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2 **Dust Under The Radar: Rethinking How to Evaluate the Impacts of Dust Events on Air Quality in**
3 **the United States**
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17 **Abstract**

18 Dust is an important and complex constituent of the atmospheric system, having significant
19 impacts on the environment, climate, air quality, and human health. Although dust events are
20 common across many regions of the United States, their impacts are not often prioritized in air
21 quality mitigation strategies. We argue that there are at least three factors that result in
22 underestimation of the social and environmental impact of dust events, making them receive less
23 attention. These include (1) sparse monitoring stations with irregular spatial distribution in dust-
24 influenced regions, (2) inconsistency with dust sampling methods, and (3) sampling frequency
25 and schedules, which can lead to missed dust events or underestimation of dust particle
26 concentrations. Without addressing these three factors, it is challenging to characterize and
27 understand the full air quality impacts of dust events in the United States. This paper highlights
28 the need for additional monitoring to measure these events so that we can more fully evaluate
29 and understand their impacts, as they are predicted to increase with climate change.

30 **Plain Language Summary**

31 Dust is an important and complex component of the atmospheric system, having significant
32 impacts on the environment, climate, air quality, and human health. Yet, dust events are
33 underestimation and therefore do not receive the level of attention necessary to fully understand
34 their impacts. This is due in part to the fact that monitoring stations are sparse in dust-influenced
35 regions; there is a lack of consistency in dust definitions and sampling methods, and there is a
36 lack of continuous monitoring, which results in missed dust events or underestimations of their
37 particle concentration. This commentary highlights the need for additional monitoring designed
38 to measure dust so that we can more fully understand its impacts, which are particularly
39 important as dust is expected to increase with climate change.

40

41 **Main Text**

42 Dust particles are emitted and suspended in the atmosphere through entrainment by
43 strong winds across erodible surfaces. This process results in the creation of dust events known
44 as blowing dust events or dust storms; the latter is considered more severe and is defined by
45 conditions resulting in horizontal visibility below 1 km, while blowing dust events are defined by
46 visibility conditions from 1 km to 10 km (WMO, 2019). These dust events have an important
47 effect on the atmospheric system, influencing radiation (Lau et al., 2020), cloud formation (Chen
48 et al., 2019), and the atmospheric vertical electric field (Ardon-Dryer et al., 2022a). Air quality,
49 human well-being, and human health are also affected, mainly negatively (Ardon-Dryer and
50 Kelley, 2022; Tong et al., 2023a,b). Dust in North America is estimated to contribute $\sim 2.5\%$
51 ($0.3\text{--}0.9$ Tg) of the global dust loading for particles with diameters up to $20\ \mu\text{m}$ (PM_{20}) (Kok et
52 al., 2021), yet some studies estimated much higher concentrations as found by Urban et al.
53 (2018), who estimated higher dust emissions just from the Mojave Desert ($7\text{--}15$ Tg year⁻¹ for
54 PM_{20}). Further, climate models predict that with increases in aridity, dust events will likely
55 increase in the future (Pu and Ginoux, 2017; Achakulwisut et al., 2018; Brey et al., 2020).

56 Besides monitoring, other challenges exist when characterizing and managing dust levels
57 in the atmosphere. At the most basic level, a lack of a consistent, or shared, definition of dust
58 amongst scientists, the public, and regulators (Kroepsch and Clifford 2021; Tong et al., 2022)
59 limits the advancement of dust science, in part because of a lack of consensus regarding standard

