Viral infections mediate microbial controls on ecosystem responses to global warming

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Abstract

Climate change is affecting how energy and matter flow within ecosystems, altering global carbon and nutrient cycles. Microorganisms play a fundamental role in carbon and nutrient cycling and are thus an integral link between ecosystems and climate. Here, we highlight a major black box hindering our ability to anticipate ecosystem climate responses: viral infections within complex microbial food webs. We show how understanding and predicting ecosystem responses to warming could be challenging—if not impossible—without accounting for the direct and indirect effects of viral infections on different microbes (bacteria, fungi, protists) that together perform diverse ecosystem functions. Importantly, understanding how rising temperatures associated with climate change influence viruses and virus-host dynamics is crucial to this task, yet severely understudied. In this perspective, we 1) synthesize existing knowledge about virus-microbe-temperature interactions and 2) identify important gaps to guide future investigations regarding how climate change might alter microbial food web effects on ecosystem functioning. To provide real-world context, we consider how these processes may operate in peatlands—globally significant carbon sinks that are threatened by climate change. We stress that understanding how warming affects biogeochemical cycles in any ecosystem hinges on disentangling complex interactions and temperature responses within microbial food webs.

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41 ABSTRACT

42 Climate change is affecting how energy and matter flow within ecosystems, altering global 43 carbon and nutrient cycles. Microorganisms play a fundamental role in carbon and nutrient 44 cycling and are thus an integral link between ecosystems and climate. Here, we highlight a major 45 black box hindering our ability to anticipate ecosystem climate responses: viral infections within 46 complex microbial food webs. We show how understanding and predicting ecosystem responses 47 to warming could be challenging—if not impossible—without accounting for the direct and 48 indirect effects of viral infections on different microbes (bacteria, fungi, protists) that together 49 perform diverse ecosystem functions. Importantly, understanding how rising temperatures 50 associated with climate change influence viruses and virus-host dynamics is crucial to this task, 51 yet severely understudied. In this perspective, we 1) synthesize existing knowledge about virusmicrobe-temperature interactions and 2) identify important gaps to guide future investigations 52 53 regarding how climate change might alter microbial food web effects on ecosystem functioning. 54 To provide real-world context, we consider how these processes may operate in peatlands globally significant carbon sinks that are threatened by climate change. We stress that 55 56 understanding how warming affects biogeochemical cycles in any ecosystem hinges on 57 disentangling complex interactions and temperature responses within microbial food webs. 58 59 60 61

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64 INTRODUCTION

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73

66 to microbial respiration (Dorrepaal *et al.* 2009; Jassey *et al.* 2015; Page and Baird 2016; Masson-

Climate change is warming terrestrial carbon (C) reserves, making them increasingly vulnerable

- 67 Delmotte *et al.* In Press). Because microbial respiration increases with temperature (Zhou *et al.*
- 68 2012; Bradford *et al.* 2019; Smith *et al.* 2019; Wieczynski *et al.* 2021), microbes will likely

69 accelerate carbon release at ever increasing rates as Earth warms, creating a positive atmospheric

- 70 feedback loop not currently represented in predictive models of future climate (Cavicchioli *et al.*
- 71 2019). However, warming is expected to restructure microbial food webs through changes in

recies composition (Petchey *et al.* 1999) (but see (Thakur *et al.* 2021)) and species interactions

(Lurgi, López and Montoya 2012; Barbour and Gibert 2021). Additionally, microbial impacts on

74 carbon cycling are likely mediated by viral infections of both microbes and their predators

75 (Wilhelm and Suttle 1999; Weitz *et al.* 2015; Fischhoff *et al.* 2020). Despite the increasing

recognition that infectious agents like viruses are integral components of food webs (Lafferty et

al. 2008), the role they play in microbial food webs and their associated temperature

78 dependencies remain poorly understood. Identifying and understanding the temperature-

79 dependence of these biotic controls on microbial respiration is paramount to properly forecast

80 current and future ecosystem-climate feedbacks.

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Autotrophic and heterotrophic bacteria, archaea, fungi, and micro-eukaryotes play functionally unique roles in microbial communities as primary producers, nitrogen (N₂)-fixers (diazotrophs), and organic biomass decomposers. For example, microbial autotrophs provide about half of global primary production (Field *et al.* 1998; Litchman *et al.* 2015). Decomposers recycle carbon and nutrients from dead organic matter and act as major carbon emitters by respiring carbon

87 (CO₂ and CH₄) into the atmosphere (Falkowski *et al.* 2000; Canadell *et al.* 2021). The matter recycled by decomposers reaches higher trophic levels through microbial predation-a process 88 known as the "the microbial loop" (Azam et al. 1983; Fenchel 2008). Predation by protists is a 89 90 major source of mortality among microbial primary producers (Geisen et al. 2020) and 91 decomposers (Sherr and Sherr 1988; Gao et al. 2019) (Fig. 1), that can drastically impact carbon 92 and nutrient cycling by reducing microbial biomass, increasing nutrient turnover, and altering 93 microbial respiration rates (Trap et al. 2016; Geisen et al. 2018, 2021; Gao et al. 2019; Rocca et 94 al. 2021). Because of these effects, protists have been called the "puppet masters" of the 95 microbiome (Gao *et al.* 2019). Due to changes in underlying physiological processes, protist 96 predation rates are expected to change with warming (DeLong and Lyon 2020), altering species 97 interactions within microbial food webs (DeLong and Lyon 2020; Thakur et al. 2021) and influencing microbial biomass and respiration rates (O'Connor et al. 2009; Yvon-Durocher and 98 99 Allen 2012; Geisen et al. 2021). This complexity emphasizes the need for a food web 100 perspective to understand microbial responses to changing environmental conditions (Thakur and 101 Geisen 2019).

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Perhaps our biggest oversight in understanding microbial food web responses to global change is
the neglected role of viruses, who have also recently been described as "puppet masters" in the
microbiome (Breitbart *et al.* 2018). All microbes are potential hosts for viruses, which may affect
microbial food web composition and functioning by increasing microbial mortality and, in turn,
nutrient cycling (*via* the Viral Shunt) (Fuhrman 1999; Wilhelm and Suttle 1999; Weinbauer
2004; Suttle 2005). Viruses are the most abundant biological entities on Earth (Weinbauer 2004;
Suttle 2005); therefore, viral mediation of carbon and nutrient flux within microbial food webs is

likely widespread, having important consequences for ecosystem functioning at both local and
global scales (Fuhrman 1999; Wilhelm and Suttle 1999; Weinbauer 2004; Suttle 2005; Weitz *et al.* 2015). Several aspects of the viral infection cycle and virus-host dynamics could potentially
be affected by warming (Table 1), yet the effects of temperature on these processes is unclear
and severely understudied (Fig. 2), undermining our ability to predict how microbial food webs
will respond to global change.

