

1     **Changes in streamflow regimes and their response to different**  
2             **soil and water conservation measures in Loess Plateau**  
3                     **watersheds, China**

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19

## 20 **Abstract**

21 Investigating the changes in streamflow regimes is useful for understanding the mechanisms  
22 associated with hydrological processes in different watersheds and for providing information  
23 to facilitate water resources management. In this study, we selected three watersheds, i.e.,  
24 Sandu River, Hulu River, and Dali River on the Loess Plateau, to examine the changes in the  
25 streamflow regimes and to determine their responses to different soil and water conservation  
26 measures (terracing, afforestation, and damming). The daily runoff was collected  
27 continuously by three hydrological gauges close to the outlets of the three watersheds from  
28 1965 to 2016. The eco-surplus, eco-deficit, and degree of hydrological change were assessed  
29 to detect hydrological alterations. The Budyko water balance equation was applied to estimate  
30 the potential impacts of climate change and human activities on the hydrological regime  
31 changes. Significant decreasing trends ( $P < 0.05$ ) were detected in the annual streamflow in  
32 the Sandu and Dali River watersheds, but not in the Hulu River watershed where afforestation  
33 dominated. The annual eco-surplus levels were low and they decreased slightly at three  
34 stations, whereas the eco-deficit exhibited dramatic increasing trends in the Sandu and Dali  
35 River watersheds. In the Sandu River watershed (dominated by terraces), the runoff exhibited  
36 the most significant reduction and the eco-deficit was the highest among the three watersheds.  
37 The integral degrees of hydrological change were higher in the Sandu River watershed than  
38 the other two watersheds, thereby suggesting substantial variations in the magnitude, duration,  
39 frequency, timing, and rate of change in the daily streamflow. In the Dali River watershed  
40 (dominated by damming), the changes in the extreme flow were characterized by a decreasing  
41 number appearing in high flow. In these watersheds, human activities accounted for 74.1%

42 and 91.78% of the runoff reductions, respectively. In the Hulu River watershed (dominated by  
43 afforestation), the annual runoff exhibited an insignificant decreasing trend but with a  
44 significant increase in the low flow duration. Rainfall changes accounted for 64.30% of the  
45 runoff reduction.

46 **Keywords:** climate variability; human activities; indicators of hydrologic alteration (IHA);

47 streamflow regime

48

## 49 **1. Introduction**

50 Streamflows are essential for environmental health, economic prosperity, and human well-  
51 being. They provide the water required to produce energy, crops, and industrial products, and  
52 for maintaining terrestrial and aquatic environment systems (Curmi et al., 2013). The water  
53 resources in rivers determine the size, shape, structure, and dynamics of aquatic ecosystems.  
54 The streamflow regime in a river is considered to be mainly responsible for the variations in  
55 many other components of a river ecosystem, e.g., fish populations and nutrient cycling  
56 (Richter et al., 2003). Streamflow variability influences hydrological functions and it is  
57 important for maintaining biodiversity in rivers and the integrity of ecosystems (Arthington et  
58 al., 2006; Jovanovic et al., 2016; Vogel et al., 2007).

59 Recent studies have demonstrated that streamflow regimes have exhibited obvious  
60 changes in many rivers around the world. Approximately 24% of the world's large rivers  
61 appear to have exhibited significant changes in the water flux according to observations of the  
62 streamflow in 4399 rivers (Li et al., 2020). Forest clearing and large-scale agricultural  
63 activities have led to dramatic changes in the streamflow in Amazonian rivers over the last 40  
64 years (Aldrich et al., 2012; Latrubesse et al., 2009; Souza-Filho et al., 2016). Milly et al.  
65 (2002) predicted increased streamflows for some areas of equatorial Africa, the La Plata  
66 Basin, and high-latitude North America and Eurasia, but decreases in southern Africa,  
67 southern Europe, the Middle East, and mid-latitude western North America by the year 2050.  
68 The global mean annual temperature has increased by 0.8°C since 1880 (Flato et al., 2013),  
69 and the humidity and precipitation increased by around 2% in the last century (Huntington,  
70 2006; Wise, 2010), which may have altered the balance of the global water circulation and

71 energy cycle, thereby resulting in changes in the magnitude and spatiotemporal distribution of  
72 streamflows (Nam et al., 2015). Climate change affects streamflow regimes by increasing  
73 temperatures but also by altering precipitation patterns, rates of evaporation, transpiration, and  
74 soil moisture contents (Merritt et al., 2006). Human activities may alter streamflow regimes  
75 through land use changes, large-scale infrastructure such as reservoirs, water abstraction,  
76 urbanization, and ecological restoration projects (Dey et al., 2017; Liang et al., 2015; Tamm et  
77 al., 2018; Wang et al., 2016b). Both of these possible causes of changes in streamflow  
78 regimes have been investigated widely (Li et al., 2021; Ellis, 2011; Guo et al., 2019; Munoz et  
79 al., 2018; Zhang et al., 2017; Smith et al., 2016). It is necessary to explore streamflow  
80 variations and related factors to better understand the mechanisms associated with streamflow  
81 changes (Li and Fang, 2017; Wang et al., 2020), thereby providing useful reference data to  
82 facilitate climate adaptation and improved water resources management.