60 dust measurements. For example, airborne dust is defined in different ways. Some studies define
61 dust based on the concentration of PM_{10} (mass of particles with diameters less than 10 μm ; Lei
62 and Wang, 2014; Eagar et al., 2017). Other studies have used coarse mass ($PM_{10} - PM_{2.5}$) (where
63 $PM_{2.5}$ is the mass of particles with diameters less than 2.5 μm ; Hand et al., 2017), or the ratio of
64 $PM_{2.5}$ and PM_{10} (Tong et al., 2017) to indicate dust impacts. Fine dust, based on elemental
65 composition, has also been used to characterize dust impacts (e.g., Malm et al., 1994; Chow et
66 al., 2015; Hand et al., 2017; Liu et al., 2022). The definition of the severity of the dust impact is
67 another challenge, as ‘blowing dust event’ and ‘dust storm’ are often used interchangeably or
68 broadly, which can cause confusion or limit the interpretation of the impacts of dust events,
69 especially with respect to health applications (Ardon-Dryer et al., 2023). Another challenge is
70 that dust is often dismissed in policy and management decisions (Clifford, 2022). For example, a
71 rule within the Clean Air Act allows dust events to be removed from regulatory datasets when
72 considered an “exceptional” event (Clifford, 2021). Such data exclusions occur in both rural and
73 urban settings, with fine and coarse particles, and while more common in the arid West, are a
74 potential issue across the country. These challenges increase the difficulty of accurately tracking
75 spatial and temporal trends in dust and its impacts on air quality.

76 We highlight three major factors that interfere with accurately estimating the impacts of
77 dust events on air quality. These issues likely lead to an underestimation of the actual dust
78 concentrations, which limits our understanding of the air quality impacts and the many issues
79 associated with and caused by dust across the United States. These three factors are described
80 below:

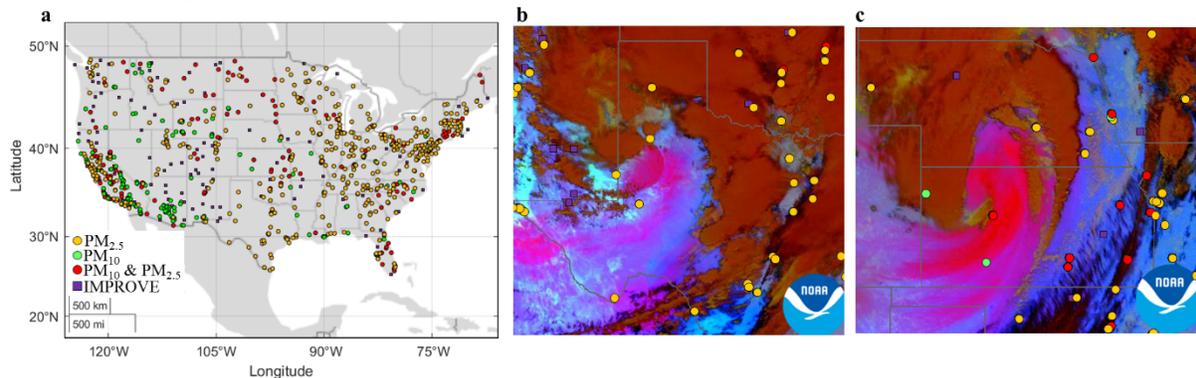
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82 ***Large spatial gaps in data result in unmonitored events***

83 Dust data are often based on the impact of dust on particulate matter (PM) measurements
84 of $PM_{2.5}$ and PM_{10} gravimetric mass, which also include contributions from other aerosol species.
85 $PM_{2.5}$ and PM_{10} are routinely monitored by the United States Environmental Protection Agency
86 (EPA) as part of large-scale monitoring networks in mostly urban settings in support of the
87 National Ambient Air Quality Standards (NAAQS). In addition, the EPA operates the large-scale
88 Chemical Speciation Network (CSN) for health exposure studies (Solomon et al., 2014), with
89 sites mainly in urban and suburban settings. Aerosol speciation measurements also occur at
90 remote and rural sites by the Interagency Monitoring of Protected Visual Environments
91 (IMPROVE) aerosol speciation network for the purpose of monitoring visibility (Malm et al.,
92 1994; see Figure 1a for map of network site locations). While the networks often have similar
93 measurement strategies and sampling schedules, they were not initially designed as dust
94 monitoring networks. However, combining data from several hundred sites across the networks
95 has led to a better understanding of large-scale spatial patterns and seasonality of dust across the
96 United States (Hand et al., 2017; 2019; Tong et al., 2017; Aryal and Evans, 2022). Nevertheless,
97 it is clear from these studies that the spatial gaps between and within monitoring networks are
98 limiting our knowledge of dust impacts on air quality.

