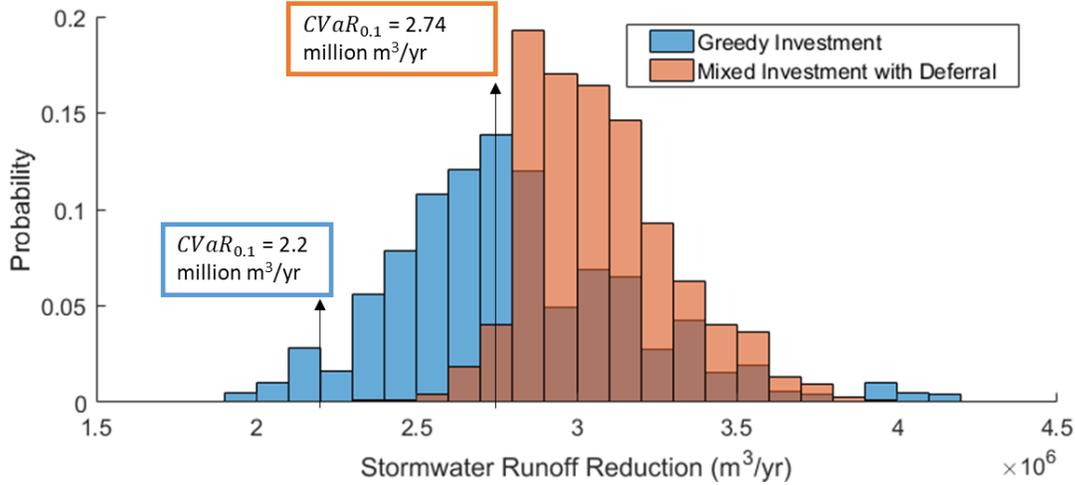
842 4.2.2. Second Risk-averse Strategy: Mixed Investment with Deferral (CVaR is between 2.6 and
 843 2.74 million m³/yr).

844 A further increase in the risk target from 2.6 million m³/yr to 2.74 million m³/yr (representing
 845 an increased aversion to poor stormwater reduction outcomes) causes the model to suggest diver-
 846 sifying the mix of first stage investments in RG, IT, PP and RB (Figure 5a). 2.74 million m³/yr
 847 is the highest CVAR value that is achievable under the \$180 million budget. It also recommends
 848 reducing first stage investment (Figure 5d). This strategy takes advantage of learning in second-
 849 stage decisions in order to lower the risk of very poor stormwater reduction outcomes, but at the
 850 expense of lower expected stormwater reductions over the entire time horizon.

851 From Figure 5b, we can see that as the CVaR target increases, the expected stormwater reduction
 852 in the second stage (“Stage II average”) increases but the objective (“program average”) decreas-
 853 es. This means that much less of the stormwater reduction is happening in the first five years.
 854 Moreover, we can see in Figure 5c that the reduction from the first stage investment is most pro-
 855 nounced in the Lower subcatchment. This is because the greater original cost-effectiveness in the
 856 Lower catchment means that the learning-based improvement would be worth more in both ex-
 857 pected stormwater reduction and CVaR.

858 Meanwhile, for the most extreme risk aversion (CVaR target = 2.74 million m³/yr), the invest-
 859 ment in PP (\$0.35 million) is for advanced learning in the Lower subcatchment. That learning
 860 would also provide a basic level of learning for the other subcatchments, so that the investment
 861 in the second stage in any subcatchment can provide higher stormwater reduction with lower un-
 862 certainty. Under this strategy, there is very little risk, since the CVaR of 2.74 million m³/yr is
 863 very close to the overall expected value of 2.78. Thus, a 6.4% sacrifice in expected value (com-
 864 pared the 2.97 million m³/yr if there is no CVaR constraint) has increased the expected perfor-
 865 mance of the 10% worse outcomes by 24.5% (from 2.2 million m³/yr to 2.74 million m³/yr).