83 Hydrologic alterations represent the changes in streamflow regimes at different temporal  
84 and spatial scales due to regulation and water extraction via human activities. Over recent  
85 decades, numerous methods have been developed to assess the streamflow variations in  
86 rivers. The Newtonian and Darwinian approaches are currently applied widely from  
87 microscopic and macroscopic perspectives, respectively (Omer et al., 2020; Wang et al.,  
88 2020). Hydrological models (e.g., SWAT, TOPMODEL, and VIC) are mainly categorized as  
89 Newtonian approaches, and they are useful tools for hydrologic alteration assessment and  
90 decision making (Tamm et al., 2018). A representative Darwinian approach is the Budyko-  
91 based method, which has been widely applied to quantify how climate change and human  
92 activities might affect streamflow changes (Liang et al., 2015; Wang et al., 2019).

93 Many methods have been developed to assess streamflow regime changes. In particular,  
94 Richter et al. (1996) proposed the indicators of hydrologic alteration (IHA) method for  
95 quantifying streamflow regime changes using long-term observed daily flows in the Roanoke  
96 River in North Carolina, USA. Furthermore, the range of variability approach (RVA) was  
97 developed to meet the water requirements of ecosystems and achieve ecological sustainability,  
98 where the degree to which the shape of a river's natural flow regime can be altered is a well-  
99 established approach for quantifying flow regime alteration (Olden and Poff, 2003).

100 Many studies have demonstrated that hydrologic regimes are associated with dam  
101 construction (Yang et al., 2008). Lai et al. (2013) and Wang et al. (2016b) found that the  
102 operation of the Three Gorges Project aggravated hydrological droughts downstream, thereby  
103 leading to extremely low water levels and environmental flow deficits, with substantial effects  
104 on the streamflow regimes downstream of the Yangtze River. By contrast, Du et al. (2020)  
105 suggested that the construction of dams can stabilize hydrological regimes and reduce the  
106 flood peaks.

107 The Yellow River is one of China's largest rivers and the water it carries is utilized by  
108 8% of the population. The lower reaches of the Yellow River appeared to have zero flows for  
109 21 of the 27 years from 1972 to 1998, which contributed to the severe ecological damage to  
110 the river (Chen et al., 2020). Droughts, water shortages, and soil and water losses are the most  
111 severe environmental problems that have affected the social and economic development of the  
112 Yellow River basin (Wu et al., 2004). The Loess Plateau is located in the middle reaches of  
113 the Yellow River basin. Over 20 rivers flow into the Yellow River and they contribute nearly  
114 9% of the sediment and 40% of the streamflow into the river basin. Due to severe soil erosion,

115 various soil and water conservation measures have been implemented on the Loess Plateau  
116 since the 1970s (Zhao et al., 2014). The main measures comprise afforestation, check dam  
117 construction, and terracing, which have increased rainfall infiltration and reduced the flood  
118 peaks, soil erosion, and sediment transported into the rivers. Many studies have been  
119 conducted on the Loess Plateau to estimate the changes in runoff and the sediment load, as  
120 well as their potential causes (Gao et al., 2015; Guo et al., 2020; Zhang et al., 2017; Zhang et  
121 al., 2018; Zhou et al., 2020). These studies have clarified the effects of climate changes and  
122 human activities on runoff and sediment load reductions. However, few studies have assessed  
123 the alterations in the streamflow regimes and their responses to different soil and water  
124 conservation measures. Due to the limited availability of information regarding conservation  
125 measures, few studies have also compared the hydrological responses of river flows to  
126 individual conservation measures. Therefore, the objectives of this study were: (1) to examine  
127 the changes in the streamflow regimes in three watersheds with different conservation  
128 measures; and (2) to quantify the effects of climate variability and human activities on the  
129 streamflow changes.

## 130 **2. Study area and data**

### 131 **2.1. Study area**

132 More than 20 tributaries in the middle reaches of the Yellow River flow through the  
133 Loess Plateau and they discharge approximately 90% of the sediment that enters the Yellow  
134 River. Since the 1950s, extensive soil and water conservation measures have been  
135 implemented in these watersheds to control severe soil erosion, including afforestation, grass

136 planting, terracing, and check dam construction. According to the spatial distribution and  
137 types of soil and water conservation measures on the Loess Plateau, we selected the following  
138 three different watersheds and analyzed the changes in their streamflow regimes: Sandu River  
139 watershed (dominated by terracing), Hulu River watershed (affected by afforestation), and  
140 Dali River watershed (dominated by check dams). Figure 1 shows the locations of the  
141 watersheds and hydro-meteorological gauges.

