

1 Using observed soil moisture to constrain the uncertainty of simulated hydrological fluxes

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9

10 **Running title:** Constraining simulated fluxes with soil moisture

11 **Keywords:** Land surface models; soil moisture; Uncertainty; Pedotransfer functions

12

13 **Abstract**

14 Using data from five long-term field sites measuring soil moisture, we show the limitations
15 of using soil moisture observations alone to constrain modelled hydrological fluxes. We test
16 a land surface model, MESH/CLASS, with two configurations: one where the soil hydraulic
17 properties are determined using a pedotransfer function (the texture-based calibration) and
18 one where they are assigned directly (the hydraulic properties-based calibration). The
19 hydraulic properties-based calibration outperforms the texture-based calibration in terms of
20 reproducing changes in soil moisture storage within a 1.6 m deep profile at each site, but
21 both perform reasonably well, especially in the summer months. When the models are
22 constrained using observations of changes in soil moisture, the predicted hydrological fluxes
23 are subject to very large uncertainties associated with equifinality. The uncertainty is larger
24 for the hydraulic properties-based calibration, even though the performance was better. We
25 argue that since the pedotransfer functions constrain the model parameters in the texture-
26 based calibrations in an unrealistic way, the texture-based calibration underestimates the
27 uncertainty in the fluxes. We recommend that reproducing observed cumulative changes in
28 soil moisture storage should be considered a necessary but insufficient criterion of model
29 success. Additional sources of information are needed to reduce uncertainties, and these
30 could include improved estimation of the soil hydraulic properties and direct observations of
31 fluxes, particularly evapotranspiration.

32 1. Introduction

33 Land surface models simulate the vertical exchanges of water and energy between the land
34 and the atmosphere. These coupled models integrate many different interacting processes
35 and are capable of outputting a large number of state variables and fluxes. Soil moisture is
36 the largest store of water on the land surface (aside from lakes) and is probably the easiest
37 hydrological variable to measure. That is not to say that obtaining representative
38 measurements of soil moisture is straightforward – particular challenges are to measure soil
39 moisture at the appropriate scale and to capture the spatial variability (Vereecken et al.,
40 2008, Peterson et al., 2016, 2019, Pan et al., 2017). On the other hand, Mälicke et al. (2020)
41 looked at networks of soil moisture sensors and showed considerable organization in the
42 spatial response of soil moisture changes, concluding that it is the temporal information
43 from a few sensors, rather than the spatial variation from a large number of sensors, that is
44 most valuable for capturing catchment scale moisture dynamics. Often the most useful
45 metric of soil moisture is field scale (10^4 - 10^6 m² area) root zone (0.5 – 2 m depth) storage, S
46 (mm), where $S = 10^3 V_w / A$ and V_w (m³) is the volume of water and A (m²) is the land surface
47 area. Such metrics can be useful for irrigation planning, crop yield assessment, and other
48 vegetation productivity measures (Vereecken et al., 2008).

49

50 Consider the simple soil water balance equation

$$\frac{dS}{dt} = I - T - E - D \quad (1)$$

51

52 Where I, T, E and D are infiltration, transpiration, evaporation and drainage, all with units
53 (mm/d). Changes in soil moisture over some time increment are obtained by

$$\Delta S_i = \int_{t=t_i}^{t_{i+1}} (I - T - E - D) dt \quad (2)$$

54

55 Subsequent values of storage are given by

$$S_{i+1} = S_i + \Delta S_i \quad (3)$$

56

57 The water balance error, ε (mm), over some time increment is given by

$$\varepsilon = \Delta S_i - \int_{t=t_i}^{t_{i+1}} (I - T - E - D) dt \quad (4)$$

58

59 Equation 4 may be solved over long-term intervals (monthly, annually or multi-year) for
60 observations and field estimates of fluxes to assess consistency of those data, and large
61 errors (e.g. ± 50 mm/year) can be expected (e.g. Pan et al., 2017). It may also be solved for
62 models over the entire simulation period, to assess model errors in the simulated states and
63 fluxes, and in this case the errors should be very small (e.g. $\pm 10^{-3}$ mm/year).

64

65 Each of the fluxes in the above equations are simulated by the model – that is to say, they
66 are not boundary conditions (most cannot be directly measured, Vereecken et al., 2008, Li
67 et al., 2019), and they are all conditioned by the assumptions within the model and subject
68 to model errors and uncertainties. It is clear from Equations 1-3 that information about the
69 changes in soil moisture, ΔS , does not uniquely constrain the values that the fluxes should
70 take. Different fluxes can compensate for one another: if the input flux (I) is overestimated
71 then overestimating one or more of the output fluxes can result in the correct change in
72 storage and hence soil moisture; if one output flux is overestimated, this can be
73 compensated for by underestimating another output flux. However, it is also the case that
74 each flux is in some way dependent on the soil storage: the infiltration capacity and soil
75 evaporation rate depend on the level of saturation of the soil at the ground surface;
76 drainage from a free-drainage boundary depends on the saturation of the soil at the base of
77 the soil profile; and transpiration depends on soil water stress, determined by the saturation
78 of the soil over the root zone (Seneviratne et al., 2010). Given all of this, for any given model
79 and particular field site, it is not clear how much information content observations of soil
80 moisture provide for uniquely constraining the model fluxes. We can say that models should
81 be required to reproduce observed changes in soil moisture storage (within some
82 acceptable error tolerance). We cannot say whether a model that does reproduce observed
83 changes in soil moisture storage will necessarily correctly simulate the model fluxes.

84

85 Many studies have shown the benefits of bringing soil moisture observations to bear on the
86 calibration and validation of hydrological models (Maheu et al. 2018, examples cited by
87 Vereecken et al., 2008, a remote sensing example from Nijzink et al, 2018). Few studies have

88 explicitly looked at the question of how much information is provided by using soil moisture
89 alone to constrain models. Vereecken et al. (2008) do not provide an explicit answer to this
90 question, though they do indirectly address this when they suggest that soil moisture data
91 alone is often insufficient to estimate of soil hydraulic properties, and that inclusion of
92 observed hydrological fluxes will improve the well posedness of the inverse problem. Li et
93 al. (2019) suggested a more positive answer – they suggest that soil moisture observations
94 could be used to estimate hydrological fluxes for a series of irrigated cropped land
95 experiments. However, whilst they did consider uncertainty associated with the observed
96 soil moisture and other data, they did not rigorously look at parameter uncertainty and did
97 not consider the problem of uncertainty associated with equifinality.