866 Finally, Figure 6 shows the histograms of the first and last solutions of our simulations (CVaR =
 867 2.2 million m³/yr and 2.74 million m³/yr, respectively) and their CVaR values as examples to
 868 illustrate CVaR calculation. We can see that the spread of the greedy investment distribution is
 869 wider than the mixed investment with deferral distribution and, therefore, is riskier.



870 **Figure 6.** The stormwater reduction distributions of the greedy investment strategy ($CVaR = 2.2$
 871 million m^3/yr) and mixed investment with deferral (2.74 million m^3/yr)
 872

873 4.3. The Value of Learning

874 To justify use of adaptive planning, we can calculate how decisions change and how much the
 875 objective improves when we consider learning, compared to a solution chosen without consider-
 876 ing learning. Although this calculation can be in terms of any of three objectives (stormwater re-
 877 duction, cost, and risk), here we use the monetary metric (minimize expected total cost, Eq. 13)
 878 to provide a more intuitive comparison. The original objective of the model (expected annual
 879 stormwater reduction) is then treated as a constraint in this model (Eq. 14).

880 Objective: Minimize total cost

$$881 \sum_{s \in S_I} \sum_{i \in SMP} \sum_{j \in Sub} \{ (T_I + T_{II}) x_{I,(i,j)} + T_{II} (x_{IINs,(i,j)} + x_{IIBs,(i,j)} + x_{IIAs,(i,j)}) \} / m_I \quad (13)$$

882 Subject to: Expected total stormwater reduction constraint:

$$883 \frac{T_I}{T_I + T_{II}} E[D_I C_I] x_I + \frac{T_{II}}{T_I + T_{II}} (E[D_I D_{II} C_I] x_I + E[D_{II} (C_I x_{IINs} + C_{IIBs} x_{IIBs} + C_{IIAs} x_{IIAs})]) \geq ESR$$

884 (14)

885 where ESR is the target for the expected annual reduction over the 25-year life of the program.
 886 As a consequence, learning in this case reduces the expected cost of meeting a given stormwater
 887 target. To characterize the value of learning, we define and compare three types of decision mak-
 888 ing: non-adaptive, passive adaptive, and active adaptive:

- 889 • **Non-adaptive:** The decision maker recognizes that recourse actions are possible in the
 890 future but neglects to consider learning by disregarding any information generated from the
 891 first-stage investment when making the subsequent second-stage investment decisions.
 892 That is, the prior probabilities are the basis of decisions in the second stage rather than the
 893 posterior probabilities. After making a Stage I decision, the cost-effectiveness of just those
 894 investments is randomly selected; then in Stage II , the optimization assumes that the
 895 performance of the later investments is also random, with the same distribution and
 896 independent of what cost-effectiveness occurred in Stage I .

- 897 • **Passive adaptive:** The decision maker does not consider learning in making the immediate
 898 investment but when learning takes place in stage II, the investment plan is adapted based
 899 on what is learned (i.e., based on the posterior rather than prior probabilities).
- 900 • **Active adaptive:** The decision maker knows that investments can result in learning, and
 901 evaluates first stage investments, fully recognizing the value of what might be learned in
 902 the second stage.

903 To calculate the total cost of the non-adaptive case, we simply set the learning thresholds to a
 904 large number (e.g., total budget) so that the model cannot take actions in Stage I to learn. For the
 905 passive adaptive case, we restrict the model to take just the optimal Stage I investment from the
 906 non-adaptive case as an additional equality constraint (i.e., $x_I = x_{I,non-adaptive}$). As a result,
 907 that model is forced to be ignorant in the first stage but can adapt the investment plan in the sec-
 908 ond stage based on what has been learned (posterior rather than prior probabilities). Finally, the
 909 total cost of the active adaptive case is obtained by relaxing the just-mentioned first-stage in-
 910 vestment constraint to allow the model to recognize how first-stage investments might affect the
 911 second stage's learning when making first-stage investment decisions.

912 We define the difference in the objective between the non-adaptive and the passive adaptive cas-
 913 es the **value of adaptivity** and the difference between the passive and the active adaptive cases
 914 the **value of learning**. Table 9 shows an example of the calculation of the values with the total
 915 expected stormwater (*ESR*) set to 80 million m³ and the minimum annual stormwater reduction
 916 after the program (*CVaR_{0.1}*) set to 3 million m³/yr. The resulting expected total costs are \$452
 917 million, \$407 million, and \$303 million for non-adaptive, passive adaptive and active adaptive
 918 cases, respectively.