### 142 **2.1.1. Sandu River watershed**

143 The Sandu River watershed is located upstream of the Wei River basin and it covers an  
144 area of 2484 km<sup>2</sup>. This watershed is characterized by a continental monsoon climate, with a  
145 mean annual temperature around 10.2°C and average annual precipitation of approximately  
146 467.3 mm. The average annual streamflow and sediment yields are 0.45 billion m<sup>3</sup> and 8560 t/  
147 km<sup>2</sup>, respectively. Since the 1970s, many terraces have been constructed on the hill slopes.  
148 Terraces with a total area of 1807.45 km<sup>2</sup> were built by 2017, which covered 72.7% of the  
149 whole watershed.

### 150 **2.1.2. Hulu River watershed**

151 The Hulu River is a tributary of the Beiluo River (a tributary of the Wei River, Figure 1),  
152 with a drainage area of 4715 km<sup>2</sup>. The mean annual precipitation and evaporation in this  
153 watershed are 494 mm and 1147 mm, respectively. The streamflow from July to September  
154 accounts for about 50% of the annual total. The catchment is covered by dense forest, where  
155 *Pinus tabuliformis* Carr. and *Platycladus orientalis* (Linn.) Franco are the major forest species  
156 present in this watershed. More than 90% of the sloping arable land was converted into forest  
157 or grassland after 1999 as part of the “Grain for Green” project launched by the Chinese

158 government. The vegetation cover has reached approximately 80% and the sediment yield  
159 decreased significantly from  $107 \times 10^4$  t in the 1960s to  $76.8 \times 10^4$  t in 2016.

### 160 **2.1.3. Dali River watershed**

161 The Dali River is the largest tributary of the Wuding River, with an area of 3906 km<sup>2</sup> and  
162 length of 170 km. The average annual streamflow was 1.31 billion m<sup>3</sup> from 1960 to 2016. The  
163 average annual precipitation is 429 mm and high-intensity rainstorms cause considerable soil  
164 erosion in this watershed during the summer. Since the beginning of the 1980s, many check  
165 dams have been built in the watershed to trap sediment transported from hill slopes to the  
166 Yellow River. In 2017, 11602 check dams were present in the Wuding watershed, i.e., 1155  
167 large sized check dams and 10174 medium to small sized check dams (Han et al., 2018). Due  
168 to the implementation of large-scale soil and water conservation measures in the Dali River  
169 watershed, the sediment load decreased from  $0.64 \times 10^8$  t/a in the 1970s to  $0.15 \times 10^8$  t/a from  
170 2000 to 2016 (Zhang et al., 2019).

171

172 <Figure 1>

173

## 174 **2.2. Data set**

175 The daily streamflows at the Gangu (Sandu River), Zhangcunyi (Hulu River), and Suide  
176 (Dali River) hydrological stations (Figure 1) were obtained from the Hydrological Year Book  
177 of the Yellow River, which was published by the Ministry of Water Resources of China. In  
178 total, the daily discharge measurements for 52 years from 1965 to 2016 were collected for  
179 investigation. Table 1 shows general information for the three stations in the watersheds.  
180 Monthly rainfall, air temperature, relative humidity, sunshine duration, and wind speed

181 meteorological data at 11 climate stations were obtained from the China National Climate  
182 Center. The potential evapotranspiration was calculated using the Penman–Monteith equation  
183 (Allen et al., 1998). The basin-averaged precipitation and potential evapotranspiration data  
184 were estimated, and then interpolated with the inverse distance weighted method in ArcGIS  
185 10.6 (<http://www.esri.com>). The spatial distributions of soil and water conservation measures,  
186 including terraces, check-dams, and reservoirs, in each catchment during 2017 were  
187 interpreted with Google Earth and validated based on a field survey. The homogeneity and  
188 continuity of all the measured data were checked to guarantee the data integrity and  
189 consistency before their release.

190

191

<Table 1>

192

### 193 **3. Methodology**

194 The Mann–Kendall test and accumulated anomaly method were applied to detect the  
195 changing trends and abrupt changes in the annual streamflow at the three hydrological  
196 stations. These methods have been used widely to examine changing hydro-meteorological  
197 time series (Kendall, 1975; Mann, 1945; Sagarika et al., 2014; Zhao et al., 2014), so they are  
198 not described in detail. The detailed estimation procedures were described in previous studies  
199 (Weber et al., 2010; Zhao et al., 2019). The IHA method (Richter et al., 1996) was employed  
200 to analyze the streamflow regime changes according to the daily streamflow, and the effects  
201 of climate change and human activities on streamflow variations were assessed by using the  
202 Budyko equation (He et al., 2019; Yang et al., 2014).

### 203 3.1. Changes in hydrological regimes

204 The IHA method was first developed by the Nature Conservancy and it has been applied  
205 widely. The IHA method employs 32 parameters (we excluded “the number of zero-flow  
206 days” because zero flows were never observed at the three hydrological stations) related to  
207 hydrological extremes and averages, which can be classified according to five major  
208 categories (Table 2).