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117 Although the individual effects of microbes and viruses on ecosystem functioning have been 118 discussed (Azam et al. 1983; Fenchel 2008; Quaiser et al. 2015; Ballaud et al. 2016; Stough et 119 al. 2017; Gao et al. 2019; Geisen et al. 2021), we lack a baseline understanding about how these 120 top-down controls jointly influence ecosystem processes within broader microbial food webs and 121 in response to novel climates. Here, we outline the current state of understanding regarding 122 temperature effects on infections within microbial food webs and propose ways to conceptualize 123 and address existing knowledge gaps, with a focus on potential effects of warming on carbon and 124 nutrient cycling. First, we present the current state of kno0wledge regarding the effects of 125 temperature on viruses and viral infections. Next, we integrate viruses into microbial food webs 126 to discuss how viruses might mediate the effects of warming on food web dynamics and 127 functioning. Finally, to provide real-world context for the potential effects of warming on viral 128 infections within microbial food webs, we conclude by exploring how virus-microbe responses 129 to warming may alter ecosystem processes in *Sphagnum* moss-dominated peatlands, which are 130 particularly vulnerable to future climate change (Page and Baird 2016) and, despite occupying 131 less than 3% of the Earth's surface, store ~25–30% of the world's soil carbon (Yu et al. 2010) 132 and produce 5–10% of global atmospheric methane (Blodau 2002).

134	1. TEMPERATURE EFFECTS ON VIRUSES AND VIRAL INFECTIONS
135	All components of microbial food webs can be infected by viruses. While it is recognized that
136	rising temperatures influence the ecology and physiology of microorganisms across
137	environments (Labbate et al. 2016), it is still unclear how the direct and indirect effects of
138	warming will influence viruses, their infection cycles, and how that will ultimately cascade to
139	influence microbial food web functioning. Viral infection occurs in a sequence of steps (Cann
140	2008) (Fig. 2) including 1) host cell encounter, 2) adsorption, 3) introduction of virus or genetic
141	material into the cell, 4) synthesis of viral particles, and 5) assembly and release of viral progeny.
142	Any one, and likely all, of these steps could be temperature dependent (Fig. 2, Table 1, Table
143	S2), but much research is still needed to evaluate the extent and nature of these temperature
144	dependencies. Furthermore, temperature may affect viral production directly by affecting the
145	particle itself (Nagasaki and Yamaguchi 1998) or indirectly by altering host physiology
146	(Kendrick et al. 2014). Understanding each of these temperature effects is paramount to
147	determine how warming might impact carbon and nutrient cycling within microbial food webs.
148	
149	Increasing temperature can cause a decrease in latent period (time from infection until release of
150	viral progeny) and an increase in burst size (number of viral progeny released) (Hadas et al.
151	1997; Nagasaki and Yamaguchi 1998; Demory et al. 2017; Maat et al. 2017; Piedade et al. 2018)
152	(Fig. 2), followed by a reversal of these trends past a virus-specific thermal optimum (T_{opt})
153	(Kimura et al. 2008; Demory et al. 2017). Temperature effects on burst size and latent period are
154	likely the result of host metabolism and virus synthesis kinetics, but direct evidence is lacking.
155	Based on these findings, we hypothesize that future warming may increase infection and viral

production in systems in which current *in situ* temperatures are below T_{opt}, while systems already
near or at T_{opt} may produce fewer viruses or undergo complete shutdown of viral propagation.

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159 Encounter rates between viruses and hosts depend on virus and host densities (Murray and 160 Jackson 1992), host cell size, and host motility (Wilhelm et al. 1998). Host cell sizes (Atkinson, 161 Ciotti and Montagnes 2003; Daufresne, Lengfellner and Sommer 2009; Martin et al. 2020) and 162 population densities (Savage et al. 2004; Bernhardt, Sunday and O'Connor 2018) often decrease 163 while motility increases (Crozier and Federighi 1924; Maeda et al. 1976; Dell, Pawar and Savage 164 2011, 2014; Gibert et al. 2016) with temperature. Consequently, warming could have positive or 165 negative effects on virus-host encounter rates, although more studies are needed (Table 1, Fig. 166 2). Evidence suggests that the effect of temperature on adsorption are dependent on the host-167 virus pair, in some cases increasing (Seeley and Primrose 1980; Hadas et al. 1997), decreasing 168 (Kendrick et al. 2014), or remaining unchanged (Seeley and Primrose 1980) with increases in 169 temperature (Table 1, Fig. 2). While cell membranes are more fluid and permeable at higher 170 temperatures (Marr and Ingraham 1962; Sinensky 1974), it is unknown whether this alters viral 171 infection. We are also unaware of studies that directly link temperature and virus synthesis rates 172 (Fig. 2). Seasonal changes in viral abundances (Nakayama et al. 2007; Payet and Suttle 2007; 173 Colombet et al. 2009) and community composition (Lymer et al. 2008), as well as climatic 174 differences in viral lysis rates (Mojica et al. 2016), have been observed, but confounding factors 175 such as nutrient availability and predation obscure the direct effects of temperature on viral 176 infection cycles. Variation in viral life strategies (*i.e.*, lysis vs. lysogeny in prokaryotes and/or 177 latency in multicellular eukaryotes (Correa et al. 2021)) is ecologically important (Stough et al. 178 2017) and these strategies likely exhibit unique trends with temperature that are currently

179 unresolved (*e.g.*, increasing temperatures may or may not induce lysis (Shan *et al.* 2014)),

180 exposing a crucial gap in our understanding of the temperature-dependencies of viral infection.181

182 Viral production is linked to host cell physiology (Tomaru, Kimura and Yamaguchi 2014; 183 Demory et al. 2017; Maat et al. 2017; Piedade et al. 2018) because viruses depend on and rewire 184 the metabolism of host cells (Hurwitz, Hallam and Sullivan 2013). However, viral temperature 185 ranges can be independent of, and often surpass, those of their hosts (Seeley and Primrose 1980; 186 Mojica and Brussaard 2014; Tomaru, Kimura and Yamaguchi 2014). Additionally, multiple 187 viruses that infect the same host can have different temperature optima (Tomaru, Kimura and 188 Yamaguchi 2014), potentially promoting niche differentiation and a shift in dominant viral taxa 189 with warming. This suggests that viruses could be less susceptible to extinction under warming 190 than their hosts, but more research is needed to determine the extent of this phenomenon and the resulting impacts on nutrient and carbon cycling. 191

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193 Finally, the potential consequences of viral temperature dependencies for microbial food web 194 dynamics and functioning may be complex, context-dependent, and variable across systems. For 195 example, Frenken et al. (2020) used aquatic mesocosm experiments to show that, although 196 warming advanced the seasonal timing of viral infection, it did not increase viral abundance or 197 strengthen viral control over host populations. In addition, <u>Danovaro et al. (2011)</u> predicted that 198 the effects of warming on viral abundance will vary by oceanic region and that a consistent 199 response to rising temperatures across environments is unlikely. These examples illustrate that 200 the temperature-dependent effects of viruses can manifest in different aspects of viral 201 infection/virus-host interactions and may vary by region. We argue that controlled studies (e.g.,

mesocosms, synthetic communities) and *in situ* monitoring across diverse environments can aid
in identifying and predicting complex viral responses to temperature in different environmental
contexts. Moreover, the vast majority of data available for temperature effects on viral dynamics
comes from marine environments or a select few model host-virus systems (Table 1),
highlighting the need to expand studies to different environments and new systems to better
comprehend the influences of virus-microbe interactions on ecosystem processes under warming
conditions.