98

99 In this study, we investigate the question of how useful observations of soil moisture are to
100 calibrate a land surface model and constrain the uncertainty in predictions of runoff,
101 infiltration, evapotranspiration and drainage, all of which are crucial fluxes for use in
102 hydrological and biogeochemical simulations. We apply the MESH/CLASS land surface model
103 to five diverse instrumented field sites along a south-north transect in Saskatchewan,
104 Canada. All sites are seasonally frozen, two are located in the Canadian prairies and three in
105 the southern boreal forest.

106

107 **2. Methods**

108 **2.1 Soil moisture as a metric of model performance**

109 We have available field measurements of volumetric liquid water content, θ ($\text{m}^3 \text{m}^{-3}$), at a
110 number of discrete depth intervals, in the profile, z_j (m), and recording at discrete intervals
111 in time, t_i (d). The profile storage, S (mm), is given by

$$S = \int_{z=0}^{z_N} \theta dz = \sum_{j=1}^N \theta_j \Delta z_j \quad (5)$$

112

113 A more useful metric of storage for use with calibrating models is the cumulative change in
114 storage, Ω (mm). This is defined:

$$\Omega = \int_{t=0}^{t_i} dS = S_i - S_0 \quad (6)$$

115

116 We can also define the analogous cumulative change in volumetric water content, Θ ($\text{m}^3 \text{m}^{-3}$) as

$$\Theta = \int_{t=0}^{t_i} d\theta = \theta_i - \theta_0 \quad (6)$$

118

119 The benefits of calibrating and validating models against Ω or Θ rather than S or θ are
120 twofold. Firstly, observations of soil moisture are made indirectly, and an instrument
121 calibration relationship is required for the probe that relates some measured variable
122 (typically the dielectric constant) to the volumetric (liquid) water content (Gardner et al.,
123 2003). These calibrations can introduce errors, but the errors in the change in water
124 content, measured between two points in time, are smaller than the errors in the absolute
125 water contents. Hence, we have more reliable observations of the changes in water content,
126 and hence changes in storage, than we do of the actual water contents. The second reason
127 is that in most soil moisture models, hydraulic properties are determined using a normalized
128 measure of water content, the effective saturation, given by

$$S_e = \frac{\theta - \theta_r}{\theta_p - \theta_r} \quad (7)$$

129

130 where θ_r and θ_p are the residual and saturated water content. The simulated cumulative
131 change in θ (i.e. Θ_i) is the same as the simulated cumulative change in $\theta - \theta_r$ (i.e. from
132 equation 6 we have $\theta_i - \theta_r - \theta_0 + \theta_r = \theta_i - \theta_0 = \Theta_i$). Therefore, any arbitrary value for θ_r can be
133 used to simulate Θ_i ($\theta_r = 0$ is a sensible choice) and the number of free calibration
134 parameters can be reduced by one. Moreover, after the model has been calibrated to
135 reproduce Θ_{OBS} , if desired, the actual values of θ_i can be obtained from $\theta_i = \Theta_i - \overline{\Theta_i} + \overline{\theta_{OBS}}$,
136 (and, if desired, the parameters θ_r and θ_p can be rescaled in the same way).

137

138 2.2 Field sites

139 We collected field data at five instrumented sites aligned from south to north in central
140 Saskatchewan, Canada. Saskatchewan has a continental and seasonally frozen climate and a
141 general trend of increasing mean annual precipitation and decreasing mean annual

142 temperature from south to north. Soil type and vegetation cover at each site is summarized
143 in Table 1.

144

145 The two prairie sites are in the Moist Mixed Grass ecoregion, which is correlated with semi-
146 arid moisture conditions and Dark Brown Chernozemic soils (Ecological Stratification
147 Working Group, 1996). The Kenaston site, located 90 km southeast of Saskatoon, SK, is in a
148 grazing pasture surrounded by annual crops. The topography is flat to slightly undulating,
149 with slopes of less than 2% (Burns et al., 2016). The water table resides between 3 to 5 m
150 below ground surface (Pan et al., 2017). Mean annual precipitation is about 330 mm (2009 –
151 2014), of which 70 mm falls as snow. The mean temperature in January and July is –12.9 and
152 18.8 °C, respectively. (Pan et al., 2017).

153

154 St. Denis National Wildlife Area (SDNWA) is located east of Saskatoon, SK, and
155 approximately 100 km north of the Kenaston site at the boundary between Moist Mixed
156 Grass and Aspen Parkland ecoregions. The topography of the site and surrounding
157 landscape is hummocky, with local relief of the order of 15 % (Miller et al., 1985).
158 Vegetation at the site is mainly native and non-native grasses, and riparian vegetation or
159 “willow-rings” surrounding numerous ponds. The soil moisture site is in a lowland between
160 two ephemeral ponds. Here, the water table has been observed within 1 m of the ground
161 surface. Annual precipitation is approximately 360 mm, of which 1/3 is snowfall, but is
162 highly variable (Bam et al., 2019). Annual potential evaporation exceeds precipitation in the
163 region, and annual open water evaporation is approximately 700 mm (Parsons et al., 2004).
164 Mean January and July temperatures (1991-2018) are -16.2 and 17.7 °C, respectively.

165

166 The BERMS (Boreal Ecosystem Research and Monitoring Sites) study area is approximately
167 200 km northeast of Saskatoon, SK, in the Mid-Boreal Upland ecoregion, which is
168 characterized by sub-humid climate and mixed coniferous and deciduous forest. The sites
169 are located in three mature forest stands and include flux-towers and soil moisture
170 monitoring sensors. Mean air temperature in January and July in nearby Waskesiu Lake, SK,
171 for the period 1971-2000 are -17.9 and 16.2 °C, respectively; during the same period mean
172 annual precipitation was 467 mm, 30% of which fell as snow (Barr et al., 2012).

173

174 The Old Aspen site (OAS) and surrounding landscape consists of hummocky terrain and
 175 trembling aspen overstory. The underlying geology is glacial till beneath a 10 cm organic
 176 layer and 30 cm silt loam mineral horizon. The water table is generally found 1 to 5 m below
 177 the ground surface (Barr et al., 2012). The Old Black Spruce site (OBS) has a shallow water
 178 table 0 to 1 m below the ground surface, with soil layers that comprise a 20-30 cm deep
 179 peat layer overlying poorly-drained sand (Barr et al., 2012). The Old Jack Pine (OJP) site is
 180 predominantly jack pine overstory in a well-drained sandy soil. The water table is at least 5
 181 m below the surface at OJP (Barr et al., 2012).