209 The RVA method proposed by Richter et al. (1996) was used to quantify the hydrological  
210 changes in terms of the 32 indicators. The target range for the RVA comprises IHA values that  
211 fall within the thresholds between the 25th and 75th percentile values (Huang et al., 2019).  
212 The degree to which the RVA does not reach the target range denotes the degree of  
213 hydrological change, and can be expressed as ( $D_i$ ) for each indicator. Cheng et al. (2019) and  
214 Zhang et al. (2016) described the calculation method in detail. Therefore, the following three

215 equal sized categories were employed: (1)  $(|D_i| < 33\%)$  indicating little or no alteration; (2)

216  $(33\% < |D_i| < 67\%)$  denoting a moderate degree of alteration; and (3)  $(67\% < |D_i| < 100\%)$

217 representing a large change.

218 The nondimensional eco-flow metrics comprising the eco-deficit and eco-surplus were  
219 estimated based on the flow duration curve (FDC) using the daily flow data (Vogel et al.  
220 2007) in order to provide intuitive representations of the hydrological impacts and to  
221 supplement those characterized by the IHA. The annual FDC can be regarded as a function of  
222 the excess of the daily streamflow over the probability in each year. The eco-flow metrics

223 were estimated in the following four steps. (1) According to the change point, the daily  
224 streamflow series covering the period from 1965 to 2016 was subdivided into two sections. In  
225 the baseline period, human activities were regarded as having limited effects on the  
226 streamflow regimes. (2) The 25% and 75% FDC quantiles were taken as threshold values, and  
227 the values in the intervals were the adaptive range for the river ecosystem. (3) The eco-flow  
228 was determined by comparing the annual FDC in other years with the previously obtained  
229 25% and 75% FDC quantiles. The area above the 75% quantile FDC was the eco-surplus,  
230 whereas the area below the 25% quantile FDC was the eco-deficit. Full details of the  
231 calculation procedure were provided by Gao et al. (2012) and Vogel et al. (2007).

232

233

<Table 2>

234

### 235 **3.2. Impacts of climate variability and human activities**

236 The Budyko equation is important in hydrology because it provides a concise and  
237 accurate representation of the relationship between the annual evapotranspiration and long-  
238 term average water and energy balance at catchment scales (Sposito, 2017). In a natural basin,  
239 the long-term average annual water and energy balance at the catchment scale can be  
240 expressed as follows.

241

$$R = P - E - \Delta S \quad (3)$$

242 The Budyko hypothesis (Budyko, 1974) considers the balance for precipitation ( $P$ ) between  
243 potential evapotranspiration ( $E_0$ ) and actual evapotranspiration ( $E$ ). Based on the long-term  
244 catchment water balance equation,  $\Delta S$  is assumed to be zero. By combining dimensional  
245 analysis and physical principles, Fu (1981) analytically derived the water–energy balance

246 function at the mean annual time scale, which is expressed as:

$$247 \quad \frac{E}{P} = 1 + \frac{E_0}{P} - \left[ 1 + \left( \frac{E_0}{P} \right)^\omega \right]^{\frac{1}{\omega}}, \quad (4)$$

248 where  $E_0$  is the mean annual potential evapotranspiration and the parameter  $\omega$  represents the

249 catchment landscape characteristics. The long-term  $\frac{E}{P}$  is mainly controlled by the water-

250 energy balance  $\frac{E_0}{P}$  (the dryness index). Changes in the catchment streamflow can be

251 expressed as the sum of three components defined by Schaake (1990) as the precipitation,

252 potential evapotranspiration, and catchment landscape elasticity of the streamflow, and thus

253 the new equation is expressed as:

$$254 \quad \Delta R = \varepsilon_P \frac{R}{P} \Delta P + \varepsilon_{E_0} \frac{R}{E_0} \Delta E_0 + \varepsilon_\omega \frac{R}{\omega} \Delta \omega = \Delta R_P + \Delta R_{E_0} + \Delta R_\omega, \quad (5)$$

255 where the elasticities of the streamflow are given as:

$$256 \quad \varepsilon_P = \frac{\left[ 1 + \left( E_0 / P \right)^\omega \right]^{\frac{1}{\omega} + 1} - \left( E_0 / P \right)^{\omega + 1}}{\left[ 1 + \left( E_0 / P \right)^\omega \right] \left\{ \left[ 1 + \left( E_0 / P \right)^\omega \right]^{\frac{1}{\omega}} - \left( E_0 / P \right) \right\}} \quad (6)$$

$$257 \quad \varepsilon_{E_0} = \frac{1}{\left[ 1 + \left( E_0 / P \right)^\omega \right] \left\{ 1 - \left[ 1 + \left( E_0 / P \right)^\omega \right]^{\frac{1}{\omega}} \right\}} \quad (7)$$

$$\varepsilon_{\omega} = \frac{\ln [1 + (E_0 / P)^{\omega}] + (E_0)^{\omega} \ln [1 + (E_0 / P)^{-\omega}]}{\ln [1 + (E_0 / P)^{\omega}] \left\{ 1 - [1 + (E_0 / P)^{-\omega}]^{\frac{1}{\omega}} \right\}}, \quad (8)$$

258

259 where  $\omega$  is a model parameter denoting the non-climatic effects on the water–energy balance  
 260 attributed to the soil properties, topography, and vegetation (Gunkel et al., 2017; Wang et al.,  
 261 2016a).