99 In particular, regions influenced by dust events, as observed from satellites, often do not
100 have PM monitors. For example, only 21% of counties in the United States have $PM_{2.5}$ monitors
101 (Sullivan and Krupnick, 2018), and most of the PM monitoring that supports NAAQS and other
102 health exposure studies are in areas with high population density, resulting in spatial gaps in
103 many rural areas that often experience dust events (see examples of dust events in Figure 1b and

104 1c). In recent years, commercially available, low-cost air quality sensors have become common,
 105 and regulatory agencies started using them to obtain more granular information on air quality
 106 spatial and temporal distribution (Jaffe et al., 2023). However, the accuracy and precision of
 107 these sensors need to be characterized (Zheng et al., 2018), the sensors require calibrations
 108 (Ardon-Dryer et al., 2020) and recent work suggests they are unable to accurately characterize
 109 coarse particles ($> 2.5 \mu\text{m}$) (Jaffe et al., 2023; Kaur and Kelly, 2023; Rueda et al., 2023) and they
 110 still contain spatial gaps. These spatial gaps limit our ability to fully quantify the number of dust
 111 events and their subsequent impacts. While satellite observations provide useful spatial and
 112 temporal information regarding large dust events, as well as identify affected regions void of
 113 monitors, satellites cannot replace ground-based monitoring to adequately characterize dust
 114 impacts on surface air quality, and satellites may miss dust events depending on the timing of
 115 satellite coverage.



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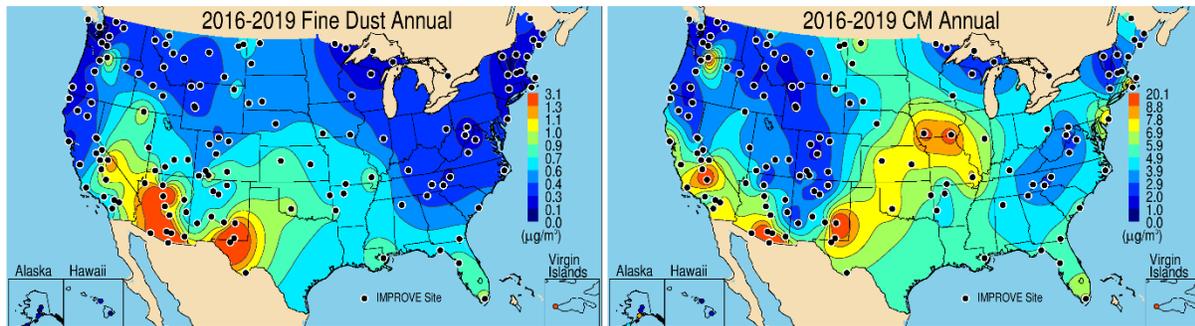
117 **Figure 1.** (a) Large-scale monitoring networks, including EPA's Federal Reference Method
 118 (FRM) networks for only PM_{2.5} (orange), only PM₁₀ (green), collocated PM_{2.5} and PM₁₀ (red), as
 119 well as IMPROVE sites (PM_{2.5} and PM₁₀, purple) that were active in 2020. (b) Examples of dust
 120 events (in pink) captured by the GOES-16 satellite over Texas (22 March 2021), and (c) over
 121 Kansas (15 December 2021), highlighting areas void of sensors, with PM_{2.5} (orange), PM₁₀
 122 (green), PM_{2.5} with PM₁₀ (red), and IMPROVE sites (purple). Satellite images retrieved from
 123 AerosolWatch (<https://www.star.nesdis.noaa.gov/smcd/spb/aq/AerosolWatch/>, accessed on 11
 124 February 2023).