919 **Table 9**

920 *The Calculation of the Value of Adaptivity and the Value of Learning*

Unit: \$ Million	Objective (Total Cost)	Value of Adaptivity Cal- culation	Value of Learning Cal- culation
Non-adaptive	452	452	-
Passive Adaptive	407	-407	407
Active Adaptive	303	-	-303
Sum	-	+45	+104

921 The value of adaptivity means that if planning is flexible such that second-stage decisions can
 922 adapt to information gained, then second-stage choices will improve, saving \$45 million. Where-
 923 as the value of learning means that the planning process can save an additional \$104 million if
 924 first-stage decisions are made by recognizing how such decisions affect learning and how that
 925 learning will affect second-stage decisions. Although the definitions of the two values resemble
 926 in some respects the well-known concepts of the value of stochastic solutions (VSS) and the
 927 value of perfect information (VPI) (Birge, 1982), there are important differences that we believe
 928 are more realistic and useful for decision making. In particular, unlike the calculation of VSS,
 929 the non-adaptive solution is also a stochastic solution with recourse but no learning. Further,
 930 unlike VPI calculations, the passive and active adaptive solutions only learn imperfect rather
 931 than perfect information. This example shows that the adaptive approaches can save 9.7% to
 932 33.0 % of the total cost compared to the non-adaptive solution.

933 These savings are, of course, case-dependent, varying considerably depending on the prior
 934 knowledge about SMP performance, risk aversion levels, and watershed and climate characteris-
 935 tics. Nonetheless, our framework provides a way of evaluating the use of adaptive approaches
 936 while managing the risk of failing to meet management goals. The intent is to lessen the barriers
 937 that uncertain performance present to adoption of GI.

938 **5. Summary and Conclusion**

939 We propose an adaptive management framework for GI evaluation and investment planning that
 940 consists of technology efficacy assessment, optimization, and risk characterization and manage-
 941 ment. The assessment of individual SMP cost-effectiveness involves cost estimation, use of hy-
 942 drologic modeling to account for each SMP's stormwater reduction capacity, and consideration
 943 of performance deterioration as investments age. The methodology's goal is not to provide high-
 944 ly precise results for detailed design, as data for calibration and validation often not available.
 945 Rather, the intent is to provide insights on the magnitude of uncertainties and the value of adap-
 946 tation and learning in GI planning.

947 We rely on EPA SWMM to assess the SMPs' efficacy in reducing stormwater runoff because of
 948 its GI functionality and its wide use in urban stormwater management. By applying Monte Carlo
 949 methods, the SWMM hydrological simulation can be used to characterize the uncertainty of the
 950 SMPs' capabilities to control stormwater. Although our SWMM parameter calibration relies on
 951 values from the literature and expert judgment rather than field validation, the results generated
 952 from this method are traceable and are consistent among the SMPs of interest.

953 Our estimates of uncertainty in capital and maintenance costs and performance deterioration also
 954 rely on expert opinion and the literature. These estimates are combined with the results of
 955 SWMM simulations to calculate the SMPs' cost-effectiveness (in $\text{m}^3/\text{\$/yr}$). The results are then
 956 used in the adaptive GI investment model, which is structured as a decision tree. The decision
 957 tree includes prior probabilities that feed into first-stage decisions (year 0). Those first-stage
 958 ("here-and-now") decisions result in learning while second-stage (year 5, "wait and see") deci-
 959 sions take advantage of that learning to adjust the mix of GIs. The learning consists of updated
 960 probabilities (posterior) based on opportunities to learn from experience (investment) as well as
 961 learning curve-type uncertainty reductions and performance improvement.

962 The adaptive GI investment model is solved using two-stage stochastic programming. The objec-
 963 tive that is maximized is the probability-weighted stormwater reduction. This maximization is
 964 subject to constraints on two other objectives: an upper limit to the amount spent over the time
 965 horizon, and a lower limit on stormwater reduction under the worst possible outcomes of the un-
 966 certain variables, quantified as the Conditional Value at Risk ($CVaR_\alpha$).