262 Changes in the observed mean annual runoff depth  $\Delta R^T$  can be estimated between the  
 263 baseline period and changing period, and they can be attributed to climate variability  $\Delta R^C$  a  
 264 nd human activities  $\Delta R^k$ . The streamflow change due to climate variation ( $\Delta R^C$ ) includes the  
 265 streamflow changes due to precipitation variation ( $\Delta R_p$ ) and potential evaporation variation ( $\Delta R_{E_0}$ ) (Koster & Suarez, 1999; Milly & Dunne, 2002). The contributions to annual

266 streamflow changes due to climate change and human activities, respectively, can be  
 267 approximated as follows.

269

$$\eta_C = \frac{\Delta R^C}{|\Delta R^C| + |\Delta R^k|} \times 100\% \quad (11)$$

270

$$\eta_k = \frac{\Delta R^k}{|\Delta R^C| + |\Delta R^k|} \times 100\% \quad (12)$$

271

## 272 **4. Results**

### 273 **4.1. Changes in annual streamflows**

#### 274 **4.1.1. Temporal variations in annual streamflows**

275 Figure 2 shows the linear trends in the annual streamflows in the three watersheds.  
276 Overall, the annual streamflows tended to decrease at all stations, with relatively high  
277 reductions at Gangu station (0.159 mm/10a) and Suide station (0.138 mm/10a) during 1965–  
278 2016. For example, at Gangu station, the annual average streamflow was only  $0.21 \times 10^8 \text{ m}^3/\text{a}$   
279 from 1995 to 2016, which was much lower than that during 1965–1994 ( $0.629 \times 10^8 \text{ m}^3/\text{a}$ ).  
280 The Mann–Kendall test showed that the annual streamflows at Gangu (Figure 2a) and Suide  
281 stations (Figure 2c) decreased significantly ( $P < 0.01$ ), but not significantly at Zhangcunyi  
282 station (Figure 2b). Comparisons of the annual streamflow fluctuations at the three stations  
283 showed that the variability in the annual streamflow was higher at Zhangcunyi ( $C_V = 0.34$ ).

284

285

<Figure 2>

286

#### 287 **4.1.2. Abrupt changes in annual streamflows**

288 As shown in Figure 3, abrupt changes in the annual streamflows mostly occurred during  
289 the mid-1990s and they were mainly attributable to the large-scale soil and water conservation  
290 measures implemented in the middle reaches of the Yellow River. Abrupt changes were  
291 detected during 1994 at Gangu station (Figure 3a), 1990 at Zhangcunyi station (Figure 3b),  
292 and 1996 at Suide station (Figure 3c). Therefore, the total time series at each station were  
293 divided into two periods based on these breakpoints. The first period represented the baseline

294 period with very limited human activities or none. The second period represented the change  
295 period when the watershed experienced substantial changes in land use,  
296 afforestation/deforestation, and dam construction.

297

298

<Figure 3>

299

## 300 **4.2. Changes in IHA metrics**

301 The IHA indicators in group 1 (G1) represent the magnitudes of the average monthly  
302 median streamflow, which clearly varied at the three gauging stations. Significant decreases  
303 were found in all of the areas, but particularly at Gangu station, with average change rates  
304 lower than  $-50\%$  ( $P < 0.001$ ). The median 1-, 3-, 7-, 30-, and 90-day annual  
305 minimum/maximum flows (Group 2, G2) tended to decrease. The highest reduction in G2 was  
306 detected at Gangu station and the lowest at Zhangcunyi station. In addition, the 1- and 3-day  
307 minimum flows increased at Suide station but not significantly. The Group 3 (G3) indicators  
308 represent changes in the timing of extreme flows. The minimum and maximum dates tended  
309 to increase at all three stations, thereby suggesting time lag effects of the soil and water  
310 conservation measures. In particular, at Gangu station, the low flow date changed most  
311 greatly and reached up to 90.9%. In the Group 4 (G4), the durations of the low pulses tended  
312 to increase at all stations during the changing period, but especially at Gangu and Zhangcunyi  
313 stations, with higher change rates ( $P < 0.05$ ) that increased from 3 to 6.5 days and 3.5 to 7.25  
314 days, respectively. The durations of the high pulses remained relatively stable at all stations.  
315 Different trends were found at Suide station where the median duration of high pulses during

316 the change period decreased to 2.5 days and by 9.09% compared with that of 2.75 days before  
317 the baseline period. The flow fall rates all decreased significantly at three stations ( $P < 0.05$ ).