182

	Prairie sites		Forest sites		
	Kenaston	St. Denis National Wildlife Area	Old Jack Pine	Old Black Spruce	Old Aspen
Location	51.382°N 106.416°W	52.208°N 106.093°W	53.916°N 104.690°W	53.987°N 105.117°W	53.629 °N 106.200°W
Elevation (m ASL)	596	557.5	579.3	628.9	600.6
Vegetation/ ground cover	Wheatgrasses (<i>Agropyron</i> sp.) and needle grasses (<i>Stipa</i> sp.)	Native prairie grasses, smooth brome (<i>Bromus inermis</i>), cattails (<i>Typha latifolia</i>)	Jack pine (<i>Pinus banksiana</i> Lamb) with lichen, exposed soil understory	Black spruce (<i>Picea mariana</i>) with exposed soil, moss, herbs understory	Trembling aspen (<i>Populus tremuloides</i>) with dense hazel understory
Vegetation Type	Grass	Grass, wetland riparian vegetation	Evergreen needle-leaf	Evergreen needle-leaf	Deciduous broad-leaf
Stand density (trees ha⁻¹)			1320	4330	980
Canopy height			14 m	11 m	21 m
Soil Layer	Mineral soil: loam to clay loam	Mineral soil: dark brown loam to gravelly-loam soils	Mineral soil: Fine sand	Mineral soil: sandy clay Organic soil: peat	Mineral soil: loam to clay Organic soil: litter, fibric and humic
Drainage	Imperfect to poor	Imperfect to poor	Very well	Imperfect to poor	Well to moderately well

183

184 Table 1. Field site characteristics (Pan et al., 2017; Bam et al., 2019; Barr et al., 2012).

185

186

FIG 1 HERE

187 Figure 1. Field site locations (circles) and major cities (stars) in Saskatchewan, Canada.
 188 Ecoregions of central Saskatchewan are shown in shading, each aligning with different soils,
 189 landforms, and associated plant communities (Ecological Stratification Working Group,
 190 1996).

191

192 2.3 Instrumentation and data

193 In this study, meteorological data were used to drive the model and soil moisture data were
 194 used to assess the model performance and constrain the simulated fluxes. The specific
 195 instruments and measurement heights/depths are summarized in Table 2.

196
 197

Variable	Kenaston	SDNWA	OJP	OBS	OAS
Volumetric water content (m^3/m^3)	Stevens HydraProbe sensors	Stevens HydraProbe sensors	CS615 water content reflectometers	CS615 water content reflectometers	Time domain reflectometers
Depths (cm bgl)	0, 20, 50, 75, 100, 130, 160	0, 5, 20, 50, 100, 200, and 300	0-15, 15-30, 30-60, 60-90, 90-120, and 120-150	2.5, 7.5, 22.5, 45 and 60-90	0-15, 15-30, 30-60, 60-90, and 90-120
Air temperature ($^{\circ}\text{C}$) and Relative humidity (%)	Vaisala HMP45C 4 m above ground	Vaisala HMP45C 1.5 m above ground	Vaisala HMP45C 14 m above canopy	Vaisala HMP45C 14 m above canopy	Vaisala HMP45C 16 m above canopy
Wind speed (m/s)	CS CSAT3 tri-axial sonic anemometer 8 m above ground	CS CSAT3 tri-axial sonic anemometer 10 m above ground	CS CSAT3 tri-axial sonic anemometer 29 m above ground	Gill R3 or R3-50 tri-axial sonic anemometer 26 m above ground	Gill R3 or R3-50 tri-axial sonic anemometer 38 m above ground
Solar radiation (W/m^2)	Kipp and Zonen CNR1 four-component radiometer	Kipp and Zonen CNR4 four-component radiometer	Kipp and Zonen CM11 paired pyranometer 9 - 14 m above canopy	Kipp and Zonen CM11 Paired pyranometer 9 - 14 m above canopy	Kipp and Zonen CM11 paired pyranometer 10 - 16 m above canopy
Longwave radiation (W/m^2)	Kipp and Zonen CNR1 four-component radiometer	Kipp and Zonen CNR4 four-component radiometer	PIR paired pyrgeometer 9 - 14 m above canopy	PIR paired pyrgeometer 9 - 14 m above canopy	PIR paired pyrgeometer 10 - 16 m above canopy
Precipitation (mm)	Geonor T200-B weighing gauge	Geonor T200-B weighing gauge	Belfort 3000 accumulating	Belfort 3000 accumulating	Belfort 3000 accumulating

			gauge	gauge	gauge
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198
199 Table 2. Instrumentation at field sites (Pan et al., 2017; Barr et al., 2012; Bam et al., 2019)

200
201 The driving data are plotted in Figure 2, which shows the interannual, seasonal and diurnal
202 variation in these datasets. Precipitation is measured at each site with a weighing
203 precipitation gauge that captures both rainfall and snowfall. It is notable that the two prairie
204 sites, Kenaston and SDNWA, have higher windspeeds, which is because the forest sites are
205 more sheltered. As a result, the prairie sites are subject to significant redistribution of snow
206 laterally by the wind, and the snowfall observations from the gauges are not representative
207 of the snow on the ground, which could be higher or lower than the snowfall (Pan et al.,
208 2017). To correct for this, snow survey data from the prairie sites, which measure the peak
209 snow on the ground prior to melt, were used to adjust the snowfall data such that the snow
210 on the ground simulated by MESH would match the snow survey observations. For each
211 winter (typically November to February), every snowfall event was scaled by a constant
212 correction factor that was determined for each year by trial and error. The precipitation
213 data in Figure 2 have been corrected in this way for the prairie sites. Precipitation data at
214 the forest sites were not modified.

215
216 FIG 2 HERE

217 Figure 2. Driving data from meteorological stations. All variables are plotted as average
218 values per time increment dt , where dt is indicated in the plot titles for each column.

219
220 The calibration/validation soil moisture data are shown in Figure 3. Soil moisture
221 observations were available at different depths at different sites (Table 2). We used these
222 data to estimate the average volumetric water content in each of the soil depth increments
223 that correspond to the three layers in the model (i.e. 0 - 0.1 , 0.1 - 0.35 and 0.35 - 1.6 m
224 depths), as shown in Figure 3. The data are considered high quality, with minimal gaps or
225 data quality problems. Soil freezing is clearly evident at the prairie sites (Kenaston and
226 SDNWA), where the water content in the first two layers drops rapidly at the start of the
227 winter, and rises rapidly in the spring - this is because the water turns to ice, which is not
228 detected by the dielectric probes. The water content in layer 3 at SDNWA is much higher

229 than at Kenaston, which is because the water table is much closer to the ground surface at
230 SDNWA, and hence the soils are not as free-draining. The same is true of the OBS sites,
231 which is in a low-lying part of the landscape where again the water table is close to the
232 ground surface, and the upper layer has a high organic matter content and a high porosity,
233 with a lot of variation in storage. The OJP site has very sandy soil, which has poor water
234 retention properties, and has a deep water table, and hence the water contents are
235 consistently low, with a small range of variation seasonally. The OAS site has relatively fine
236 grained mineral soils that drain reasonably well.