967 The study area is the Wingohocking sewershed in Philadelphia, Pennsylvania, which we divide
 968 into three subcatchments to capture the spatial variability of sewershed characteristics. The mod-
 969 eling results presented in Section 4 show how the framework can be used to develop investment
 970 strategies and provide the economic justification of the adaptive approach (*value of learning*). In
 971 the first modeling experiment (Section 4.1), we show how decisions to invest immediately or
 972 wait until the second stage depend on how much is learned, and whether that learning would af-
 973 fect later decisions. We also see that as the total budget increases, the optimal investment portfo-
 974 lio would change in one or more of three ways (Figure 4): diversification of types of GIs at in-
 975 termediate budget levels (*hedging*), switching from the most cost-efficient SMPs to the SMPs

976 with a higher stormwater reduction per treated area (*technology switch*), and, eventually, in-
 977 creased first-stage investment in order to learn (*learning for stimulating technology improve-*
 978 *ment*).

979 In the second modeling experiment (Section 4.2), we identified two strategies that are resorted
 980 to as the manager becomes more risk averse. Under a moderately stringent CVaR constraint, a
 981 *greedy investment* strategy is adopted. It devotes most of its investment to SMPs with the highest
 982 expected stormwater reduction in the first stage, and the remainder is either invested immediately
 983 in other SMPs or deferred until the second stage in order to manage risks. A more risk averse
 984 manager would specify a higher CVaR lower bound, which yields a *mixed investment with defer-*
 985 *ral strategy*. That strategy improves the CVaR value by deferring major investments to the sec-
 986 ond stage, investing mainly for learning in the first stage. The result is a pronounced tradeoff be-
 987 tween the expected performance objective and CVaR, in which the CVaR value can be increased
 988 from 2.2 million m³/yr to 2.74 million m³/yr stormwater reduction at a cost of sacrificing 6%
 989 expected stormwater reduction objective.

990 The final experiments quantify the economic values of adaptivity and learning. These can inform
 991 stormwater manager about how performance can be improved by adaptive management, and the
 992 tradeoffs between costs and benefits of deliberate learning through research and monitoring
 993 (Walters, 1997; Williams, 2011). If the net value of learning is positive (its benefits exceed its
 994 cost), the stormwater manager should consider options for learning when planning, which may
 995 result in investments in diversification, monitoring, or deliberate experimentation. The value of
 996 adaptivity, on the other hand, represents the expected total cost saving by recognizing and taking
 997 advantage of flexibility to modify plans over time in response to changing circumstances. The
 998 decision tree/stochastic programming framework we propose can quantify these values, which
 999 are considerable in our GI case study.

1000 Future research can make our adaptive GI investment model more realistic and more reflective of
 1001 long-term costs by, for instance, considering how the dynamics of maintenance and deterioration
 1002 affect decisions concerning SMP renovation or replacement in later stages. Consideration should
 1003 also be given to additional objectives of GI planning, such as the “ancillary benefits” of energy
 1004 savings, environmental amenities, and heat island mitigation (CNT, 2009).

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 1009 tioned in the publication. We thank S. Szalay for his guidance on cost and performance deterio-
 1010 ration estimation. No new data are presented. Land use data were downloaded from
 1011 <https://www.opendataphilly.org>. Imperviousness data are retrieved from
 1012 <http://metadata.phila.gov/>. Digital elevation data are downloaded from the USGS website:
 1013 <https://lta.cr.usgs.gov/NED>. Precipitation data are downloaded from the NOAA Climatic Data Cen-
 1014 ter (www.ncdc.noaa.gov/cdo-web/). The soil survey data is retrieved from
 1015 <https://websoilsurvey.sc.egov.usda.gov/App/WebSoilSurvey.aspx>. The source code of the SWMM
 1016 Simulation in Python 2.7 can be found at Github:
 1017 https://github.com/hfengwei/Wingohocking_SWMM_MCMC, whereas the scripts of the adap-
 1018 tive GI planning optimization model in Matlab R2016b are published in another Github reposi-
 1019 tory: <https://github.com/hfengwei/Adaptvie-GI-Investment-Planning-Model>.