125

126 *Monitors may miss dust events due to particle size*

127 Dust particle size can range across several orders of magnitude ($\sim 100 \text{ nm}$ to $100 \mu\text{m}$,
 128 Scheuven and Kandler, 2014; Ardon-Dryer and Kelley, 2022; Ardon-Dryer et al., 2022b). Some
 129 of these inhalable particles have serious health effects (Martinelli et al., 2013; Tobias et al.,
 130 2019). Measurements of PM_{2.5} or dust concentrations derived from PM_{2.5} elemental speciation
 131 measurements could lead to underestimates of dust impacts on air quality as they miss the
 132 fraction of dust particles which are often associated with the coarse aerosol mode (PM₁₀-PM_{2.5}),
 133 or even giant mode dust (van der Does et al., 2018). Differences in spatial and seasonal
 134 variability of reconstructed fine dust and coarse mass suggest that knowledge of dust size
 135 distribution and coarse mass composition may be critical for understanding and reconciling
 136 PM_{2.5} and PM₁₀ data that are influenced by dust events (Hand et al., 2023). Annual mean fine
 137 dust concentrations from 2016 through 2019 at sites from the IMPROVE network are shown in
 138 Figure 2a, compared to the annual mean coarse mass at the same sites (Figure 2b). Differences in

139 spatial patterns, especially at sites in the central United States, may be due to dust size
 140 distribution, or different sources and composition of coarse mass (Malm et al., 2007; Bondy et
 141 al., 2018). Without these additional measurements, we can only speculate regarding the sources
 142 and transport of dust. In addition, biases in fine dust concentrations derived from measurements
 143 using collocated but different samplers from the CSN and IMPROVE networks are likely due to
 144 the sharpness of the cut point of the sampler that allows varying amounts of coarse mode dust to
 145 be collected on the PM_{2.5} filter (Solomon et al., 2014; Gorham et al., 2020; Hand et al., 2012;
 146 2023).



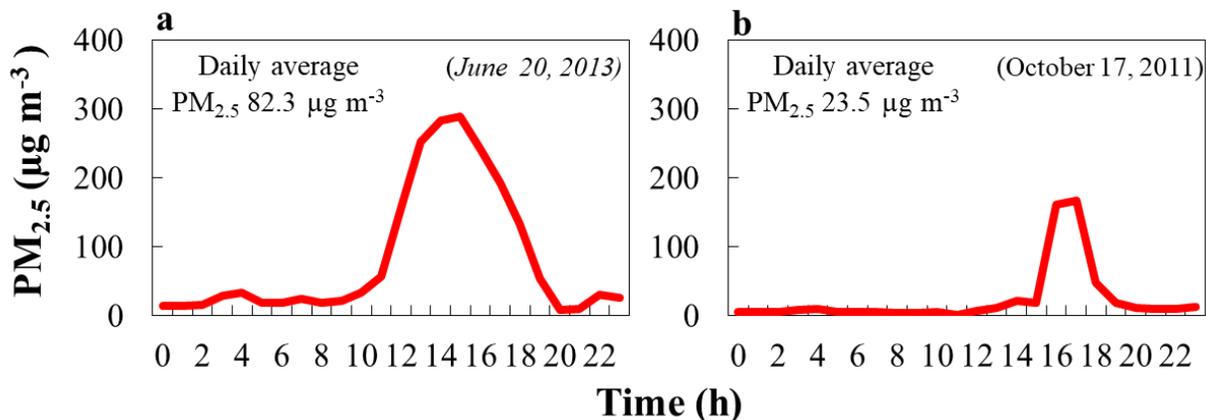
147
 148 **Figure 2.** 2016-2019 IMPROVE annual mean (a) reconstructed fine (PM_{2.5}) dust ($\mu\text{g m}^{-3}$) and
 149 (b) gravimetric coarse mass (PM₁₀-PM_{2.5}, $\mu\text{g m}^{-3}$), adapted from Hand et al., 2023.