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210 2. INTEGRATING VIRAL INFECTIONS WITHIN MICROBIAL FOOD WEBS UNDER 211 WARMING

212 Although viruses are known to impact carbon and nutrient cycling directly, namely *via* the viral 213 shunt (Wilhelm and Suttle 1999; Sullivan, Weitz and Wilhelm 2017), how viruses might mediate 214 microbial responses to warming is poorly understood. Microbes account for a substantial fraction 215 of the biomass on Earth (Bar-On, Phillips and Milo 2018) and place major controls on carbon 216 and nutrient cycling in terrestrial (Schimel and Schaeffer 2012), freshwater (Kayranli et al. 217 2010), and marine (Zhang et al. 2018) ecosystems worldwide. Microbial communities are 218 complex, functionally-diverse, multi-trophic food webs (Bengtsson, Setälä and Zheng 1996; 219 Petchey et al. 1999; Gao et al. 2019; Thakur and Geisen 2019) in which energy and matter flow 220 between organisms that occupy different trophic positions and play a variety of functional roles 221 (Fenchel 2008; Steinberg and Landry 2017). Ecosystem responses to climate change are thus 222 likely regulated by changes in overall microbial food web dynamics and organization (Thakur 223 and Geisen 2019; Kuppardt-Kirmse and Chatzinotas 2020). Viruses could play important roles in 224 these changes that depend on i) the relative infection rates of hosts in different functional groups,

ii) the temperature dependencies of the viral infection cycle, iii) thermal matching between virushost pairs, and iv) changes in host physiology, population dynamics, and species interactions
associated with viral infection.

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229 Broadly speaking, how viruses mediate microbial controls on ecosystem responses to warming 230 hinges on how they impact the overall balance of carbon and nutrient uptake (via photosynthesis 231 and decomposition), storage in biomass, sequestration in sediment, and release (via respiration) 232 (Box 1, Figs. 2, 3). Respiration and decomposition rates are expected to increase with warming 233 (Petchey et al. 1999; Kirschbaum 2000; Smith et al. 2019) and may be more sensitive to 234 temperature change than photosynthetic rates (Allen, Gillooly and Brown 2005) (although a 235 great deal of variation exists in temperature sensitivities among different microbial groups 236 (Smith et al. 2019)). This suggests that warming could tip ecosystems from productivity-237 dominant carbon sinks (storing carbon in biomass and sediment) to respiration-dominant carbon 238 sources (releasing carbon into the atmosphere) (Yvon-Durocher and Allen 2012). However, 239 increases in microbial primary productivity should at least partially offset this uneven increase in 240 carbon release (Zhou et al. 2012; Wyatt et al. 2021). Furthermore, warming is expected to alter 241 the biomass and composition of microbial food webs, affecting ecosystem processes like CO2 242 release via respiration (Geisen et al. 2021; Rocca et al. 2022). How viruses mediate this balance 243 between carbon uptake and release under warming is poorly understood, but will likely involve 244 complex and differential impacts on the dynamics and mortality of hosts that perform different 245 ecosystem functions (Sarmento et al. 2010; Danovaro et al. 2011; Vaqué et al. 2019). Based on 246 preliminary model results, we hypothesize that warming could strengthen viral controls on 247 decomposers, N-fixers, and protists, leading to reduced microbial biomass, increased nutrient

248 cycling and respiration, shorter mean residence time of carbon in microbial food web

249 compartments, and shifts in the balance of carbon sequestration and release into the atmosphere

250 (Box 1, Fig. B2d). However, the generality of these effects is very difficult to judge given how

251 much uncertainty remains about the effects of temperature on viral infection, virus-host

252 dynamics, and the impacts of viruses on microbial food web structure.

253

3. PEATLANDS AS A MODEL SYSTEM TO STUDY HOW VIRAL INFECTIONS MEDIATE MICROBIAL FOOD WEB RESPONSES TO WARMING

256 We use peatland microbial food webs as a real-world case study to explore how viral infections 257 may influence the effects of microbial activity on carbon and nutrient cycling in a warming 258 world. Peatlands are typically dominated by Sphagnum peat mosses, storing more carbon (in 259 both living biomass and peat)—and therefore arguably having a greater influence on global 260 carbon cycling and climate—than any other single genus of plants (Clymo and Hayward 1982; 261 Gorham 1991). While Sphagnum plays a primary role in carbon dynamics (Slate, Sullivan and 262 Callaway 2019), it serves a secondary role by insulating permafrost, thus dampening the impacts 263 of rising temperatures on vast amounts of carbon stored in the arctic tundra (Camill and Clark 264 1998). Peatland microbial food webs are uniquely well-suited systems for studying ecosystem 265 responses to global change due to 1) their net impact on the global carbon cycle (Gorham 1991; 266 Dorrepaal et al. 2009; Yu et al. 2010; Bu et al. 2011), 2) the functional diversity of their 267 constituent microbial taxa (Gilbert et al. 1998; Trap et al. 2016; Geisen et al. 2018; Thakur and 268 Geisen 2019), 3) their vulnerability to changes in temperature (Richardson et al. 2018; Norby et 269 al. 2019; Smith et al. 2019; Geisen et al. 2021), and 4) the ability to grow and study Sphagnum 270 moss and associated microbial communities in the laboratory (Altermatt et al. 2015; Geisen et al. 2018; Carrell *et al.* 2019, 2022b) Doing so, however, will require a multifaceted approach—
including characterization of microbial communities in the field, microbial experiments in the
laboratory, -omics approaches, and mathematical modeling (Singh *et al.* 2010; Geisen *et al.*2017), all of which can benefit from cross-scale integration.

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276 We propose that the response of *Sphagnum*-dominated peatlands to warming is regulated by 277 poorly understood controls on carbon and nutrient cycling from microbes and viral infections 278 (Fig. 1, Box 1). Microbes play diverse functional roles in peatlands (Gilbert et al. 1998; Gilbert 279 and Mitchell 2006; Lara et al. 2011; Kostka et al. 2016; Carrell et al. 2022a) (Fig. 3). For 280 example, bacterial and fungal decomposers are primarily responsible for breaking down dead 281 organic material stored within peatlands (Gilbert et al. 1998; Gilbert and Mitchell 2006), a 282 process being accelerated by warming (Dorrepaal et al. 2009). Additionally, Sphagnum's ability 283 to persist in harsh peatland habitats with extremely low mineral nitrogen availability depends on 284 symbiotic interactions with microbial associates (Lindo, Nilsson and Gundale 2013; Kostka et al. 285 2016; Carrell et al. 2022a)-including diazotrophs that colonize the cell surface and water-filled hyaline cells in host plants (Kostka et al. 2016) (Fig. 3). Bacterial methanotrophs are also 286 287 prevalent in boreal peat bogs (Liebner and Svenning 2013; Vile et al. 2014) and not only fix N₂, 288 but supply 5%–20% of CO₂ necessary for Sphagnum photosynthesis via methane oxidation 289 (Larmola et al. 2014). Sphagnum's microbial community composition varies widely with climate 290 (Singer *et al.* 2019) and is expected to shift considerably under warming (Carrell *et al.* 2019; 291 Basińska et al. 2020), likely altering associated microbial food webs (Bengtsson, Setälä and 292 Zheng 1996; Petchey et al. 1999; Geisen et al. 2018; Gao et al. 2019; Thakur and Geisen 2019). 293