1020

1021 **Reference**

- 1022 Artzner, P., Delbaen, F., Eber, J.-M., & Heath, D. (1999). Coherent measures of risk.
1023 *Mathematical Finance*, 9(3), 203–228. <https://doi.org/10.1111/1467-9965.00068>
- 1024 Askarizadeh, A., Rippey, M. A., Fletcher, T. D., Feldman, D. L., Peng, J., Bowler, P., Mehring,
1025 A. S., Winfrey, B. K., Vrugt, J. A., Aghakouchak, A., Jiang, S. C., Sanders, B. F., Levin, L.
1026 A., Taylor, S., & Grant, S. B. (2015). From rain tanks to catchments: Use of low-impact
1027 development to address hydrologic symptoms of the urban stream syndrome.
1028 *Environmental Science and Technology*, 49(19), 11264–11280.
1029 <https://doi.org/10.1021/acs.est.5b01635>
- 1030 Asleson, B. C., Nestingen, R. S., Gulliver, J. S., Hozalski, R. M., & Nieber, J. L. (2009).
1031 Performance assessment of rain gardens. *Journal of the American Water Resources*
1032 *Association*, 45(4), 1019–1031.
- 1033 Avellaneda, P. M., Jefferson, A. J., Grieser, J. M., & Bush, S. A. (2017). Simulation of the
1034 cumulative hydrological response to green infrastructure. *Water Resources Research*, 53(4),
1035 3087–3101. <https://doi.org/10.1002/2016WR019836>
- 1036 Bates, B. C., & Campbell, E. P. (2001). A Markov Chain Monte Carlo scheme for parameter
1037 estimation and inference in conceptual rainfall-runoff modeling. *Water Resources Research*,
1038 37(4), 937–947. <https://doi.org/10.1029/2000WR900363>
- 1039 Bergman, M., Hedegaard, M. R., Petersen, M. F., Binning, P., Mark, O., & Mikkelsen, P. S.
1040 (2011). Evaluation of two stormwater infiltration trenches in central Copenhagen after 15
1041 years of operation. *Water Science & Technology*, 63(10), 2279.
1042 <https://doi.org/10.2166/wst.2011.158>
- 1043 Birge, J. R. (1982). The value of the stochastic solution in stochastic linear programs with fixed
1044 recourse. *Mathematical Programming*, 24(1), 314–325.
1045 <https://doi.org/10.1007/BF01585113>
- 1046 Chow, V. Te, Maidment, D. R., & Mays, L. W. (1988). *Applied Hydrology*. New York:
1047 McGraw-Hill.
- 1048 CNT. (2009). *National green value calculator methodology*. Retrieved from
1049 <http://greenvalues.cnt.org/national/downloads/methodology.pdf>
- 1050 Copeland, C. (2014). *Green infrastructure and issues in managing urban stormwater*.
1051 *Congressional Research Service*. Congressional Research Service. Retrieved from
1052 <https://fas.org/sgp/crs/misc/R43131.pdf>
- 1053 Deb, K. (2014). Multi-objective optimization. *Search Methodologies: Introductory Tutorials in*
1054 *Optimization and Decision Support Techniques* (pp. 403–449). Boston, MA: Springer US.
1055 https://doi.org/10.1007/978-1-4614-6940-7_15
- 1056 Dhakal, K. P., & Chevalier, L. R. (2017). Managing urban stormwater for urban sustainability:
1057 Barriers and policy solutions for green infrastructure application. *Journal of Environmental*
1058 *Management*, 203, 171–181. <https://doi.org/10.1016/j.jenvman.2017.07.065>
- 1059 Dietz, M. E. (2007). Low impact development practices: A review of current research and