237

238

239

FIG 3 HERE

240 Figure 3. Soil volumetric liquid water content observations from the five field sites, averaged
241 over the three model layer depth increments. The shaded cyan areas represent times of the
242 year when the observed soil temperature in layer 1 was below zero °C.

243

244

245 **2.4 Model description**

246 MESH (Modélisation Environnementale communautaire - Surface Hydrology) is a physically
247 based hydrological land-surface scheme built by Environment and Climate Change Canada
248 (ECCC) (Pietroniro et al. 2007). MESH is a configuration of the Modelling the Environment
249 Community (MEC) surface model that couples the Canadian Land Surface Scheme (CLASS)
250 (Bartlett et al., 2003, Verseghy, 2017) with hydrological routing scheme WATFLOOD
251 (Kouwen 1988, recently described in Pomeroy, 2016). MESH relies on a mosaic of Group
252 Response Unit's (GRUs) to represent the heterogeneity and hydrological processes of the
253 landscape. A GRU is a grouping of hydrological response units with similar soil and/or
254 vegetation attributes (Xu et al., 2017). In this study, we used a single grid cell with a single
255 GRU to represent the point scale vertical processes. MESH employs CLASS to simulate
256 vertical water fluxes and energy balances for each GRU (Verseghy, 2017). CLASS divides the
257 soil column into three layers and the vertical movement of water between each soil layer is
258 governed by a finite difference solution of one-dimensional Richards' equation for
259 unsaturated flow in porous media (Soulis, 2000). The hydraulic properties in CLASS adopt

260 the Clapp and Hornberger (1978) model to determine the relationship between hydraulic
 261 conductivity, K (m/d), matric potential, ψ (m), and soil moisture, where
 262

$$\psi = \psi_s \left(\frac{\theta}{\theta_p} \right)^{-b} \quad (8)$$

$$K = K_s \left(\frac{\theta}{\theta_p} \right)^{2b+3} \quad (9)$$

263 where ψ_s (m) is the saturated matric potential, θ_p is the saturated water content, K_s (m/d) is
 264 the saturated hydraulic conductivity, and b (-) is a shape parameter. Note, temperature
 265 corrections are applied to K_s to account for changes in the viscosity of water (Verseghy,
 266 2017, p. 122), and in frozen conditions an additional impedance factor is applied to reduce
 267 K_s to account for the ice blockages in the pore space (Verseghy, 2017, p. 146). The
 268 parameters in Equations 8 and 9 are normally determined using the empirical pedotransfer
 269 functions of Cosby et al. (1984), whereby
 270

$$\theta_p = \frac{-0.126 X_s + 48.9}{100.0} \quad (10)$$

$$b = 0.159 X_c + 2.91 \quad (11)$$

$$\psi_s = 0.01 e^{(-0.0302 X_s + 4.33)} \quad (12)$$

$$K_s = 0.60960384 e^{(0.0352 X_s - 2.035)} \quad (13)$$

271 Where X_s (%) and X_c (%) are the percentage sand and clay, respectively, of the soil in a
 272 particular layer. Note, in Equation 13 K_s is given in units of m/d. The pedotransfer function
 273 relationships and the resulting hydraulic properties are shown in Figure 4.

274

275

FIG 4 HERE

276 Figure 4. The pedotransfer functions that define soil hydraulic properties as a function of
 277 sand and clay content, and $\psi(\theta)$ and $K(\theta)$ relationships used in CLASS for three example
 278 soil textures.

279

280 The way that the fluxes in Equation 1 are calculated in CLASS are briefly described in Table 3.

281

Flux	Calculation description
Infiltration	Rainfall and throughfall on the ground are combined with snowmelt to form the potential infiltration flux; the infiltration capacity is calculated by a Green-Ampt model and depends on the soil hydraulic properties; water that cannot infiltrate forms ponding on the ground-surface, where it may later infiltrate unless the surface ponding capacity is exceeded, in which case the excess water forms overland runoff.
Drainage	The bottom of the soil profile, i.e. the base of soil layer 3, has a free drainage boundary condition, (Soulis et al., 2000), where $D = \phi K_3(\theta_3)$, where ϕ is a drainage parameter that restricts free drainage (0-1).
Soil evaporation	Soil evaporation is the sum of evaporation from bare soil and evaporation from soil below the canopy, both of which are driven by a humidity gradient. The humidity at the soil surface, and hence the soil evaporation flux, is reduced as the water content in layer 1 drops below field capacity, and soil evaporation is limited by the availability of water in the top soil layer and surface ponding (Sun and Verseghy, 2019).
Transpiration	Transpiration is extracted from all soil layers, weighted by the root density in each layer, as long as the liquid water content is greater than 0.04. The flux rate depends on the leaf-to-air humidity gradient, the boundary layer resistance and the canopy resistance, which in turn is related to leaf stomatal resistance and leaf area index. The stomatal resistance has a reference value, $r_{s,min}$, and is modified as a function of incoming solar radiation, vapour pressure deficit, soil moisture in the wettest layer, and air temperature. Through the stomatal conductance term transpiration rates are reduced exponentially as the soil suction in the wettest layer reduces below ψ_s , i.e. as the soil moisture reduces (Verseghy, 2017, Sun and Verseghy, 2019).

Table 3. Description of soil flux calculations in CLASS

282

283

284 2.5 Monte Carlo simulations

285 To investigate how information from soil moisture observations constrains simulated fluxes,

286 we apply a simple Monte Carlo approach. We allow for uncertainty in the model parameters

287 by randomly sampling parameter values from a uniform (or log-uniform) distribution. We

288 generate 10,000 parameter combinations and run the model with each parameter set over

289 a two-year calibration period. The performance of each realization is determined by

290 calculating ϵ_j , the root mean squared error (RMSE) of the cumulative change in liquid water

291 content (equation 6), for each depth, j , as in equation 14, and then averaging these over the

292 three layers, as in equation 15

293

$$\epsilon_j = \sqrt{\frac{\sum_{i=1}^n (\Theta_{o(i,j)} - \Theta_{s(i,j)})^2}{n}} \quad (14)$$

$$\epsilon_T = \sum_{j=1}^3 \frac{\epsilon_j}{3} \quad (15)$$

294

295 where $\Theta_{o(i,j)}$ and $\Theta_{s(i,j)}$ are the cumulative change in volumetric liquid water content from
 296 the observations and simulations, respectively, at time index i and depth index j , n is the
 297 number of points in time, and ϵ_T is the overall performance metric, with units of volumetric
 298 water content (m^3/m^3). The layer average RMSE was used, so that the performance in each
 299 layer would have an equal weighting – if the total storage was used the performance would
 300 be biased towards the 3rd soil layer, as this is much thicker (1.25 m) than the 1st (0.1 m) and
 301 2nd (0.25 m) layers.