294 Peatland ecosystems also harbor a diverse group of viruses that infect prokaryotes and 295 eukaryotes (Ballaud et al. 2016; Emerson et al. 2018; Stough et al. 2018) and are correlated with 296 overall concentrations of both CO₂ and CH₄ (ter Horst *et al.* 2021). Surprisingly, the inferred 297 frequency of protist infections in the Sphagnum microbiome was found to be higher than that of 298 bacterial infection by phages (Stough et al. 2018), although the functional role of protist 299 infection in this system remains unclear. Fungal viruses can have considerable downstream 300 ecological consequences by lysing or altering the phenotypes of fungal decomposers, symbionts, 301 or pathogens in *Sphagnum* (Sutela, Poimala and Vainio 2019). In peatlands, viral community 302 composition, abundance, and lifestyle strategies are influenced by environmental factors, 303 including temperature (Ballaud et al. 2016; Emerson et al. 2018). However, how warming might 304 modify the direct (lytic release of elements) and indirect (altered host phenotype/dynamics and 305 food web processes) effects of viral infections on *Sphagnum*-associated microbial food webs— 306 and carbon and nitrogen cycling in peatlands—is not well understood. Our simple model 307 suggests that viral infections and microbial activity may jointly accelerate the positive effects of 308 warming on C sequestration in peatlands (Box 1, Fig. B2). However, this simple conceptual 309 model is intended as a first attempt to generate hypotheses about the potential impacts of 310 warming, rather than predict future scenarios. Indeed, the mechanisms and parameters governing 311 such interactions between temperature, viruses, protists, and prokaryotes in this model—and the 312 magnitude and direction of resulting changes in carbon cycling—have little empirical 313 verification and will require much more experimental investigation to resolve, thus highlighting 314 the importance of these missing data. A deeper understanding about how these ecological 315 interactions occur in nature and how they are influenced by warming is direly needed, but 316 peatland microbial food webs provide a promising system to begin to develop this understanding.

318	CONCLUSIONS
319	Microbial food webs play a central role in the global carbon cycle by processing and storing vast
320	amounts of carbon. We suggest that viral infections within microbial food web components that
321	play distinct functional roles, and their associated temperature-dependencies, could control
322	changes in carbon cycling and storage in response to global warming. We highlight the
323	importance of studying the complex dynamics of microbial food webs to better understand and
324	predict whether rising temperatures will lead to net carbon sequestration or release in globally
325	important ecosystems like Sphagnum-dominated peatlands. But we also stress that these
326	ecological interactions and their temperature-dependencies are poorly understood, highlighting
327	several gaps for future research. We propose the following list of questions to serve as a guide
328	moving forward:
329	1) How will warming influence different aspects of the viral infection cycle, including both
330	host-dependent and host-independent processes? (Section 1)
331	2) How will virus-host interactions be affected by warming, including virus and host
332	temperature sensitivities, niches, and matching? (Section 1)
333	3) How will warming affect virus life strategies? (Section 1)
334 335	4) How will viral infections mediate the rewiring of functionally- and trophically-diverse microbial food webs under warming? (Section 2)
336	5) How do viral infections alter host physiology, population dynamics and species
337	interactions? (Section 2)
338	6) Will viral infections of functionally distinct microbial groups affect how warming shifts
339	the balance of carbon uptake, storage, and release? (Section 2)
340	7) What are the relative viral abundances and infection rates across microbial hosts in real
341	ecosystems like peatlands? (Section 3)
342	8) How can we leverage empirical data and models to study the coordinated impacts of
343	warming and viral infection on microbial carbon and nutrient cycling? (Section 3)
344	
345	Resolving these uncertainties will require a combination of empirical and theoretical analyses
346	that specifically evaluate temperature-dependencies and virus-host interactions within microbial

347	food webs. The effects of these important processes on microbial population dynamics and
348	carbon flow may then shed light on the broader impacts of warming on carbon cycling and
349	storage within and across whole ecosystems.
350	
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Box 1.

Climate-driven shifts in nutrient and carbon cycling can be studied using mathematical models that track the collective responses of several essential organisms within microbial food webs (Fig. B1). Each organism plays a unique role in carbon and nutrient cycling depending on its metabolic requirements, trophic mode (autotroph, heterotroph), trophic position, stoichiometry, temperature sensitivity, etc. The fate of carbon—storage in biomass, storage in sediment, or respiration into the atmosphere—is therefore controlled by the composition and organization of microbial food webs. Here we develop a conceptual model describing a simplified, example microbial food web from the *Sphagnum*-dominated peatland system and examine potential impacts of warming on ecosystem functioning.

<u>Organisms</u>

- **Decomposers** like heterotrophic bacteria and fungi recycle dead organic matter produced primarily by plants (C uptake) and are major contributors to microbial respiration (C release) and soil organic carbon via mortality (C sequestration).
- **Nitrogen-fixers** like cyanobacteria, methanogenic archaea, and some heterotrophic bacteria transform atmospheric nitrogen (N₂) into biologically usable forms that are metabolically required by all organisms and photosynthetic nitrogen-fixers also require carbon dioxide for photosynthesis (C uptake).
- **Predators** include protists such as heterotrophic flagellates, ciliates, and mixotrophs that consume both decomposers and nitrogen-fixers, altering elemental flows by reducing prey biomass and potentially increasing respiration (C release) and storing recycled carbon and nutrients in predator biomass (C uptake). We use the term "predators" here to differentiate these protists from those that also eat other protists (termed "top predators" below).
- **Eukaryotic algae** include protists that use carbon dioxide for photosynthesis (C uptake) and may represent a significant offset to microbial respiration.
- **Top predators** constitute a subnetwork within the overall food web and include larger protists (*e.g.*, testate amoebae) that consume recycled carbon via predation on all trophic levels, altering biomass and elemental flows throughout (C uptake or release).
- Viruses impact elemental flows directly through lysis (C release) and indirectly by altering host biochemistry and population dynamics (C uptake or release)

Essential elements

- **Inorganic carbon** from the atmosphere (CO₂) is fixed and stored in biomass during photosynthesis and is released through respiration.
- **Organic carbon** is produced by mortality and viral lysis/decay and is transferred between organisms through decomposition and predation.
- **Essential nutrients** like nitrogen and phosphorus are required by all organisms and can affect competitive and trophic dynamics depending on the stoichiometric requirements

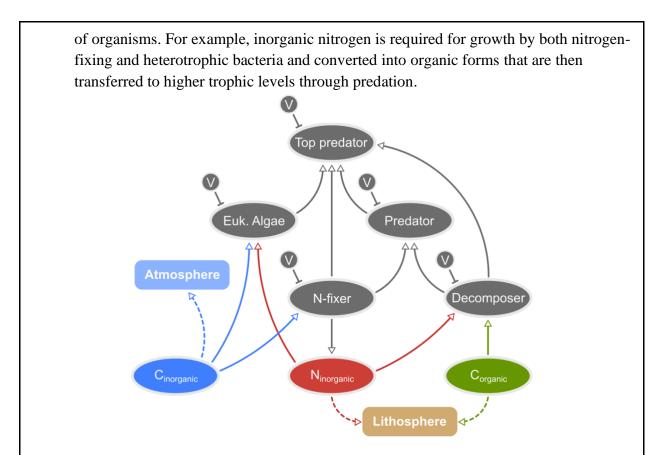


Figure B1. Hypothetical microbial food web in *Sphagnum* peatlands including organisms and nitrogen and carbon flow. Arrows represent flow between components. Each type of organism consumes elements or other organisms based on its unique stoichiometric requirements and is also subject to infection by viruses (V). Unused elements are released into the atmosphere or stored in the lithosphere.