- 1060 recommendations for future directions. *Water, Air, and Soil Pollution*, 186(1–4), 351–363.
1061 <https://doi.org/10.1007/s11270-007-9484-z>
- 1062 Dong, X., Guo, H., & Zeng, S. (2017). Enhancing future resilience in urban drainage system:
1063 Green versus grey infrastructure. *Water Research*, 124, 280–289.
1064 <https://doi.org/10.1016/j.watres.2017.07.038>
- 1065 Drake, J., & Bradford, A. (2013). Assessing the potential for restoration of surface permeability
1066 for permeable pavements through maintenance. *Water Science and Technology*, 68(9),
1067 1950–1958. <https://doi.org/10.2166/wst.2013.450>
- 1068 Eckart, K., McPhee, Z., & Bolisetti, T. (2017). Performance and implementation of low impact
1069 development – A review. *Science of the Total Environment*, 607–608, 413–432.
1070 <https://doi.org/10.1016/j.scitotenv.2017.06.254>
- 1071 Failing, L., Horn, G., & Higgins, P. (2004). Using expert judgment and stakeholder values to
1072 evaluate adaptive management options. *Ecology and Society*, 9(1). [https://doi.org/10.1890/1052-3175\(2004\)09\[0113:UEJASV\]2.0.CO;2](https://doi.org/10.1890/1052-3175(2004)09[0113:UEJASV]2.0.CO;2)
- 1073 Fletcher, T. D., Andrieu, H., & Hamel, P. (2013). Understanding, management and modelling of
1074 urban hydrology and its consequences for receiving waters: A state of the art. *Advances in*
1075 *Water Resources*, 51, 261–279. <https://doi.org/10.1016/j.advwatres.2012.09.001>
- 1076 Freni, G., Mannina, G., & Viviani, G. (2010). Urban storm-water quality management:
1077 Centralized versus source control. *Journal of Water Resources Planning and Management*,
1078 136(2), 268–278. [https://doi.org/10.1061/\(ASCE\)0733-9496\(2010\)136:2\(268\)](https://doi.org/10.1061/(ASCE)0733-9496(2010)136:2(268))
- 1079 Hinman, C., & Wulkan, B. (2012). *Low impact development technical guidance manual for*
1080 *Puget Sound*. Retrieved from
1081 http://www.psp.wa.gov/downloads/LID/20121221_LIDmanual_FINAL_secure.pdf
- 1082 Holling, C. S. (1978). *Adaptive environmental assessment and management*. John Wiley & Sons.
1083 <https://doi.org/10.1080/00139157.1986.9928829>
- 1084 Hung, F., & Hobbs, B. F. (2019). How can learning-by-doing improve decisions in stormwater
1085 management? A Bayesian-based optimization model for planning urban green infrastructure
1086 investments. *Environmental Modelling and Software*, 113, 59–72.
1087 <https://doi.org/10.1016/j.envsoft.2018.12.005>
- 1088 Jackisch, N., & Weiler, M. (2017). The hydrologic outcome of a Low Impact Development
1089 (LID) site including superposition with streamflow peaks. *Urban Water Journal*, 14(2),
1090 143–159. <https://doi.org/10.1080/1573062X.2015.1080735>
- 1091 Jarden, K. M., Jefferson, A. J., & Grieser, J. M. (2016). Assessing the effects of catchment-scale
1092 urban green infrastructure retrofits on hydrograph characteristics. *Hydrological Processes*,
1093 30(10), 1536–1550. <https://doi.org/10.1002/hyp.10736>
- 1094 Johnson, F. A., Smith, B. J., Bonneau, M., Martin, J., Romagosa, C., Mazzotti, F., Waddle, H.,
1095 Reed, R. N., Eckles, J. K., & Vitt, L. J. (2017). Expert elicitation, uncertainty, and the value
1096 of information in controlling invasive species. *Ecological Economics*, 137, 83–90.
1097 <https://doi.org/10.1016/j.ecolecon.2017.03.004>
- 1098 Krokhmal, P., Uryasev, T., Palmquist, J., Uryasev, S., Palmquist, J., & Uryasev, S. (2001).
1099 Portfolio optimization with conditional value-at-risk objective and constraints. *The Journal*
1100 *of Risk*, 4(2), 43–68. <https://doi.org/10.21314/JOR.2002.057>