302

303 As described in section 2.3, the CLASS model is typically run using the Cosby et al (1984)
 304 pedotransfer function, which determines the soil hydraulic properties from soil texture –
 305 specifically sand percentage, X_s and clay percentage X_c . We perform two separate
 306 calibration experiments here: the first using these pedotransfer functions samples the
 307 parameters X_s and X_c for each soil layer – i.e. six free parameters; the second does not use
 308 the pedotransfer functions, and samples the four hydraulic properties (θ_p , b , ψ_s , K_s) directly
 309 for each soil layer – i.e. twelve free parameters. Hereafter we describe these two
 310 experiments as the texture-based calibration and hydraulic properties-based calibration. In
 311 addition to the six/twelve soil parameters, we also sample four additional parameters that
 312 are understood to have a strong control on the simulated fluxes (Nazarbakhsh et al., 2020):
 313 the minimum and maximum annual leaf area index, L_{min} and L_{max} (-), the minimum stomatal
 314 conductance, $r_{s,min}$ (s/m) and a drainage index, ϕ (-) (Table 4). To prevent the random
 315 parameter sampling procedure from generating $L_{min} > L_{max}$ we instead sample L_{max} and the
 316 factor f_L such that $L_{min} = \min(L_{max}, L_{max} f_L)$. f_L is randomly sampled from a uniform
 317 distribution between 0.5 – 1.25, which ensures that around one-third of combinations will
 318 have $L_{min} = L_{max}$. For the texture sampling, we need to specify ranges in sand (X_s), clay (X_c)
 319 and silt (X_L) and ensure that the sum of the three equals 100%, which is done by randomly
 320 sampling X_s , X_c and X_L from uniform distributions (see ranges in Table 4), and then

321 rescaling each by the same scale factor such that $X_S + X_C + X_L = 100$. The texture-based and
322 hydraulic properties-based calibrations sample 10 and 16 free parameters, respectively. The
323 parameter ranges considered for each site were based on knowledge of the soils and
324 vegetation characteristics at each site, and are shown in Table 4.

325

	Kenaston	SDNWA	OJP	OBS	OAS
<i>Texture-based calibration parameters:</i>					
X_s (%)	20–70	20–70	45–100	20–70	20–70
X_L (%)	30–80	30–80	0–55	30–80	0–50
X_C (%)	0–60	0–60	0–55	0–60	5–35
<i>Hydraulic properties-based calibration parameters:</i>					
θ_p (-)	0.2–0.5	0.2–0.5	0.1–0.4	0.2–0.7	0.2–0.5
b (-)	3–18	3–18	3–18	3–18	3–18
ψ_s (m)	0.05–3	0.05–3	0.05–3	0.05–3	0.05–3
K_s (m/s)	$10^{-7} - 10^{-4}$	$10^{-7} - 10^{-4}$	$10^{-7} - 10^{-4}$	$10^{-7} - 10^{-4}$	$10^{-7} - 10^{-4}$
<i>Additional parameters:</i>					
L_{max} (-)	0.5–3	0.5–3	2–4	2–4	1–4
f_L (-)	0.5–1.25	0.5–1.25	0.5–1.25	0.5–1.25	0.5–1.25
$r_{s,min}$ (s/m)	50–300	50–300	50–300	50–300	50–300
ϕ (-)	0–1	0–1	0–1	0–1	0–1

327 Table 4. Parameter ranges considered. All parameters are sampled from a uniform
328 distribution except K_s which is sampled from a log-uniform distribution.

329

330 The models were all initialized on 1st August 2013, and the calibration period ends on 30th
331 September 2015 (two complete hydrological years) while the validation period ends on 30th
332 September 2017 (two additional complete hydrological years). Initializing the models in
333 August eliminates the need to specify initial soil ice content or the initial snowpack. The
334 initial water content of the model was based on observed estimates of the initial saturation,
335 combined with the current realization value of θ_p , i.e.

336

$$\theta_{j,ini} = \frac{\theta_{O,j,ini} - \min(\theta_{O,j})}{\max(\theta_{O,j}) - \min(\theta_{O,j})} \theta_{p,j} \quad (16)$$

337

338 where θ_o is the observed liquid volumetric water content, and subscripts j is the depth
339 index and t_{ini} is the initial time. This approach ensures that the initial relative saturation is
340 always the same, regardless of the porosity value that was sampled randomly in the Monte
341 Carlo simulation.

342

343 For the validation run, only the 30 best ranked parameter sets were considered (due to the
344 excessive computational expense of running 10,000 models for a 4 year period). We seek to
345 validate the model validation performance with the Ω observations, and to explore the
346 uncertainty in the model fluxes associated with these runs.

347

348 **3. Results and Discussion**

349 Figure 5 shows the performance of the model at each of the five sites, for the two sets of
350 Monte Carlo runs: i) sampling the soil texture parameters; and ii) sampling the soil hydraulic
351 properties. The results are shown in units of the cumulative change in storage in the profile
352 (i.e. $\Omega \dot{z}$). In the calibration period, the range of Ω from all 10,000 simulations is shown, along
353 with the range from the best 30 simulations, with the ranking based on ϵ_T (equation 14 and
354 15). Dotty plots for each simulation are provided in the Appendix. The model performance is
355 variable between the different sites, and between the texture based and hydraulic
356 properties-based calibration. The rise in storage in the spring (March-April) is characteristic
357 of seasonally frozen soils, and is a complex combination of snowmelt infiltration and soil
358 thaw (whereby ice becomes liquid water, leading to an apparent observed increase in liquid
359 water content, which may or may not be associated with an actual increase in total, that is
360 ice plus liquid, water content). The timing of this rise is delayed in all models, suggesting
361 there may be some limitations with either the snowmelt, the infiltration flux or the soil
362 thawing. However, the magnitude of the rise, and the overall seasonal changes in storage
363 are generally well captured in the models. The worst performance is at the OJP field site,
364 with the texture-based calibration. The reason this is poor is because the sandy soils at OJP
365 have a very small range of moisture variation, never rising as high as 0.2 (Figure 3). The
366 texture-based properties only allow for a variation in θ_p , the saturated water content, of
367 between 0.36 and 0.49, as shown in Figure 4 (top left), and therefore the model always
368 over-estimates the range of variation of water content. This restriction is removed using the
369 hydraulic properties-based calibration. This improvement in performance using the

370 hydraulic properties-based calibration is true for all field sites. This is because the
371 pedotransfer functions used in the texture-based calibration artificially constrain the agility
372 of the model, as discussed by Mendoza et al (2015). To elaborate, when using pedotransfer
373 functions, for a given parameter value of θ_p there will always be a unique parameter value
374 of K_s , and there is no way to explore deviations in this combination. With the hydraulic
375 properties-based calibration applied here, this constraint is removed, and θ_p and K_s are
376 treated as completely independent of one another. These two cases therefore represent
377 two extreme possibilities in terms of the possible parameter space that is explored by the
378 model.