The impacts of global warming on the carbon cycle will ultimately depend on the temperature dependencies of several different processes within microbial food webs, including photosynthesis, respiration, predation, viral infection, and mortality (Fig. 1), many of which are poorly understood for most of these organisms (Figs. 1&4). However, photosynthesis is generally less sensitive to increases in temperature (activation energy of ~0.32eV (Allen, Gillooly and Brown 2005; López-Urrutia *et al.* 2006; O'Connor *et al.* 2009; Yvon-Durocher and Allen 2012)) than respiration and predation (~0.65eV (Brown *et al.* 2004; Dell, Pawar and Savage 2011, 2014)), while mortality lies somewhere in between (~0.45eV (Brown *et al.* 2004; Savage *et al.* 2004)).

Accounting for these temperature dependencies in our hypothetical food web suggests that warming will have little effect on the balance of carbon storage and release in systems composed of only decomposers, fungi, and protists—where carbon released into the atmosphere ($C_{Inorganic}$) is expected to exceed carbon stored in the sediment ($C_{Organic}$) (Fig. B2 a&c). Protists significantly increase the amount of carbon stored but also reduce the amount of

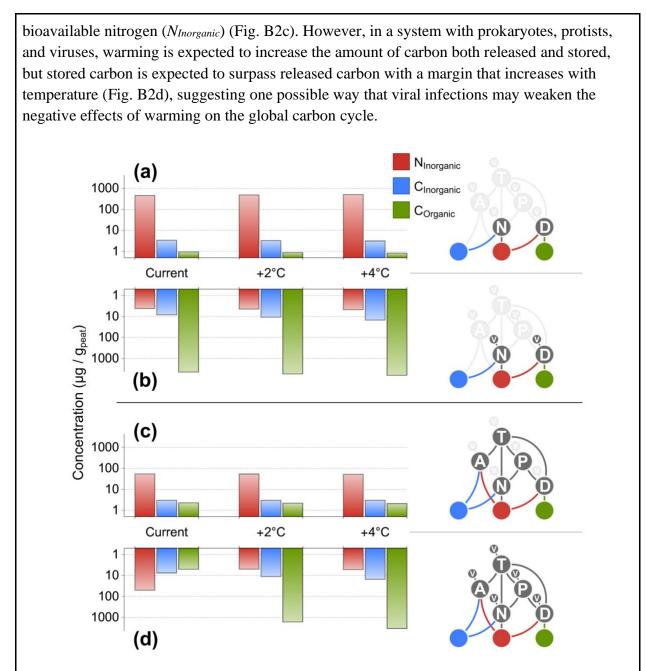


Figure B2. The effects of warming on equilibrium concentrations of nitrogen and carbon in the model microbial food web from Fig. B1. Four scenarios are shown to assess the influences of different food web components: (a) non-protists only (N + D), (b) non-protists + viruses (N + D + V), (c) non-protists + protists (N + D + A + P + T), and (d) all organisms and viruses.

These results are merely suggestions based on limited knowledge of parameter space and many simplifying assumptions. True temperature responses will depend on changes in the composition and structure of specific microbial food webs, several temperature-dependencies that are poorly understood across organisms (Figs. 1&4), possible changes in size across taxa

that could change predation rates (Brose et al. 2012), and temperature-dependence at all stages
of viral infection (Table 1). We stress that all of the parameters, interactions among organisms,
and temperature dependencies outlined in this model are poorly understood and should be the
subject of much-needed future investigation. Hence, the primary role of this model is to
provide a roadmap that identifies the components of microbial food webs that could have
important impacts on carbon flux. We advocate that investigating these unknowns is a critical
step towards more accurately predicting ecosystem responses to climate change.

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386 TABLES

- **Table 1.** Select published studies of temperature effects on viruses. A more detailed description
- 388 of each study, including summarized results, can be found in Table S2.

Process	Temperature Effects	Location or Host-Virus System
Viral decay	Increases with temperature	 Backwater system of Danube River (Field) (Mathias, Kirschner and Velimirov 1995)¹ <i>Heterosigma akashiwo</i> (H93616, NM96) / Hav (HaV01, HaV08) (Lab) (Nagasaki and Yamaguchi 1998)² Bacteriophage 9A isolated from Arctic seawater (Lab) (Wells and Deming 2006)³ Samples from Western Pacific Ocean (Lab) (Wei <i>et al.</i> 2018)⁴
	Increases with temperature	 <i>Escherichia coli</i> / coliphage isolates from the River Swift (Lab) (Seeley and Primrose 1980)⁵ <i>Escherichia coli</i> / T4 (Lab) (Hadas <i>et al.</i> 1997)⁶ <i>Chaetoceros tenuissimus</i> / Cten DNAV and Cten RNAV (Lab) (Tomaru, Kimura and Yamaguchi 2014)⁷
Adsorption	Decreases with temperature	 Chaetoceros tenuissimus / Cten DNAV and Cten RNAV (Lab) (Tomaru, Kimura and Yamaguchi 2014)⁷ Emiliana huyxleyi CCMP374 / EhV86 (Lab) (Kendrick et al. 2014)⁸
	No effect of temperature	 Escherichia coli / coliphage isolates from the River Swift (Lab) (Seeley and Primrose 1980)⁵
Burst size	Increases with temperature	 Backwater system of Danube River (Field) (Mathias, Kirschner and Velimirov 1995)¹ Escherichia coli / T4 (Lab) (Hadas et al. 1997)⁶ Micromonas sp. MicA, MicB, MicC / MicVA, MicVB, MicVC (Lab) (Demory et al. 2017)⁹ Micromonas polaris / MpoV (Lab) (Maat et al. 2017)¹⁰ Micromonas polaris strain RCC2257, strain RCC2258 / Mpov-45T (Lab) (Piedade et al. 2018)¹
	Decreases with temperature	 Backwater system of Danube River (Field) (Mathias, Kirschner and Velimirov 1995)¹ <i>Micromonas</i> sp. MicA, MicB, MicC / MicVA, MicVB, MicVC (Lab) (Demory <i>et al.</i> 2017)⁹
	Increases with temperature	 <i>Micromonas</i> sp. MicA, MicB, MicC / MicVA, MicVB, MicVC (Lab) (Demory <i>et al.</i> 2017)⁹ <i>Escherichia coli</i> / coliphage (Lab) (Ellis and Delbrück 1939)¹²
Latency period	bcy period Decreases with temperature	 Heterosigma akashiwo (H93616, NM96) / Hav (HaV01, HaV08) (Lab) (Nagasaki and Yamaguchi 1998)² Escherichia coli / T4 (Lab) (Hadas et al. 1997)⁶ Micromonas sp. MicA, MicB, MicC / MicVA, MicVB, MicVC (Lab) (Demory et al. 2017)⁹ Micromonas polaris / MpoV (Lab) (Maat et al.