150

151 ***Sampling frequency and duration limit the ability to detect the impact of dust events on air***
 152 ***quality***

153 While some of the EPA FRM PM_{2.5} and PM₁₀ sensors provide hourly values (EPA,
 154 2020), most networks collect aerosol on filters for 24 hours every third, sixth, or twelfth day,
 155 depending on the site. Given these schedules, dust events are often missed. Dust events can
 156 happen on time scales of hours or less, and some may exceed EPA daily thresholds, EPA daily
 157 thresholds are $35 \mu\text{g m}^{-3}$ for PM_{2.5} and $150 \mu\text{g m}^{-3}$ for PM₁₀ (Figure 3a). Integrated samples or 24-
 158 hour averages of hourly data may mask the severity and true impact of these events, leading to a
 159 low daily threshold (Figure 3b). In a recent study, Ardon-Dryer and Kelley (2022) showed that
 160 even hourly PM measurements may mask the contribution of PM concentrations during short-
 161 duration (sub-hourly) dust events, leading to underestimation of the PM concentrations during
 162 these short and intense events. While overlooked, these short-duration dust events still carry
 163 important impacts; studies found that exposure to high PM concentrations (of fine and coarse
 164 PM) causes health issues (Pérez et al., 2008; Malig and Ostro, 2009; Martinelli et al., 2013;
 165 Tobias et al., 2019). Yet, it is still unclear what the health consequences are of such intense but
 166 relatively short exposure to dust particles (e.g., acute exposure). This raises questions about
 167 whether we might need different air quality standards that account for such short-duration dust
 168 emissions (Bouet et al., 2019), especially as it is speculated these are more common across the
 169 United States compared to dust events around the world.

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171
 172 **Figure 3.** Hourly gravimetric PM_{2.5} measurements of dust events measured by TCEQ from
 173 Lubbock, Texas with daily values (a) Example of the hourly values during longer-term dust
 174 events show high daily PM_{2.5} values. (b) Example of short-term dust events that do not exceed
 175 the daily PM_{2.5} threshold.

176
 177 ***Moving Forward***

178 Dust can contribute significantly to particulate matter in the United States, especially on a
 179 seasonal basis. Current estimates suggest that half of the PM_{2.5} mass during spring in the
 180 Southwest is fine dust, and coarse mass can contribute more than 70% to PM₁₀ across the West
 181 on an annual basis (Hand et al., 2017; 2019, 2023). However, without consistent, more frequent
 182 measurements in dust-influenced regions, these are likely underestimates. Recent studies suggest
 183 that dust loading will increase due to climate change (Pu and Ginoux, 2017; Achakulwisut et al.,
 184 2018) but without additional data, we are limited in our ability to fully characterize its impacts
 185 on air quality and therefore on health. Even designing appropriate mitigation strategies to reduce
 186 its future impacts will require accurate characterizations of dust events. The first step is to
 187 enhance the monitoring network with additional monitors that sample more frequently,
 188 especially in dust-influenced regions. Consistent measurements of particle size and composition
 189 would also help characterize the environmental, climate, and human health impacts of dust
 190 events. Because these additions are likely cost prohibitive, the development and deployment of
 191 low-cost sensors that can accurately measure coarse PM concentrations during dust events may
 192 be a path forward to help understand the true impacts of dust on air quality in the United States.

193
 194 **Acknowledgment**

195 This research did not receive any specific grant from funding agencies in the public, commercial,
 196 or not-for profit sectors. The authors would like to thank IMPROVE network for the use of the
 197 network data. IMPROVE is a collaborative association of state, tribal, and federal agencies, and
 198 international partners. Funded by the U.S. Environmental Protection Agency (EPA) with
 199 research support from the National Park Service.

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