- 1101 Lee, J. G., Selvakumar, A., Alvi, K., Riverson, J., Zhen, J. X., Shoemaker, L., & Lai, F. (2012).
 1102 A watershed-scale design optimization model for stormwater best management practices.
 1103 *Environmental Modelling and Software*, *37*, 6–18.
 1104 <https://doi.org/10.1016/j.envsoft.2012.04.011>
- 1105 Markewich, H. W., Pavich, M. J., & Buell, G. R. (1990). Contrasting soils and landscapes of the
 1106 Piedmont and Coastal Plain, eastern United States. *Geomorphology*, *3*(3–4), 417–447.
 1107 [https://doi.org/10.1016/0169-555X\(90\)90015-I](https://doi.org/10.1016/0169-555X(90)90015-I)
- 1108 McGarity, A. E. (2013). Watershed systems analysis for urban storm-water management to
 1109 Achieve Water Quality Goals. *Journal of Water Resources Planning and Management*,
 1110 *139*(5), 464–477.
- 1111 Medema, W., Mcintosh, B. S., & Jeffrey, P. J. (2008). From premise to practice : A critical
 1112 assessment of integrated water resources management and adaptive management
 1113 approaches in the water sector. *Ecology And Society*, *13*(2). <https://doi.org/29>
- 1114 Muleta, M. K., McMillan, J., Amenu, G. G., & Burian, S. J. (2013). Bayesian approach for
 1115 uncertainty analysis of an urban storm water model and its application to a heavily
 1116 urbanized watershed. *Journal of Hydrologic Engineering*, *18*(10), 1360–1371.
 1117 [https://doi.org/10.1061/\(ASCE\)HE.1943-5584.0000705](https://doi.org/10.1061/(ASCE)HE.1943-5584.0000705)
- 1118 Palla, A., & Gnecco, I. (2015). Hydrologic modeling of Low Impact Development systems at the
 1119 urban catchment scale. *Journal of Hydrology*, *528*, 361–368.
 1120 <https://doi.org/10.1016/j.jhydrol.2015.06.050>
- 1121 Petrucci, G., & Tassin, B. (2015). A simple model of flow-rate attenuation in sewer systems.
 1122 Application to urban stormwater source control. *Journal of Hydrology*, *522*, 534–543.
 1123 <https://doi.org/10.1016/j.jhydrol.2015.01.012>
- 1124 Philadelphia Water Department. (2015). *Green stormwater infrastructure design requirements*
 1125 *and guidelines packet*. Philadelphia.
- 1126 Probert, W. J. M., Hauser, C. E., McDonald-Madden, E., Runge, M. C., Baxter, P. W. J., &
 1127 Possingham, H. P. (2011). Managing and learning with multiple models: Objectives and
 1128 optimization algorithms. *Biological Conservation*, *144*(4), 1237–1245.
 1129 <https://doi.org/10.1016/j.biocon.2010.07.031>
- 1130 Rist, L., Campbell, B. M., & Frost, P. (2013). Adaptive management: where are we now?
 1131 *Environmental Conservation*, *40*(01), 5–18. <https://doi.org/10.1017/S0376892912000240>
- 1132 Rockafellar, R., & Uryasev, S. (2000). Optimization of conditional value at risk. *Journal of Risk*,
 1133 *2*(3), 21–41.
- 1134 Rossman, L. a. (2015). *Storm water management model user's manual version 5.1*. United States
 1135 *Environment Protection Agency*. Washington, D.C. <https://doi.org/PNR61>
- 1136 Rossman, L., & Huber, W. C. (2016). *Storm water management model reference manual volume*
 1137 *I – Hydrology* (Vol. I). Washington, D.C. Retrieved from
 1138 https://cfpub.epa.gov/si/si_public_record_report.cfm?Lab=NRMRL&dirEntryId=309346
- 1139 Runge, M. C., Converse, S. J., & Lyons, J. E. (2011). Which uncertainty? Using expert
 1140 elicitation and expected value of information to design an adaptive program. *Biological*
 1141 *Conservation*, *144*(4), 1214–1223. <https://doi.org/10.1016/j.biocon.2010.12.020>