			 2017)¹⁰ <i>Micromonas polaris</i> strain RCC2257, strain RCC2258 / Mpov-45T (Lab) (Piedade <i>et al.</i> 2018)¹¹ <i>Escherichia coli</i> / coliphage (Lab) (Ellis and Delbrück 1939)¹² <i>Staphylococcus aureus</i> / S. aureus phage (Lab) (Krueger and Fong 1937)¹³
	Virus abundance	Temperature effects unclear	 Backwater system of Danube River (Field) (Mathias, Kirschner and Velimirov 1995)¹ Southern Beaufort Sea and Amundsen Gulf (Field) (Payet and Suttle 2007)¹⁴ Lake Pavin (Field) (Colombet <i>et al.</i> 2009)¹⁵ Japanese paddy field (Field) (Nakayama <i>et al.</i> 2006)¹⁶ Michigan agricultural soils (Field) (Roy <i>et al.</i> 2020)¹⁷ Metadata (Danovaro <i>et al.</i> 2011¹⁸; Williamson <i>et al.</i> 2017¹⁹)
	Lysis thermal range	Temperature effects are host- dependent	 Heterosigma akashiwo (H93616, NM96) / Hav (HaV01, HaV08) (Lab) (Nagasaki and Yamaguchi 1998)² Bacteriophage 9A isolated from Arctic seawater (Lab) (Wells and Deming 2006)³ Escherichia coli / coliphage isolates from the River Swift (Lab) (Seeley and Primrose 1980)⁵ Metadata (Mojica and Brussaard 2014)
	Virus-induced host mortality	Increases with temperature	- North Atlantic Ocean (Field) (Mojica <i>et al.</i> 2016)
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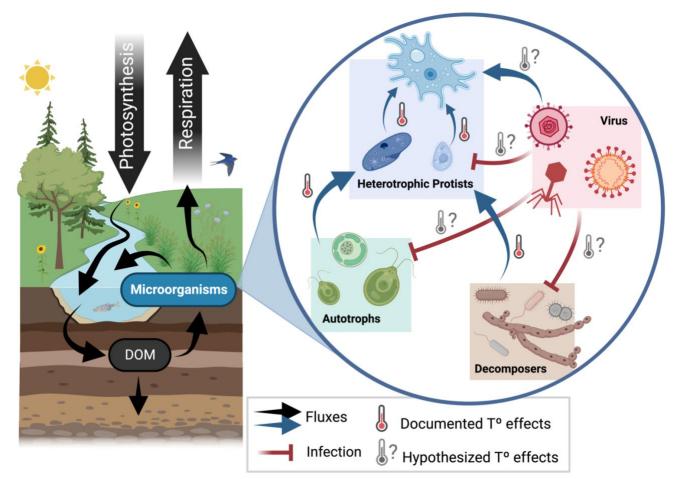


Figure 1. Conceptual diagram outlining the documented and hypothesized temperature effects
on processes influencing global carbon cycling, including the impacts of decomposers
(heterotrophic bacteria, archaea, and fungi), autotrophs (cyanobacteria and eukaryotic algae),
heterotrophic protists that consume all organisms, and viruses that infect all organisms. Note that
some organisms (prokaryotes and eukaryotes) can occupy both autotrophic and heterotrophic
compartments (mixotrophs).

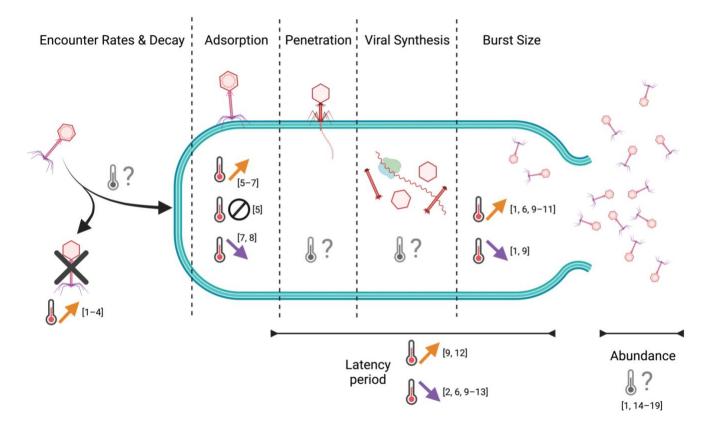




Figure 2. Stages of the viral lytic infection cycle and published temperature effects. Orange
arrows indicate a positive effect, purple arrows indicate a negative effect, and interdictory
symbols indicate no effect with warming. Gray thermometers indicate stages of the viral
infection cycle that either have no published experimental data or published effects are
confounded by other environmental/biological factors (*e.g.* abundances from field studies).
Numbers correspond to references in Table 1. More details from these studies can be found in
Table S2.

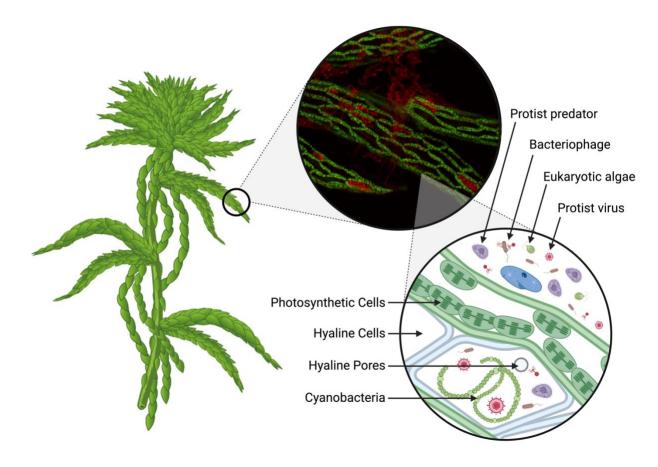


Figure 3. *Sphagnum* moss and associated microbial food web. Microbial species inhabit both
water-filled hyaline cells of *Sphagnum* tissue and the external aquatic habitat. First inset shows
cyanobacteria (in red) living inside *Sphagnum* tissue (in green, image taken using a Zeiss LSM
710 laser scanning confocal microscope, image credit: Andrea Timm and Collin Timm).

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SUPPORTING INFORMATION

Supplementary methods

Microbial food web model

To illustrate the potential impacts of temperature, microbial food web structure, and viral infection on the carbon and nutrient cycling, we developed a mathematical model to study the dynamics of an assortment of organisms that exist at different trophic levels and play distinct functional roles within microbial food webs—including N-fixers (NF), decomposers (D), eukaryotic algae (A), protist grazers (G), protist top predators (P), and viruses (V_i) that exclusively infect each organism (Box 1, Figure B1). The model also includes pools (external to organisms) of relevant essential elements-including, inorganic nitrogen (N1; converted from N2 by N-fixers), inorganic carbon (C_I ; i.e., carbon fraction of CO₂), and organic carbon (C_O ; carbon fraction of dead organic matter). These pools of essential elements are available for use by organisms and their concentrations are influenced by biological processes (e.g., photosynthesis, respiration, and mortality). Biological populations and elemental pools are referred to in terms of mass concentrations standardized by units of peat mass (units of $\mu g / g$ of peat). The dynamics of all components are governed by a system of ordinary differential equations (Eqns. S1-S13). Variable and parameter definitions, units, and values used for analysis are given in Table S2. Parameter values were chosen such that all organisms exhibited non-zero equilibrium densities using the same parameter values across all biological scenarios shown in Figure B2, allowing for more direct comparison of biological scenarios.