318 The variations in the hydrologic indicators tended to differ among the stations, as shown  
319 in Figure 4b. We found that 13 indicators exhibited moderate alterations at Suide station, 18  
320 indicators had low-degree alterations at Zhangcunyi station, and 75% of the indicators had  
321 high degree alterations at Gangu station. Clearly, the fall rates exhibited the greatest changes  
322 ( $|Di| > 67%$ ) in the three watersheds and they indicated declines in the streamflows, where  
323 more of the conditions in the post-period were below the lower limit of the RVA threshold  
324 than those in the pre-period. At Suide station, the  $Di$  value was related to the flood season.  
325 Indicators such as the duration of low pulses ( $D_i = -78.57%$ ), baseflow index ( $D_i = -70%$ ),  
326 and number of reversal stations were assigned to the high-degree alteration category at  
327 Zhangcunyi.

328

329

<Figure 4>

330

### 331 **4.3. Changes in eco-flow metrics**

332 In Figure 5, the blue and red curves correspond to the 25th percentile and 75th percentile  
333 FDC, respectively, during the baseline period at each station. Compared with the daily  
334 streamflow indices in the baseline period, the high and low flows decreased significantly at  
335 the three stations, where the reductions in the low flow ( $Q_{90}$ ) rates were lower than those in  
336 the high flow ( $Q_{10}$ ) rates at Gangu and Zhangcunyi stations. In particular, the low flow  
337 declined greatly by 69.57% at Gangu station. By contrast, the daily streamflow remained  
338 relatively stable at Zhangcunyi station, where the  $Q_{10}$  and  $Q_{90}$  components decreased by

339 23.48% and 18.4%, respectively. At Suide station, the reduction in the low flow component  
340 (44.64%) was higher than that in the high flow (23.33%) during the changing period.

341

342 <Figure 5>

343

344 Figure 6 shows the annual eco-flow metrics (eco-surplus and eco-deficit) obtained based  
345 on the annual FDC and temporal variations in the annual precipitation anomaly at the three  
346 stations. Overall, the annual eco-surplus and annual eco-deficit tended to fluctuate greatly.  
347 Remarkably, the variations in the annual eco-deficit were more substantial than those in the  
348 annual eco-surplus. Figures 6b, 6d, and 6e show the eco-surplus and eco-deficit results for the  
349 three stations in different decades. From the 1990s, the annual eco-deficit increased  
350 dramatically at Gangu station (a), but there was no apparent variation at Zhangcunyi (b) and  
351 only a slight increase at Suide station (c). The eco-surplus tended to decrease at the three  
352 stations. At Gangu station, persistent and high peaks were detected in the eco-deficit and eco-  
353 surplus during the early baseline period and later changing period. In addition, the negative  
354 deviation in the precipitation explained the eco-deficit during 1994–2002. At Zhangcunyi  
355 station, the eco-flow metrics varied consistently with the changes in precipitation, thereby  
356 suggesting that afforestation did not change the eco-flow metrics in the watershed and climate  
357 change may have contributed more to the streamflow changes. At Suide station, the eco-flow  
358 metrics and precipitation were strongly correlated. During the changing period, the eco-deficit  
359 increased because of the implementation of soil and water conservation measures, although  
360 the precipitation increased. According to Figure 6, the low flow rate contributed more to the

361 eco-surplus whereas the high flow rate contributed more to the eco-deficit. These results are  
362 similar to those reported by Du et al. (2020).

363

364 <Figure 6>

365

#### 366 **4.4. Attribution of streamflow variations to climate variability and** 367 **human activities**

368 Significant reductions in the annual streamflow could be attributed to climate change and  
369 human activities. We employed the Budyko equation to quantify the effects of climate change  
370 and land surface changes on the streamflow variations. After comparing the annual runoff  
371 rates in the changing period and baseline period, we found that the streamflow decreased by  
372 16.83 mm, 4.16 mm, and 9.95 mm at Gangu, Zhangcunyi, and Suide stations, respectively.  
373 Thus, the annual streamflow reduction was lowest at Zhangcunyi station in the forest  
374 dominated watershed. As shown in Table 3, climate change only accounted for 25.9% and  
375 8.22% of the reductions in the Sandu and Dali River watersheds, respectively. Human  
376 activities were mainly responsible for the runoff reductions, particularly in the Dali River  
377 watershed (91.78%). In contrast to the Sandu and Dali River watersheds, the streamflow  
378 reduction in the Hulu River basin was attributed primarily to climate change (64.30%), and  
379 human activities (mainly afforestation) were only responsible for the other 35.70%.

380

381 <Table3>

382

## 383 **5. Discussion**

### 384 **5.1. Impacts of climate change and human activities on streamflows**

385 Using hydro-climatic data, the elasticity/sensitivity method based on the Budyko  
386 equation has been widely applied to quantify the effects of climate variability and human  
387 activities on streamflow changes. Wang et al. (2020) found that human activities accounted  
388 for over 50% of the runoff reductions in four catchments in the Yellow River basin from 1960  
389 to 2015. Gu et al. (2019) showed that precipitation and human activities contributed  
390 approximately 20% and 80%, respectively, to the runoff reductions in both river sources and  
391 the middle reaches of the Yellow River basin. Similar results were obtained in our analyses of  
392 watersheds dominated by terraces and check dams, but different results in the afforested  
393 watershed (Hulu River watershed).