379

380

FIG 5 HERE

381 Figure 5. Model calibration and validation performance in terms of cumulative change in
382 storage over the profile (mm) plotted for all 10,000 runs and the best 30 runs, ranked by ϵ_T

383

384 Figure 6 shows box plots of each of the soil water fluxes (Equation 1), plotted as annual
385 fluxes over the calibration period of the model, for both all simulations and the best 30. In
386 the box plots, the whiskers represent the complete range of the data, i.e. minimum to
387 maximum data points, which is taken as a simple measure of the overall uncertainty. The
388 range of the “all” plots represents the uncertainty of the simulated flux in the absence of
389 any constraints, while the range of the “best” plots represents the uncertainty of the
390 simulated flux when constrained by observed changes in soil moisture (as in Figure 5 and
391 the dotted plots in the appendix). If we consider each of the 30 best realizations as equally
392 credible, then the spread in the fluxes from these models is attributable to equifinality –
393 that is, different parameter sets that provide the same performance in one metric, but
394 provide different outcomes in terms of some other state or flux. The texture-based
395 calibration, which had a lower performance, appears to be slightly better at constraining the
396 fluxes than the hydraulic properties-based calibration. For example, at Kenaston, the range
397 of uncertainty in soil evaporation, E , is reduced from 300 mm/year (unconstrained) to 176
398 mm/year (constrained) using soil texture-based calibration. Using the hydraulic properties-
399 based calibration, uncertainty was reduced from 394 mm/year (unconstrained) to 327 mm/
400 year (constrained). We consistently see higher uncertainties when we calibrate the
401 hydraulic properties rather than the texture, which is expected because there are more

402 degrees of freedom in the hydraulic properties-based calibration. The infiltration flux has
403 the highest value and the lowest uncertainty of all of the fluxes, though the uncertainty
404 increases markedly for the hydraulic properties-based calibration. This is because the
405 infiltration capacity of the soil depends on the hydraulic properties directly, and these
406 properties are overly restricted in the texture-based calibration. At different sites for the
407 constrained (best) simulations the relative size of uncertainty in the soil evaporation,
408 transpiration and drainage does vary, but is in almost all cases too large to be useful. The
409 possible exception to this is at OBS for the texture-based calibration - where the fluxes are
410 most effectively constrained. Here the median/uncertainty range values for the constrained
411 fluxes are: I 443/27 mm/year; E 90/36mm/year; T 347/39 mm/year; D 0/1 mm/year. These
412 uncertainties are perhaps low enough that the model predictions could be considered
413 useful, but note that we are only reporting here the uncertainty based on soil moisture
414 observations, and an assessment of the performance of these fluxes against flux
415 observations is still needed.

416

417

418

FIG 6 HERE

419 Figure 6. Box plots of annual modelled soil fluxes including infiltration, I , soil evaporation, E ,
420 transpiration, T , and drainage, D , for all 10,000 runs and the best 30 runs, ranked by ϵ_T . The
421 whiskers represent the entire range of the data (i.e. minimum to maximum) and the boxes
422 represent the 1st, 2nd and 3rd quartiles.

423

424 Figure 7 shows cumulative fluxes of evapotranspiration (which includes soil evaporation plus
425 transpiration plus canopy evaporation and snow sublimation), runoff and drainage, for both
426 the calibration and validation period. As in Figure 6, this plot shows that soil moisture
427 overall is acting as a poor constraint on all of these fluxes. In the prairie sites (Kenaston and
428 SDNWA) we see evapotranspiration is the dominant flux; simulated runoff is associated with
429 the melt period; and drainage is relatively small but highly uncertain. At OJP we see the
430 largest uncertainties, particularly in the drainage fluxes that could be anywhere from zero to
431 > 200 mm/year. At OBS, where again the uncertainties are lowest overall, evaporation is
432 relatively well constrained, but drainage and runoff are still quite uncertain, especially in

433 2017. At OAS, the evaporation is somewhat well constrained, runoff is small, but drainage is
434 highly uncertain.

435

436 FIG 7 HERE

437 Figure 7. Simulated cumulative fluxes from the unconstrained (all 10,000) and constrained
438 (best 30) model runs for the calibration and validation periods.

439

440 4. Conclusions

441 We have shown that the MESH/CLASS model is capable of simulating the changes in soil
442 moisture within a 1.6 m deep profile at five diverse prairie/forest field sites relatively well,
443 albeit with some limitations associated with the timing of the rise in liquid water content
444 during the spring melt period. However, this relatively good performance at simulating
445 water content did not result in well constrained predictions of hydrological fluxes. From a
446 simple and qualitative assessment of Figure 6 and 7 we conclude that the information
447 content in soil moisture data is relatively low. The texture-based calibration is apparently
448 slightly better at constraining the fluxes than the hydraulic properties-based calibration,
449 which demonstrates that uncertainties in models can be reduced by embedding
450 assumptions within the models. However, this also requires that these embedded
451 assumptions are reasonable. In this case, the embedded assumption is that the hydraulic
452 properties can be determined from pedotransfer functions. Since we see in Figure 5 that
453 doing this degrades the performance of the model in reproducing observed changes in soil
454 moisture storage, this is not deemed a reasonable assumption in this case; the reduction in
455 uncertainty is considered misleading. We therefore conclude that the very wide uncertainty
456 bounds predicted by the hydraulic properties-based calibration are in fact a more accurate
457 reflection of the true uncertainty in the system. It is important to recognize that we are only
458 looking at one form of uncertainty here: uncertainty associated with parameters. There are
459 still other sources of uncertainty, most importantly uncertainty in the input data and
460 uncertainty in the model structure, that are not addressed here. We conclude that soil
461 moisture observations, while valuable in combination with other data, on their own are
462 inadequate for calibration of land surface models. Reproducing the cumulative change in
463 soil moisture storage is a necessary but insufficient criterion for model success. Uncertainty
464 is reduced by bringing in more information. If we had knowledge of the hydraulic properties,

465 for example from direct field observations, we would have a reduction in the parametric
466 uncertainty, resulting in less spread in the simulated fluxes. However, representative
467 parameter values are hard to obtain. Another valuable source of information comes from
468 observations of fluxes, which can be used to better constrain the model, and we
469 recommend that multi-objective calibration, using ET estimates from flux towers, will
470 improve the situation.