In this model, all basal organisms (i.e., organisms that do not consume other organisms; NF, D, A) grow logistically and consume elements from external pools (N_I , C_I , C_O) according to their modes of energy acquisition: autotrophs (NF and A) use C_I , non-N-fixers (D and A) use N_I , and decomposers (D) use Co. Element uptake rates follow Michaelis-Menten kinetics. Biomass production rates in all organisms is reduced by inefficient conversion of resources (ε_i). Conversion efficiency in consumers is also reduced according to the lowest stoichiometric ratio (carbon or nitrogen) between a given resource organism and its consumer (*qelement, resource* / *q*element, consumer; i.e., Leipig's law of the minimum). All organisms are infected by viruses that are specific to each host. All elemental pools operate as chemostats with an inflow rate (α_k) and an outflow rate (δ_k). Inorganic nitrogen (N_l) increases with respiration and decreases with growth of decomposers (D) and eukaryotic algae (A). Inorganic carbon (C_1) increases with respiration and decreases with growth of N-fixers (NF) and eukaryotic algae (A). Organic carbon increases with mortality (*m*, organisms and viruses) and viral lysis (ϕ) and decreases with growth of decomposers (D). All temperature dependencies follow Sharpe-Schoolfield functional forms (Schoolfield et al. 1981) (Eqn. S14) with activation energies that are specific to each rate: respiration (0.65eV (Brown et al. 2004)), photosynthesis (0.32eV (Allen et al. 2005)), mortality (0.45eV (Savage et al. 2004)), and consumption (0.65eV (Brown et al. 2004; Dell et al. 2011a)). Viral lysis rates and burst sizes were assumed to follow established activation energies of consumption (0.65eV). Although we assume these temperature sensitivities here for simplicity, we note that a great deal of variation exists in the activation energies of various metabolic processes and across taxa (Dell et al. 2011b; Smith et al. 2019) and that this variation could

affect overall food web responses to warming. More specific temperature responses could easily be incorporated in future models by replaced those used here.

Nitrogen-fixer:	$\dot{NF} = NF\left(\varepsilon_{N}\mu_{NF}(T)\frac{C_{I}}{h_{C_{I},NF} + C_{I}}\left(1 - \frac{NF}{K_{NF}}\right) - a_{NF,G}(T)G - a_{NF,P}(T)P - \phi_{NF}(T)V_{NF} - r_{NF}(T) - m_{NF}(T)\right)$	(S1)
Decomposer:	$\dot{D} = D\left(\varepsilon_D \mu_D(T) \frac{N_I}{h_{N_I,D} + N_I} \frac{C_O}{h_{C_O,D} + C_O} \left(1 - \frac{D}{K_D}\right) - a_{D,G}(T)G - a_{D,P}(T)P - \phi_D(T)V_D - r_D(T) - m_D(T)\right)$	(S2)
Eukaryotic Algae:	$\dot{A} = A\left(\varepsilon_{A}\mu_{A}(T)\frac{N_{I}}{h_{N_{I},A} + N_{I}}\frac{C_{I}}{h_{C_{I},A} + C_{I}}\left(1 - \frac{A}{K_{A}}\right) - a_{A,P}(T)P - \phi_{A}(T)V_{A} - r_{A}(T) - m_{A}(T)\right)$	(S3)
Grazer:	$\dot{G} = G\left(\varepsilon_{G}min\left(\frac{q_{N,NF}}{q_{N,G}}, \frac{q_{C,NF}}{q_{C,G}}\right)a_{NF,G}(T)NF + \varepsilon_{G}min\left(\frac{q_{N,D}}{q_{N,G}}, \frac{q_{C,D}}{q_{C,G}}\right)a_{D,G}(T)D - a_{G,P}(T)P - \phi_{G}(T)V_{G} - r_{G}(T) - m_{G}(T)\right)$	(S4)
Predator:	$\begin{split} \dot{P} &= P\left(\varepsilon_{P}min\left(\frac{q_{N,NF}}{q_{N,P}},\frac{q_{C,NF}}{q_{C,P}}\right)a_{NF,P}(T)NF + \varepsilon_{P}min\left(\frac{q_{N,D}}{q_{N,P}},\frac{q_{C,D}}{q_{C,P}}\right)a_{D,P}(T)D + \varepsilon_{P}min\left(\frac{q_{N,A}}{q_{N,P}},\frac{q_{C,A}}{q_{C,P}}\right)a_{A,P}(T)A \\ &+ \varepsilon_{P}min\left(\frac{q_{N,G}}{q_{N,P}},\frac{q_{C,G}}{q_{C,P}}\right)a_{G,P}(T)G - \phi_{P}(T)V_{P} - r_{P}(T) - m_{P}(T)\right) \end{split}$	(\$5)
Virus (N-fixer):	$\dot{V}_{NF} = V_{NF} \left(\beta_{NF}(T) \phi_{NF}(T) NF - m_V(T) \right)$	(S6)
Virus (Decomposer):	$\dot{V}_D = V_D \big(\beta_D(T) \phi_D(T) D - m_V(T) \big)$	(S7)
Virus (Algae):	$\dot{V}_A = V_A \big(\beta_A(T) \phi_A(T) A - m_V(T) \big)$	(S8)
Virus (Grazer):	$\dot{V}_G = V_G \big(\beta_G(T) \phi_G(T) G - m_V(T) \big)$	(S9)
Virus (Predator):	$\dot{V}_{P} = V_{P} \big(\beta_{P}(T) \phi_{P}(T) P - m_{V}(T) \big)$	(S10)
Inorganic Nitrogen (N _l):	$\dot{N}_{I} = \alpha_{N_{I}} + r_{NF}(T)q_{N,NF}NF + r_{D}(T)q_{N,D}D + r_{A}(T)q_{N,A}A + r_{G}(T)q_{N,G}G + r_{P}(T)q_{N,P}P - q_{N,D}\mu_{D}(T)\frac{N_{I}}{h_{N_{I},D} + N_{I}}\frac{C_{o}}{h_{C_{o},D} + C_{o}}D - q_{N,A}\mu_{A}(T)\frac{N_{I}}{h_{N_{I},A} + N_{I}}\frac{C_{I}}{h_{C_{I},A} + C_{I}}A - \delta_{N_{I}}N_{I}$	(\$11)
Inorganic Carbon (<i>C</i> _l):	$\dot{C}_{I} = \alpha_{C_{I}} + r_{NF}(T)q_{C,NF}NF + r_{D}(T)q_{C,D}D + r_{A}(T)q_{C,A}A + r_{G}(T)q_{C,G}G + r_{P}(T)q_{C,P}P - q_{C,NF}\mu_{NF}(T)\frac{C_{I}}{h_{C_{I},NF} + C_{I}}NF - q_{C,A}\mu_{A}(T)\frac{N_{I}}{h_{N_{I}A} + N_{I}}\frac{C_{I}}{h_{C_{I}A} + C_{I}}A - \delta_{C_{I}}C_{I}$	(812)
Organic Carbon (<i>Co</i>):	$ \dot{C}_{o} = \alpha_{C_{o}} + m_{NF}(T)q_{C,NF}NF + m_{D}(T)q_{C,D}D + m_{A}(T)q_{C,A}A + m_{G}(T)q_{C,G}G + m_{P}(T)q_{C,P}P + m_{V}(T)q_{C,V}(V_{NF} + V_{D} + V_{A} + V_{G} + V_{P}) + \phi_{NF}(T)V_{NF}q_{C,NF}NF + \phi_{D}(T)V_{D}q_{C,D}D + \phi_{A}(T)V_{A}q_{C,A}A + \phi_{G}(T)V_{G}q_{C,G}G + \phi_{P}(T)V_{P}q_{C,P}P - q_{C,D}\mu_{D}(T)\frac{N_{I}}{h_{N_{I},D} + N_{I}}\frac{C_{O}}{h_{C_{O},D} + C_{O}}D - \delta_{C_{O}}C_{O} $	(\$13)