- 1142 Sadegh, M., & Vrugt, J. A. (2014). Approximate Bayesian computation using Markov Chain
1143 Monte Carlo simulation: DREAM (ABC). *Water Resources Research*, 50(8), 6767–6787.
1144 <https://doi.org/10.1002/2014WR015386>
- 1145 Schueler, T., & Claytor, R. (2009). *2000 Maryland Stormwater design manual Volumes I & II*.
1146 Baltimore. Retrieved from
1147 http://mde.maryland.gov/programs/water/StormwaterManagementProgram/Pages/stormwater_design.aspx
1148
- 1149 Sebti, A., Fuamba, M., & Bennis, S. (2016). Optimization model for BMP selection and
1150 placement in a combined sewer. *Journal of Water Resources Planning and Management*,
1151 142(3), 04015068. [https://doi.org/10.1061/\(ASCE\)WR.1943-5452.0000620](https://doi.org/10.1061/(ASCE)WR.1943-5452.0000620)
- 1152 Stow, C. A., Reckhow, K. H., Qian, S. S., Lamon, E. C., Arhonditsis, G. B., Borsuk, M. E., &
1153 Seo, D. (2007). Approaches to evaluate water quality model parameter uncertainty for
1154 adaptive TMDL implementation. *Journal of the American Water Resources Association*,
1155 43(6), 1499–1507. <https://doi.org/10.1111/j.1752-1688.2007.00123.x>
- 1156 US EPA. (2010). *Green infrastructure case studies: Municipal policies for managing stormwater
1157 with green infrastructure*. Washington, DC. Retrieved from
1158 <http://nepis.epa.gov/Exe/ZyPURL.cgi?Dockey=P100FTEM.txt>
- 1159 US EPA. (2013). *The importance of operation and maintenance for the long-term success of
1160 green infrastructure: A review of green infrastructure O&M practices in ARRA Clean
1161 Water State Revolving Fund Projects*. Retrieved from
1162 [https://www.epa.gov/sites/production/files/2015-04/documents/green_infrastructure-
1163 om_report.pdf](https://www.epa.gov/sites/production/files/2015-04/documents/green_infrastructure-om_report.pdf)
- 1164 Walters, C. (1997). Challenges in adaptive management of riparian and coastal ecosystems.
1165 *Conservation Ecology*, 1(2). <https://doi.org/10.1111/j.1526-100X.2008.00478.x>
- 1166 Williams, B. K. (2011). Adaptive management of natural resources—framework and issues.
1167 *Journal of Environmental Management*, 92(5), 1346–1353.
1168 <https://doi.org/10.1016/j.jenvman.2010.10.041>
- 1169 Williams, B. K., & Brown, E. D. (2014). Adaptive management: From more talk to real action.
1170 *Environmental Management*, 53(2), 465–479. <https://doi.org/10.1007/s00267-013-0205-7>
- 1171 Williams, B. K., & Johnson, F. A. (2015). Value of information and natural resources decision-
1172 making. *Wildlife Society Bulletin*, 39(3), 488–496. <https://doi.org/10.1002/wsb.575>
- 1173 Wise, S., Braden, J., Ghalayini, D., Grant, J., Kloss, C., MacMullan, E., Morse, S., Montalto, F.,
1174 Nees, D., Nowak, D., Peck, S., Shaikh, S., & Yu, C. (2010). Integrating Valuation Methods
1175 to Recognize Green Infrastructure’s Multiple Benefits. In *Low Impact Development 2010*
1176 (pp. 1123–1143). Reston, VA: American Society of Civil Engineers.
1177 [https://doi.org/10.1061/41099\(367\)98](https://doi.org/10.1061/41099(367)98)
- 1178 Wright, T. J., Liu, Y., Carroll, N. J., Ahiablame, L. M., & Engel, B. A. (2016). Retrofitting LID
1179 Practices into Existing Neighborhoods: Is It Worth It? *Environmental Management*, 57(4),
1180 856–867. <https://doi.org/10.1007/s00267-015-0651-5>
1181
1182