471

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479 Climate Change Canada for her advice about the Kenaston datasets. We thank Dan Princz
480 from Environment and Climate Change Canada for his excellent assistance with setting up
481 the MESH/CLASS model.

482

483 **Data Availability Statement**

484 On acceptance of the paper, all field data used will be uploaded to the Canadian Federated
485 Research Data Repository at: <https://www.frdr-dfdr.ca/repo/> and assigned a DOI number.

486

487 **References**

488 Bam, E. K. P., Brannen, R., Budhathoki, S., Ireson, A. M., & Spence, C. (2019). Meteorological,
489 soil moisture, surface water, and groundwater data from the St. Denis National Wildlife
490 Area, Saskatchewan, Canada. 11.

491 Barr, A. G., van der Kamp, G., Black, T. A., McCaughey, J. H., & Nesic, Z. (2012). Energy
492 balance closure at the BERMS flux towers in relation to the water balance of the White Gull
493 Creek watershed 1999–2009. *Agricultural and Forest Meteorology*, 153, 3–13.
494 <https://doi.org/10.1016/j.agrformet.2011.05.017>

495 Bartlett, P. A., Harry McCaughey, J., Lafleur, P. M., & Verseghy, D. L. (2003). Modelling
496 evapotranspiration at three boreal forest stands using the CLASS: Tests of parameterizations

497 for canopy conductance and soil evaporation. *International Journal of Climatology*, 23(4),
498 427–451. <https://doi.org/10.1002/joc.884>

499 Burns, T. T., Berg, A. A., Cockburn, J., & Tetlock, E. (2016). Regional scale spatial and
500 temporal variability of soil moisture in a prairie region: Regional Scale Soil Moisture
501 Variability. *Hydrological Processes*, 30(20), 3639–3649. <https://doi.org/10.1002/hyp.10954>

502 Clapp, R. B., & Hornberger, G. M. (1978). Empirical equations for some soil hydraulic
503 properties. *Water Resources Research*, 14(4), 601–604.
504 <https://doi.org/10.1029/WR014i004p00601>

505 Cosby, B. J., Hornberger, G. M., Clapp, R. B., & Ginn, T. R. (1984). A Statistical Exploration of
506 the Relationships of Soil Moisture Characteristics to the Physical Properties of Soils. *Water*
507 *Resources Research*, 20(6), 682–690. <https://doi.org/10.1029/WR020i006p00682>

508 Ecological Stratification Working Group. (1996). A national ecological framework for Canada.
509 Ottawa: Agriculture and Agri-Food Canada, Research Branch, Centre for Land and Biological
510 Resources Research; and Ottawa: Environment Canada, State of the Environment
511 Directorate, Ecozone Analysis Branch.

512 Gardner, C. M. K., Robinson, D., Blyth, K., & Cooper, J. D. (2003). Soil water content. In *Soil*
513 *and environmental analysis, physical methods*. Marcel Dekker.

514 Kouwen, N. (1988). WATFLOOD: A Micro-Computer Based Flood Forecasting System Based
515 on Real-Time Weather Radar. *Canadian Water Resources Journal*, 13(1), 62–77.
516 <https://doi.org/10.4296/cwrj1301062>

517 Li, Z., Liu, H., Zhao, W., Yang, Q., Yang, R., & Liu, J. (2019). Quantification of soil water
518 balance components based on continuous soil moisture measurement and the Richards
519 equation in an irrigated agricultural field of a desert oasis. *Hydrology and Earth System*
520 *Sciences*, 23(11), 4685–4706. <https://doi.org/10.5194/hess-23-4685-2019>

521 Maheu, A., Anctil, F., Gaborit, É., Fortin, V., Nadeau, D. F., & Therrien, R. (2018). A field
522 evaluation of soil moisture modelling with the Soil, Vegetation, and Snow (SVS) land surface
523 model using evapotranspiration observations as forcing data. *Journal of Hydrology*, 558,
524 532–545. <https://doi.org/10.1016/j.jhydrol.2018.01.065>

525 Mälicke, M., Hassler, S. K., Blume, T., Weiler, M., & Zehe, E. (2020). Soil moisture: Variable in
526 space but redundant in time. *Hydrology and Earth System Sciences*, 24(5), 2633–2653.
527 <https://doi.org/10.5194/hess-24-2633-2020>

528 Mendoza, P. A., Clark, M. P., Barlage, M., Rajagopalan, B., Samaniego, L., Abramowitz, G., &
529 Gupta, H. (2015). Are we unnecessarily constraining the agility of complex process-based
530 models? *Water Resources Research*, 51(1), 716–728.
531 <https://doi.org/10.1002/2014WR015820>

532 Miller, J. J., Acton, D. F., & St. Arnaud, R. J. (1985). The effect of groundwater on soil
533 formation in a morainal landscape in saskatchewan. *Canadian Journal of Soil Science*, 65(2),
534 293–307. <https://doi.org/10.4141/cjss85-033>

535 Nazarbakhsh, M., Ireson, A. M., & Barr, A. G. (2020). Controls on evapotranspiration from
536 jack pine forests in the Boreal Plains Ecozone. *Hydrological Processes*, 34(4), 927–940.
537 <https://doi.org/10.1002/hyp.13674>

538 Nijzink, R. C., Almeida, S., Pechlivanidis, I. G., Capell, R., Gustafssons, D., Arheimer, B.,
539 Parajka, J., Freer, J., Han, D., Wagener, T., Nooijen, R. R. P. van, Savenije, H. H. G., &
540 Hrachowitz, M. (2018). Constraining Conceptual Hydrological Models With Multiple
541 Information Sources. *Water Resources Research*, 54(10), 8332–8362.
542 <https://doi.org/10.1029/2017WR021895>

543 Pan, X., Helgason, W., Ireson, A., & Wheeler, H. (2017). Field-scale water balance closure in
544 seasonally frozen conditions. *Hydrology and Earth System Sciences*, 21(11), 5401–5413.
545 <https://doi.org/10.5194/hess-21-5401-2017>

546 Parsons, D. F., Hayashi, M., & van der Kamp, G. (2004). Infiltration and solute transport
547 under a seasonal wetland: Bromide tracer experiments in Saskatoon, Canada. *Hydrological*
548 *Processes*, 18(11), 2011–2027. <https://doi.org/10.1002/hyp.1345>

549 Peterson, A. M., Helgason, W. H., & Ireson, A. M. (2019). How Spatial Patterns of Soil
550 Moisture Dynamics Can Explain Field–Scale Soil Moisture Variability: Observations From a
551 Sodic Landscape. *Water Resources Research*, 55(5), 4410–4426.
552 <https://doi.org/10.1029/2018WR023329>

553 Peterson, Amber M., Helgason, W. D., & Ireson, A. M. (2016). Estimating field-scale root
554 zone soil moisture using the cosmic-ray neutron probe. *Hydrology and Earth System*
555 *Sciences*, 20(4), 1373–1385. <https://doi.org/10.5194/hess-20-1373-2016>

556 Pomeroy, J. W., MacDonald, M. K., Dornes, P. F., & Armstrong, R. (2016). Water budgets in
557 ecosystems. In *A Biogeoscience Approach to Ecosystems* (pp. 88–131). Cambridge University
558 Press.