Supplementary Tables

Variable/Parameter	Definition	Units	Value
(NF, D, A, G, P)	Biomass conc.	$\mu g g_{peat}^{-1}$	na
(N_I, C_I, C_O)	Nutrient conc.	$\mu g g_{peat}^{-1}$	na
ε _i	Production efficiency	na	0.8
$\mu_i(T)$	Max growth rate	d ⁻¹	2.5
$h_{k,i}$	Half-saturation constant	g	10
K _i	Carrying capacity	$\mu g g_{peat}^{-1}$	$K_{NF}, K_A = 500$ $K_D = 1000$
$a_{i,j}(T)$	Consumption rate	$d^{-1} (\mu g \ g_{peat})^{-1}$	$a_{NF,G}, a_{D,G} = 0.01$ $a_{NF,P}, a_{D,P} = 0.0001$ $a_{A,P} = 0.001$ $a_{G,P} = 0.08$
$\phi_i(T)$	Lysis rate	$d^{-1} (\mu g g_{peat})^{-1}$	0.01
$r_i(T)$	Respiration rate	d-1	$r_{NF}, r_A = 0.05$ $r_D = 0.09$ $r_G = 0.2$ $r_P = 0.3$
$m_i(T)$	Mortality rate	d ⁻¹	$m_{NF} = 0.05$ $m_D = 0.01$ $m_A, m_G, m_P = 0.1$
$q_{k,i}$	Elemental content	g g ⁻¹	$q_{N,NF}, q_{N,D} = 0.05$ $q_{N,A}, q_{N,G} = 0.03$ $q_{N,P} = 0.08$ $q_{C} = 0.5$
$\beta_i(T)$	Burst size	$d^{-1} (\mu g g_{peat})^{-1}$	$\beta_{NF}, \beta_D = 0.05$ $\beta_A, \beta_G, \beta_P = 0.03$
α_k	Inflow rate	$\mu g g_{peat}^{-1} d^{-1}$	$\alpha_{N_I} = 6$ $\alpha_{C_I} = 100$ $\alpha_{C_O} = 30$
δ_k	Outflow rate	d-1	$\delta_{N_I}, \delta_{C_I}, \delta_{C_O} = 0.01$

Table S1. Variables and parameters used in the microbial food web model. For parameters that are functions of temperature (f(T)), values are given at a reference temperature of 20°C.

Table S2. Detailed description and summarized results for select published studies of temperature effects on viruses.

Type of Study	Location or Host- Virus system	Observed Temperature Effects	Reference
Environmental	Backwater system of Danube River	 Higher temperature induced higher viral decay rates Viral abundance was tightly correlated with seasonal bacterial abundance one year, but not the next The lowest percentage of bacteria infected by phage were observed at 23-26°C, the highest at 6-22°C, and between at ≤ 5°C Burst size was temperature dependent 	(Mathias <i>et al.</i> 1995)
Laboratory	Heterosigma akashiwo (H93616, NM96) / Hav (HaV01, HaV08)	 Decay rates increased with increasing temperature Latent phase decreased with increasing temperature Thermal ranges of lysis by virus were unique for different host-virus pairs 	(Nagasaki & Yamaguchi 1998)
Laboratory	Bacteriophage 9A isolated from Arctic seawater	• The half-life of infective phages decreased with increasing temperature	(Wells & Deming 2006)
Laboratory	Samples from Western Pacific Ocean	 Increases in temperature and photosynthetic radiation resulted in higher virus decay rates Low fluorescence viruses were more sensitive to warming and increased PAR than high fluorescence viruses 	(Wei <i>et al.</i> 2018)
Metadata	N/A	• Temperatures at which most marine viruses are inactivated fall outside of the host temperature range	(Mojica & Brussaard 2014)
Laboratory	<i>Escherichia coli /</i> coliphage isolates from the River Swift	 Temperature range of phages were independent of host growth temperature Temperature was seen to affect the adsorption of 2 phages and the multiplication of another 2 	(Seeley & Primrose 1980)
Laboratory	Escherichia coli / T4	 Adsorptions rates increased with increasing growth rate and positively correlated with cell size The rate of phage release and burst size increased with growth rate, but the length of the eclipse and latent periods decreased with growth rate Burst size was dependent on both growth rate and time until lysis 	(Hadas <i>et al.</i> 1997)
Laboratory	<i>Emiliana huyxleyi</i> CCMP374 / EhV86	• 3°C increase in temperature induces a viral resistant host phenotype	(Kendrick <i>et al.</i> 2014)
Laboratory	Chaetoceros tenuissimus / Cten DNAV and Cten RNAV	 Susceptibility of all strains to Cten DNAV increased with temperature up to T_{opt} Temperature range and degree of susceptibility to Cten RNAV was strain dependent Maximum burst size of Cten DNAV and minimum burst size of Cten RNAV were both observed between 15-20°C 	(Tomaru <i>et al.</i> 2014)

Laboratory	Staphylococcus aureus / S. aureus phage	• The rate of phage production is related to the growth rate of the host. Higher growth rates up to $T(opt)$ result in shorter latency periods, though $T > T(opt)$ result in longer latency periods	(Krueger & Fong 1937)
Laboratory	<i>Escherichia coli /</i> coliphage	• Latency period decreases with increasing temperature and is directly inversely proportional to the division rate of bacteria	(Ellis & Delbrück 1939)
Laboratory	<i>Micromonas</i> sp. MicA, MicB, MicC / MicVA, MicVB, MicVC	 At temperatures < T_{opt}, latent periods were increased, host cell lysis was delayed, and viral yield was reduced Cell lysis did not usually occur at temperatures > T_{opt} At temperatures slightly above T_{opt}, chronic infection (viral production with no cell lysis) was observed At temperatures much above T_{opt}, no viral progeny were produced 	(Demory <i>et al.</i> 2017)
Laboratory	<i>Micromonas polaris</i> / MpoV	• Higher temperatures resulted in shorter latent periods and increased burst sizes	(Maat <i>et al.</i> 2017)
Laboratory	<i>Micromonas polaris</i> strain RCC2257, strain RCC2258 / Mpov-45T	• Higher temperature (7°C vs. 3°C) caused earlier cell lysis and increased burst size, except in low light conditions	(Piedade <i>et al.</i> 2018)
Environmental	Southern Beaufort Sea and Amundsen Gulf	 Seasonal and spatial variation in virus concentrations were correlated with Chl-a concentration, bacterial abundance and composition, temperature, salinity, and depth Percentage of variance explained by temperature was inconsistent between seasons 	(Payet & Suttle 2007)
Environmental	Lake Pavin	 Virus abundances correlated most closely with host abundance Surface bacterial abundances were largely influenced by temperature while monimolimnion bacterial abundances likely influenced by organic matter export during surface blooms 	(Colombet <i>et al.</i> 2009)
Metadata	N/A	 Positive relationships were observed between viral abundance and temperature within all distinct oceanic regions examined, however a global decreasing trend was seen across these regions when all data was assessed together Water column viral production increased with temperature in polar and cold temperate regions, but decreased with temperature in warm temperate systems 	(Danovaro <i>et al.</i> 2011)
Environmental	Japanese paddy field flood waters	• Viral abundance changed seasonally, but was highly correlated with bacterial abundance	(Nakayama <i>et al.</i> 2007)
Environmental	North Atlantic Ocean	• Shift from virus-induced to grazing-induced phytoplankton mortality with increased latitude (decreased temperature)	(Mojica <i>et al.</i> 2016)

Environmental	Michigan agricultural soils	-Viral abundance changed seasonally; abundance was highly correlated to bacterial abundance, organic carbon content and total nitrogen	(Roy <i>et al.</i> 2020)
Metadata	Global	-Viral abundances are several orders of magnitude higher in cold deserts compared to hot deserts	(Williamson <i>et al.</i> 2017)

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