394 To further verify our results, we applied the double mass curve method to quantify the  
395 effects of climate change and human activities on the runoff changes (Chang et al., 2015; Gao  
396 et al., 2017). Figure 7 shows that changes were found in the cumulative annual runoff curves  
397 and precipitation at Suide and Gangu stations, thereby indicating that the runoff changes were  
398 more significant in these two watersheds. The estimates indicate that human activities  
399 accounted for 86.85% of the runoff reduction at Gangu station and 97.11% at Suide station.  
400 For the Hulu River watershed, human activities accounted for 35.70% of the streamflow  
401 reduction and the remaining 64.30% was attributed to climate change. Thus, the results  
402 obtained using the double mass curve were consistent with those produced by the Budyko  
403 method.

404

405

<Figure 7>

406

407 **5.2. Impacts of soil and water conservation measures on streamflow**  
408 **regime changes**

409 Soil and water conservation measures have been implemented in the upper and middle  
410 reaches of the Yellow River basin since the late 1950s (Zhang et al., 2018). These measures  
411 include biological measures comprising afforestation and grassing, and engineering measures  
412 comprising terracing and check dams. Approximately 58000 check dams have been built,  
413 including 5546 large sized dams. The area covered by terraces is  $5.5 \times 10^4$  km<sup>2</sup>, mostly  
414 upstream of the Wei River and in the middle reaches of the Yellow River basin.

415 The proportion of vegetation cover in the study area was 28.6% in the 1980s and it  
416 increased to 63.2% by 2018. These obvious changes in the land surface cover were  
417 responsible for the significantly reduced river streamflow. Vegetation plays a vital role in  
418 regulating terrestrial water flows, where forests can fix and store carbon as well as regulating  
419 water functionalities (Bai et al., 2020; Ellison et al., 2017; Farooqi et al., 2020). The “Grain  
420 for Green” project launched in 1999 greatly increased the vegetation cover and reduced soil  
421 erosion on the Loess Plateau (Zhou et al., 2015). Figure 8 shows the vegetation cover at  
422 different levels (high, medium, and low vegetation cover) in the three watersheds. Changes in  
423 the medium and high vegetation cover levels occurred in the three watersheds. In general, the  
424 Hulu River watershed had a high vegetation cover rate, and transformations occurred from  
425 medium vegetation to high vegetation cover in the other two watersheds from 1998 to 2016  
426 (Figures 8b and 8c). These results are consistent with previous reports of great increases in the

427 vegetation cover in the middle reaches of the Yellow River basin (Wang et al., 2019), and the  
428 vegetation cover has increased significantly since 2002 (Xin et al. 2008).

429

430

<Figure 8>

431

432 The hydrological responses to afforestation differed among the three river basins. We  
433 found significant reductions in the runoff in the Sandu and Dali River watersheds, and a  
434 constant increase in the eco-deficit values after the abrupt change. The results suggest that  
435 engineering measures had more significant effects on runoff reduction. A non-significant  
436 decrease in the annual runoff was found in the forest-dominated basin (Hulu River basin),  
437 thereby confirming that afforestation affected the river streamflow by increasing terrestrial  
438 interception and evapotranspiration (*ET*). Moreover, large differences in the eco-flow metrics  
439 and degree of hydrologic alteration according to most of the IHA indicators were found  
440 between the Hulu River basin and other watersheds. Previous studies by Zhou et al. (2015),  
441 Ellison et al. (2017), and Evaristo et al. (2019) showed that afforestation can effectively  
442 moderate floods by storing or recycling substantial amounts of water via interception,  
443 infiltration, transpiration, evaporation, and groundwater recharge. Vegetation cover can help  
444 to maintain a low flow by moderating the streamflow and conserving water. The change in  
445 precipitation was also an important factor that affected the variations in vegetation cover.  
446 Thus, the relatively stable streamflow regime in the Hulu River basin was caused by  
447 afforestation.

448 However, compared with the changes in the annual runoff in the humid region, the dry  
449 region tended to decrease due to both climate variability and afforestation. Zhou et al. (2020)

450 reported that the water conserving effects of afforestation were higher than those of increased  
451 terrestrial interception and evapotranspiration in a humid region. However, this is the opposite  
452 of the effect found in dry areas, such as the Hulu River watershed. Afforestation can increase  
453 low flows and reduce high floods through more interception and infiltration, but it also results  
454 in higher evapotranspiration. This may explain the decreased runoff at Zhangcunyi station.