559 Seneviratne, S. I., Corti, T., Davin, E. L., Hirschi, M., Jaeger, E. B., Lehner, I., Orlowsky, B., &
560 Teuling, A. J. (2010). Investigating soil moisture–climate interactions in a changing climate: A
561 review. *Earth-Science Reviews*, 99(3), 125–161.
562 <https://doi.org/10.1016/j.earscirev.2010.02.004>

563 Soulis, E. D., Snelgrove, K. R., Kouwen, N., Seglenieks, F., & Verseghy, D. L. (2000). Towards
564 closing the vertical water balance in Canadian atmospheric models: Coupling of the land
565 surface scheme class with the distributed hydrological model watflood. *Atmosphere-Ocean*,
566 38(1), 251–269. <https://doi.org/10.1080/07055900.2000.9649648>

567 Sun, S., & Verseghy, D. (2019). Introducing water-stressed shrubland into the Canadian Land
568 Surface Scheme. *Journal of Hydrology*, 579, 124157.
569 <https://doi.org/10.1016/j.jhydrol.2019.124157>

570 Vereecken, H., Huisman, J. A., Bogaen, H., Vanderborght, J., Vrugt, J. A., & Hopmans, J. W.
571 (2008). On the value of soil moisture measurements in vadose zone hydrology: A review.
572 Water Resources Research, 44(4). <https://doi.org/10.1029/2008WR006829>

573 Versegny, D. (2017). CLASS -- The Canadian Land Surface Scheme (Version 3.6.1).
574 Environment Canada.

575 Xu, X., Tolson, B. A., Li, J., & Davison, B. (2017). Assimilation of Synthetic Remotely Sensed
576 Soil Moisture in Environment Canada's MESH Model. IEEE Journal of Selected Topics in
577 Applied Earth Observations and Remote Sensing, 10(4), 1317-1327.
578 <https://doi.org/10.1109/JSTARS.2016.2626256>

579

580 **Appendix: Parameter identifiability**

581 Figures A1 – A10 present the dot plots for the texture-based and hydraulic properties-
582 based calibrations for each of the five sites. On the x-axes are plotted the parameter values
583 (refer to sections 2.4 and 2.5 for symbols), and on the y-axes are plotted the objective
584 function values, i.e. ϵ_T (Equations 14 and 15). For all sites, we see that the parameter
585 identifiability is better for the texture-based calibration than for the hydraulic properties-
586 based calibration. For the hydraulic properties-based calibration, the parameters are often
587 completely unidentifiable. One notable exception is at OJP (Figure A6) where the porosity is
588 identifiable and low values are clearly preferred. This poor identifiability is associated with
589 the equifinality in the model, and further shows that additional observations are needed to
590 constrain the parameters.

591

592

FIG A1 HERE

593 Figure A1. Dot plots for the Monte Carlo runs at Kenaston using the texture based soil
594 parameterisation. RMSE is the root mean squared error between the observed and
595 simulated cumulative change in water content for the unfrozen period, averaged between
596 the three layers, in units of volumetric water content.

597

598

FIG A2 HERE

599 Figure A2. Dotty plots for the Monte Carlo runs at Kenaston using the hydraulic properties
600 based soil parameterisation. RMSE is the root mean squared error between the observed
601 and simulated cumulative change in water content for the unfrozen period, averaged
602 between the three layers, in units of volumetric water content.

603

604

FIG A3 HERE

605 Figure A3. Dotty plots for the Monte Carlo runs at St Denis using the texture based soil
606 parameterisation. RMSE is the root mean squared error between the observed and
607 simulated cumulative change in water content for the unfrozen period, averaged between
608 the three layers, in units of volumetric water content.

609

610

FIG A4 HERE

611 Figure A4. Dotty plots for the Monte Carlo runs at St Denis using the hydraulic properties
612 based soil parameterisation. RMSE is the root mean squared error between the observed
613 and simulated cumulative change in water content for the unfrozen period, averaged
614 between the three layers, in units of volumetric water content.

615

616

FIG A5 HERE

617 Figure A5. Dotty plots for the Monte Carlo runs at OJP using the texture based soil
618 parameterisation. RMSE is the root mean squared error between the observed and
619 simulated cumulative change in water content for the unfrozen period, averaged between
620 the three layers, in units of volumetric water content.

621

622

FIG A6 HERE

623 Figure A6. Dotty plots for the Monte Carlo runs at OJP using the hydraulic properties based
624 soil parameterisation. RMSE is the root mean squared error between the observed and
625 simulated cumulative change in water content for the unfrozen period, averaged between
626 the three layers, in units of volumetric water content.

627

628

FIG A7 HERE

629 Figure A7. Dotty plots for the Monte Carlo runs at OBS using the texture based soil
630 parameterisation. RMSE is the root mean squared error between the observed and

631 simulated cumulative change in water content for the unfrozen period, averaged between
632 the three layers, in units of volumetric water content.

633

634 FIG A8 HERE

635 Figure A8. Dotty plots for the Monte Carlo runs at OBS using the hydraulic properties based
636 soil parameterisation. RMSE is the root mean squared error between the observed and
637 simulated cumulative change in water content for the unfrozen period, averaged between
638 the three layers, in units of volumetric water content.

639

640 FIG A9 HERE

641 Figure A9. Dotty plots for the Monte Carlo runs at OAS using the texture based soil
642 parameterisation. RMSE is the root mean squared error between the observed and
643 simulated cumulative change in water content for the unfrozen period, averaged between
644 the three layers, in units of volumetric water content.

645

646 FIG A10 HERE

647 Figure A10. Dotty plots for the Monte Carlo runs at OAS using the hydraulic properties based
648 soil parameterisation. RMSE is the root mean squared error between the observed and
649 simulated cumulative change in water content for the unfrozen period, averaged between
650 the three layers, in units of volumetric water content.