455

456

<Figure 9>

457

458 Terraced fields are essential measures for reducing the transport of soil eroded from  
459 upstream to downstream areas by reshaping the microtopography and increasing the slope  
460 length (Tarolli et al., 2014), as demonstrated by Zhang et al. (2008) and Chen et al. (2017). An  
461 experimental study conducted by Ran et al. (2006) suggested that terraces could reduce the  
462 runoff by 60.7% and sediment yield by 58.0% in the Wei River Basin. In the Sandu River  
463 basin, about 72.7% of the watershed is covered by terraces (Table 4), which is much higher  
464 than the coverage rates found in the Dali and Hulu River basins. The change in the  
465 streamflow regime in the Sandu River basin differed from those in the other watersheds. The  
466 annual streamflow and eco-deficit decreased most in the Sandu River basin watershed,  
467 thereby demonstrating the significant effect of terraces on the surface runoff. The Budyko  
468 method indicated that the evaporation was 40 mm higher during the changing period than the  
469 baseline period. Moreover, the duration of the low flow increased, whereas the number of  
470 high flows clearly decreased.

471

472

<Table 4>

473

474 Check dams were initially built to trap sediment and produce fertile agricultural land in  
475 the gully-dominated region of the Loess Plateau. According to Li et al. (2016), check dams  
476 are vital engineering measures for retaining flood water, trapping upstream sediment,  
477 increasing the availability of farmland for agricultural production, and reducing downstream  
478 sediment transport. In the Dali River watershed, 46.7% of the total area was controlled by  
479 check dams and reservoirs in 2017. By contrast, far fewer check dams were present in the  
480 other two watersheds. Dams and reservoirs can effectively reduce flood peaks and increase  
481 their slow time, as indicated by the IHA metrics. Dam construction significantly increases the  
482 seepage of soil water into groundwater, thereby leading to an increase in low flows.  
483 According to Martin-Rosales et al. (2007), check dams increased the infiltration of runoff by  
484 3–50% in a semiarid region of Spain. Unlike vegetation measures, engineering measures can  
485 immediately regulate surface runoff, especially in high flow events. Furthermore, the  
486 hydrological alteration degree in the Dali River watershed changed significantly in terms of  
487 the streamflow regime among the watersheds, thereby demonstrating the strong influence of  
488 dams on hydrological processes.

489 Previous studies employed the IHA/RVA approach to analyze the hydrological  
490 alterations in the middle reaches of the Yellow River. Zhang et al. (2016) determined the  
491 effects of reservoir construction and operation in the upper Yellow River basin on alterations  
492 in the ecological flow regimes, and found decreases in high flows and increases in low flows  
493 at mainstream stations. Indicators of low flow and rising and falling water conditions changed  
494 greatly, and they were consistent with our results. However, we obtained more detailed results

495 regarding the streamflow regime changes related to different soil and water conservation  
496 measures.

497

## 498 **6. Conclusion**

499 In this study, we applied the IHA/RVA method and eco-flow metrics to estimate  
500 streamflow regimes based on the daily discharge in three watersheds where different soil and  
501 water conservation measures were implemented from 1965 to 2016 on the Loess Plateau. The  
502 Budyko equation was employed to quantify the effects of climate change and human activities  
503 on the streamflow variations. Our main conclusions can be summarized as follows.

504 (1) The annual streamflows decreased significantly at Gangu station ( $Z = -6.75$ ) in the Sandu  
505 River watershed and Suide station ( $Z = -3.28$ ) in the Dali River watershed. The decrease  
506 in the annual streamflow was not significant at Zhuangcunyi station ( $Z = -1.75$ ) in the  
507 Hulu River watershed. Abrupt change points were mostly detected in the 1990s for all  
508 watersheds.

509 (2) The hydrological indicators obtained with the IHA/RVA method showed that the degrees  
510 of hydrological change in the Sandu River, Dali River, and Hulu River basins were at  
511 high, medium, and low levels, respectively, thereby suggesting that the streamflow  
512 regimes varied under different soil conservation measures. On the microscale level, the low  
513 flow duration and the number of high flows had great influences at the three stations.

514 (3) The changes in the eco-deficit were more remarkable than those in the eco-surplus at all  
515 stations. The annual eco-deficit increased dramatically after the 1980s at Gangu station,  
516 but there was no obvious change at Zhangcunyi and it only increased moderately after the

517 1990s at Suide station. The annual eco-surplus tended to decrease slightly at all three  
518 stations.

519 (4) The significant annual runoff reductions were primarily attributed to human activities in  
520 the Dali River watershed (91.7%) and Sandu River watershed (74.1%). Climate change  
521 accounted for 64.30% of the annual streamflow decrease in the Hulu River watershed  
522 where afforestation dominated.

523 The streamflow at Gangu station changed greatly due to the significant effect of  
524 numerous terraces, which apparently led to an eco-deficit. The flow regime was altered in the  
525 forest dominated watershed but the overall streamflow was relatively weaker and vulnerable  
526 to climate change. Moreover, human activities made the greatest contribution to the reduced  
527 runoff in the Suide watershed. The results obtained in this study provide novel insights into  
528 hydrological regime changes and their responses to different soil and water conservation  
529 measures.

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534 data.

535

### 536 **Data Availability Statement**

537 Data available on request due to privacy/ethical restrictions.

538

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